

UNIVERSIDADE FEDERAL DE MINAS GERAIS
Instituto de Ciências Biológicas
Programa de Pós-Graduação em Ecologia, Conservação e
Manejo de Vida Silvestre

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**A STUDY OF THE PROCESS AND PATTERNS THAT DRIVE FLORIVORY ACROSS
PLANT LINEAGES AND ENVIRONMENTAL CONTEXTS**

Belo Horizonte,
2023

Maria Gabriela Boaventura de Castro

**A STUDY OF THE PROCESS AND PATTERNS THAT DRIVE FLORIVORY ACROSS
PLANT LINEAGES AND ENVIRONMENTAL CONTEXTS**

Tese apresentada ao programa de Pós-graduação em Ecologia, Conservação e Manejo de Vida Silvestre, da Universidade Federal de Minas Gerais, como requisito parcial à obtenção do título de doutora em Ecologia.

Orientadora: Dr^a Tatiana Cornelissen

Belo Horizonte,

2023

043

Castro, Maria Gabriela Boaventura de.

A study of the process and patterns that drive florivory across plant lineages and environmental contexts [manuscrito] / Maria Gabriela Boaventura de Castro. – 2023.

116 f. : il. ; 29,5 cm.

Orientadora: Dra Tatiana Cornelissen.

Tese (doutorado) – Universidade Federal de Minas Gerais, Instituto de Ciências Biológicas. Programa de Pós-Graduação em Ecologia Conservação e Manejo da Vida Silvestre.

1. Ecologia. 2. Flores. 3. Herbivoria. 4. Urbanização. I. Cornelissen, Tatiana Garabini. II. Universidade Federal de Minas Gerais. Instituto de Ciências Biológicas. III. Título.

CDU: 502.7



UNIVERSIDADE FEDERAL DE MINAS GERAIS
INSTITUTO DE CIÊNCIAS BIOLÓGICAS
PROGRAMA DE PÓS-GRADUAÇÃO EM ECOLOGIA, CONSERVAÇÃO E MANEJO DA VIDA SILVESTRE



Ata da Defesa de Tese

Nº 220
Entrada: 2018/2

Maria Gabriela Boaventura de Castro

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FINANCIAMENTO:



U F *m* G

AGRADECIMENTOS

À minha orientadora Tatiana Cornelissen por ter acreditado no meu trabalho, por ensinar com amor, pela amizade, pela generosidade e acolhimento, pelo incentivo e apoio em todos os momentos, tanto no âmbito profissional quanto no pessoal ao longo destes dez anos de parceria. Tati você é uma luz no meu caminho e um exemplo de profissional e pessoa que eu espero ser. Sou eternamente grata a ti por tudo!

Aos colaboradores da minha pesquisa, especialmente ao professor Fernando Silveira por todo apoio e dedicação. Agradeço também ao Filipe Aguiar e a Marina Andrade, alunos de iniciação científica, pelo apoio nas coletas e pelo aprendizado compartilhado.

À Universidade Federal de Minas Gerais (UFMG) e aos professores do Programa de Pós-Graduação em Ecologia, Conservação e Manejo da Vida Silvestre (ECMVS), pela oportunidade, apoio e excelente formação acadêmica. Agradeço também aos secretários do programa, Frederico e Cristiane, por todo o apoio administrativo, paciência e carinho.

A todos os colegas e amigos do Centro de Síntese Ecológica e Conservação (CSEC) pelo companheirismo e apoio.

Aos professores, Geraldo Wilson Fernandes, Lorena Ashworth, Milton Barbosa, Irene Gélvez-Zúñiga, Gisele Mendes e Fernanda Costa por aceitarem gentilmente o convite para avaliar este trabalho.

À CAPES pela concessão da bolsa de pesquisa.

À minha família que é minha base e o meu equilíbrio. À minha mãe Liliane, ao meu pai José e meus irmãos Henrique e Guilherme pelo apoio incondicional e amor, por sempre acreditarem no meu potencial e pela ajuda nas coletas deste estudo. Ao meu companheiro Francisco pelo suporte, apoio, compreensão, amizade e amor. Ao meu filho de quatro patas Tião, por sempre fazer meus dias mais leves e felizes.

A todos os meus amigos queridos que sempre apoiaram e incentivaram essa longa caminhada, especialmente a Lorena, Bruna, Marcílio, Helbert, Isa, Nati, Red, Aninha, Júlia e Nathalia. Agradeço também ao Márcio (*in memoriam*) pelo carinho e apoio, especialmente nos momentos difíceis. Por último, agradeço a todos os outros que de alguma forma fizeram parte desse ciclo. MUITO OBRIGADA!

RESUMO

Este trabalho é o resultado de um estudo que teve como objetivo avaliar os padrões gerais e processos que determinam as interações entre flores e florívoros. No capítulo I nós resumimos as informações disponíveis até o momento sobre florivoria, através de uma revisão integrativa, que reuniu dados de estudos publicados até 2020 e dados inéditos, coletados em campo através de um protocolo padronizado. Nós comparamos os níveis de danos florais entre plantas tropicais e temperadas, examinamos os efeitos das características funcionais das plantas nos danos florais e exploramos a dinâmica eco-evolutiva das interações flor-florívoro para 180 espécies de plantas, em 64 famílias, distribuídas ao redor do mundo. Além disso, nós identificamos os principais caminhos de pesquisa que ajudarão a preencher lacunas persistentes de conhecimento, fornecemos o primeiro banco de dados global e de acesso livre com informações sobre interações flor-florívoro e propusemos um protocolo padronizado para medições de florivoria. Com flores de 61 espécies, distribuídas em 20 famílias e coletadas usando o protocolo padronizado para medições de florivoria supracitado, no capítulo II nós avaliamos a influência de alguns fatores abióticos (ambientais) e características florais nos níveis de florivoria. O objetivo deste capítulo II foi examinar se as mudanças ambientais causadas pela urbanização influenciam os níveis de florivoria e se os níveis de florivoria são impulsionados por atributos de plantas ligados a urbanização e a características funcionais e filogenéticas de flores e espécies. Neste capítulo nós fornecemos uma das primeiras avaliações de florivoria ao nível da comunidade de espécies, através de uma nova abordagem, no contexto das alterações ambientais que ocorrem devido a urbanização, preenchendo assim uma importante lacuna do conhecimento.

Palavras-chave: florivoria, herbívoros florais, dano floral, inimigos florais, florívoros, urbanização.

ABSTRACT

This work is the result of a study aimed at evaluating the general patterns and processes that determine the interactions between flowers and florivores. In Chapter I, we summarized the available information to date on florivory through an integrative review that brought together data from published studies up to 2020 and unpublished data collected in the field using a standardized protocol. We compared levels of floral damage between tropical and temperate plants, examined the effects of plant functional traits on floral damage, and explored the eco-evolutionary dynamics of flower-florivore interactions for 180 plant species in 64 families distributed worldwide. Additionally, we identified key research pathways that will help fill persistent knowledge gaps, provided the first global and freely accessible database with information on flower-florivore interactions, and proposed a standardized protocol for measuring florivory. In Chapter II, using flowers from 61 species belonging to 20 families and collected using the a standardized protocol for florivory collect and measurement's, we assessed the influence of abiotic (environmental) factors and floral traits on florivory levels. The aim of this Chapter II was to examine whether environmental changes caused by urbanization influence florivory levels and whether florivory levels are driven by plant attributes related to urbanization, as well as functional and phylogenetic traits of flowers and species. In this chapter, we provided one of the first assessments of florivory at the species community level through a novel approach, in the context of environmental changes occurring due to urbanization, thus filling an important knowledge gap.

Keywords: florivory, floral herbivory, floral damage, floral enemies, florivores, urbanization.

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INTRODUÇÃO GERAL

Interações entre insetos e plantas desempenham um papel fundamental na dinâmica das populações (Marquis, 1984), na estrutura das comunidades (Agrawal e Maron, 2022), nos processos evolutivos e serviços ecossistêmicos (Turcotte et al., 2014a; Kozlov et al., 2017). Essas interações compreendem processos que variam desde interações mutualísticas e positivas, como a polinização, ou interações antagônicas e negativas, como a herbívoros e a florívoros (Strauss et al. 1996; McCall e Irwin, 2006; Bronstein et al. 2007).

Os padrões gerais de herbívoros foliar vem sendo bem documentados nos últimos anos (Coley e Barone, 1996; Metcalfe et al., 2014; Turcotte et al., 2014b; Kozlov et al., 2015; Mendes et al., 2021), mas os padrões de florívoros, por outro lado, seguem amplamente pouco avaliados e indefinidos, já que a maioria dos estudos tende a se concentrar na herbívoros como relação antagonística e na polinização como relação mutualística (McCall e Irwin, 2006; McCall, 2008; Gorden e Adler, 2016). A florívoros é definida como qualquer tipo de dano a flores e envolve o consumo de tecidos florais por animais vertebrados ou invertebrados, como insetos (McCall e Irwin, 2006; Fig.1). Florívoros podem afetar as plantas tanto diretamente ao consumirem partes florais efetivamente ligadas à reprodução como pistilos, estames e óvulos, reduzindo o número de gametas produzidos – como indiretamente, quando ao consumirem pétalas, sépalas e botões florais eles modificam atributos florais relacionados à atratividade, interferindo na relação planta-polinizador (McCall e Irwin, 2006; Caruso et al., 2019, Fig.1). Como consequência, os florívoros podem causar uma pressão seletiva negativa sobre os atributos florais que atraem polinizadores, o que pode levar a uma redução no serviço ecológico prestados por eles (McCall e Irwin, 2006; Caruso et al., 2019; Rusman et al., 2019).



Fig.1. Florivoria e florívoros em algumas das espécies de plantas avaliadas no presente estudo. A- Flor e florívoros de *Solanum lycocarpum*. B- Florivoria em *Bauhinia variegata*. C- *Diabrotica speciosa* (Chrysomelidae), florívoro de *Turnera subulata*. D- Formiga florívora de *Perianthra mediterranea*. E- Besouro Chrysomelidae alimentando-se de pétalas de *Chamaecrista* sp. F- Florivoria em *Ceiba speciosa*. G- Besouro *Conognatha compta* alimentando-se de pétalas de *Microliceae congestiflora*. H- Florivoria em *Trembleya laniflora*. I- Florivoria em *Aspilia* sp.

Apesar do potencial impacto negativo da florivoria, ainda não temos uma compreensão completa sobre níveis gerais de dano floral ou, sobre como seus efeitos diretos e indiretos afetam as plantas e quais os mecanismos e processos que tornam certas espécies mais vulneráveis a esse antagonismo. Acreditamos que os efeitos da florivoria no *fitness* das plantas dependam de vários fatores como, da intensidade dos danos florais (quantidade de área floral removida), identidade das plantas e dos florívoros e a sua guilda alimentar e de fatores ligados ao ambiente onde plantas e florívoros estão inseridos. Entretanto, só conseguiremos avançar no conhecimento acerca da florivoria acumulando dados sobre níveis de danos e padrões de ocorrência deste antagonismo em várias linhagens de plantas e diferentes contextos ambientais e biogeográficos.

Considerando a falta de conhecimento sobre os padrões gerais que determinam as interações entre flores e florívoros, no **capítulo I** nós resumimos as informações disponíveis até o momento sobre florívoros, através de uma revisão integrativa, que reuniu dados de estudos publicados até 2020 e dados inéditos, coletados em campo através de um protocolo padronizado. Nós comparamos os níveis de danos florais entre plantas tropicais e temperadas, examinamos os efeitos das características funcionais das plantas nos danos florais e exploramos a dinâmica eco-evolutiva das interações flor-florívoro para 180 espécies de plantas, em 64 famílias, distribuídas ao redor do mundo. Além disso, nós identificamos os principais caminhos de pesquisa que ajudarão a preencher lacunas persistentes de conhecimento, fornecemos o primeiro banco de dados global e de acesso livre com informações sobre interações flor-florívoro e propusemos um protocolo padronizado para medições de florívoros.

Com flores de 61 espécies, distribuídas em 20 famílias e coletadas usando o protocolo padronizado para medições de florívoros supracitado, no **capítulo II** nós avaliamos a influência de alguns fatores abióticos (ambientais) e características florais nos níveis de florívoros. O objetivo deste capítulo II foi examinar se as mudanças ambientais causadas pela urbanização influenciam os níveis de florívoros e se os níveis de florívoros são impulsionados por atributos de plantas ligados a urbanização e a características funcionais e filogenéticas de flores e espécies. Neste capítulo nós fornecemos uma das primeiras avaliações de florívoros ao nível da comunidade de espécies, através de uma nova abordagem, no contexto das alterações ambientais que ocorrem devido a urbanização, preenchendo assim uma importante lacuna do conhecimento.

Capítulo I:

Revisiting florivory: an integrative review and global patterns of a neglected interaction

1 **Running title: An integrative review of florivory and florivores**

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4 **Revisiting florivory: an integrative review and global patterns of a neglected**
5 **interaction**

6

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Manuscript published in New Phytologist ([//doi.org/10.1111/nph.17670](https://doi.org/10.1111/nph.17670))

Research review

Revisiting florivory: an integrative review and global patterns of a neglected interaction

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Received: 6 November 2020
Accepted: 2 August 2021

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New Phytologist (2022) 233: 132–144
doi: 10.1111/nph.17670

Key words: floral herbivory, florivores,
florivory, flower damage, flower enemy.

Summary

Florivory is an ancient interaction which has rarely been quantified due to a lack of standardized protocols, thus impairing biogeographical and phylogenetic comparisons. We created a global, continuously updated, open-access database comprising 180 species and 64 families to compare floral damage between tropical and temperate plants, to examine the effects of plant traits on floral damage, and to explore the eco-evolutionary dynamics of flower–florivore interactions. Flower damage is widespread across angiosperms, but was two-fold higher in tropical vs temperate species, suggesting stronger fitness impacts in the tropics. Flowers were mostly damaged by chewers, but neither flower color nor symmetry explained differences in florivory. Herbivory and florivory levels were positively correlated within species, even though the richness of the florivore community does not affect florivory levels. We show that florivory impacts plant fitness via multiple pathways and that ignoring this interaction makes it more difficult to obtain a broad understanding of the ecology and evolution of angiosperms. Finally, we propose a standardized protocol for florivory measurements, and identify key research avenues that will help fill persistent knowledge gaps. Florivory is expected to be a central research topic in an epoch characterized by widespread decreases in insect populations that comprise both pollinators and florivores.

Introduction

The rise and diversification of angiosperms is one of the most dramatic yet controversial events in the evolution of plant life on Earth (Ramírez-Barahona *et al.*, 2020). The dominance of pollinator-driven selection in the angiosperm evolution literature largely ignores the key roles natural enemies play in the evolution of flower traits (Galen, 1999a,b; Moreira *et al.*, 2019; Ramos & Schiestl, 2019; Roddy *et al.*, 2021). For example, the fact that insect florivory in the early Cretaceous is indistinguishable from modern day patterns (Xiao *et al.*, 2021) suggests that this ancient interaction potentially affects individual fitness, with potential consequences for plant demography and trait evolution. Florivory, broadly defined as any type of damage to flowers or flower parts that

involves tissue consumption by vertebrates or invertebrates (McCall & Irwin, 2006; Box 1; Supporting Information Fig. S1), has been traditionally understood as a form of herbivory; however, current knowledge suggests that florivory lies at the intersection of herbivory and pollination (McCall & Irwin, 2006).

Florivory potentially impacts plant fitness in multiple ways including changes in resource sinks, floral display, nectar production and/or interactions with other organisms, such as predators (Knauer *et al.*, 2018), leaf herbivores (Rusman *et al.*, 2019) and pollinators (Moreira *et al.*, 2019). The effects of florivory on plant fitness depend upon several factors, including the extent of damage (amount of floral area removed), florivore identity, diet breadth and feeding guild. However, we still lack an appropriate understanding of the degree of flower consumption that will affect plant fitness,

43 **Summary**

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45 Florivory is an ancient interaction which has rarely been quantified owing to a lack of standardized
46 protocols, thus impairing biogeographical and phylogenetic comparisons. We created a global,
47 continuously-updated, open-access database comprising 180 species and 64 families to compare
48 floral damage between tropical and temperate plants, to examine the effects of plant traits on floral
49 damage, and to explore the eco-evolutionary dynamics of flower-florivore interactions. Flower
50 damage is widespread across angiosperms but was two-fold higher in tropical vs. temperate
51 species, suggesting stronger fitness impacts in the tropics. Flowers were mostly damage by
52 chewers, but neither flower color nor symmetry explained differences in florivory. Herbivory and
53 florivory levels were positively correlated within species even though the richness of the florivore
54 community does not affect florivory levels. We show that florivory impacts plant fitness through
55 multiple pathways and that ignoring this interaction hampers a broad understanding on the ecology
56 and evolution of angiosperms. Finally, we propose a standardized protocol for florivory
57 measurements, and identify key research avenues that will help fill persistent knowledge gaps.
58 Florivory is expected to be a central research topic in an epoch characterized by widespread
59 decreases in insect populations that comprise both pollinators and florivores.

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63 ***Keywords***

64 Florivores, flower damage, flower enemy, florivory, floral herbivory

65 **Introduction**

66

67 The rise and diversification of angiosperms is one of the most dramatic, yet controversial
68 events on Earth (Ramírez-Barahona *et al.*, 2020). The dominance of pollinator-driven selection in
69 the angiosperm evolution literature largely ignores the key roles natural enemies play in the
70 evolution of flower traits (Galen, 1999a,b; Moreira *et al.*, 2019; Ramos & Schiestl, 2019; Roddy
71 *et al.*, 2021). For example, the fact that insect florivory in early Cretaceous is indistinguishable
72 from modern day patterns (Xiao *et al.*, 2021) suggests that this ancient interaction potentially
73 affects individual fitness, with consequences for plant demography and trait evolution. Florivory,
74 broadly defined as any type of damage to flowers or flower parts that involves tissue consumption
75 by both vertebrates and invertebrates (McCall & Irwin, 2006; Box 1, Supporting information, Fig.
76 S1), has been traditionally understood as a form of herbivory; but current knowledge suggests that
77 florivory lies at the intersection between herbivory and pollination (McCall & Irwin, 2006).

78 Florivory potentially impacts plant fitness in multiple ways including changes in resource
79 sinks, floral display, nectar production and/or interactions with other organisms, such as predators
80 (Knauer *et al.*, 2018), leaf herbivores (Rusman *et al.*, 2019) and pollinators (Moreira *et al.*, 2019).
81 The effects of florivory on plant fitness depend upon several factors, including the extent of damage
82 (amount of floral area removed), florivore identity, diet breadth and feeding guild. However, we
83 still lack an appropriate understanding on how much flower consumption is too much to affect
84 plant fitness, mainly because the relative contribution of direct vs. indirect effects of florivory is
85 challenging to measure. The magnitude of the indirect effects of flower damage on plant fitness
86 (via shifts in pollinator behavior) are poorly understood (but see Liao *et al.*, 2013; Missagia &
87 Alves, 2017; Ye *et al.*, 2017). Additionally, because florivory levels have been rarely quantified in
88 the field, the effects of florivores on plant fitness cannot yet be extrapolated to different ecological
89 and biogeographic contexts. In a field experiment, McCall (2008) showed that both naturally
90 damaged flowers and artificially damaged flowers set fewer fruit or seeds relative to undamaged
91 control flowers.

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Box 1. Flowers beyond pollinators: the role of flower enemies

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Florivory has been broadly defined as flower consumption by animals or flower herbivores, although the different flower enemies involved in this interaction have different impacts on plant fitness and floral evolution. Nectar robbers and florivores have been the most frequent enemies evaluated, both through natural observations as well as through experiments manipulating flower damage artificially. We compiled definitions of terms regarding flower enemies (Galen, 1999a,b; McCall & Irwin, 2006), bearing in mind the challenges of labelling species behavior in such entangled, intermingled interactions. For example, in some plant families, flower parts are offered as the reward for legitimate pollinators, which are not included in our definition of florivores due to their overall positive outcome for plant fitness (Gottsberger, 1999):

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1. Florivores: animals that cause damage and consume flower buds, flowers and floral rewards, before fruit and seed formation. Included in this category are vertebrates and invertebrates that damage petals, sepals, bracts, pistils, stamens and/or ovules. These florivores consume floral parts or, in some cases, the whole flower. We excluded from this definition, pollen-eating pollinators which consume flower parts but act as legitimate pollinators.

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2. Nectar robbers: Nectar robbers damage flowers, usually without aiming to consume the perianth tissue. They are considered flower enemies, but not florivores.

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2.1. Primary nectar robbers: illegitimate flower visitors who usually pierce flower parts to obtain nectar without exerting any pollen removal or deposition.

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2.2. Secondary nectar robbers: illegitimate flower visitors who use perforations made by previous robbers to obtain nectar, but who exert no pollen removal or deposition functions.

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3. Nectar thieves: animals that use the natural flower opening to obtain nectar, but do not exert any pollen removal or deposition due to incompatible morphology between body and flower. Nectar thieves do not pierce flowers. Together with nectar robbers, nectar thieves are sometimes called nectar larcenists.

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4. Pre-dispersal seed or ovule predators: animals that prey upon seeds or ovules before they are dispersed. The main predators are insect larvae and adults, birds, and mammals. Their impacts on plant fitness vary widely.

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5. Flower parasites: pathogens that are transported to flowers by pollinators or other visitors, such as spores, viruses or bacteria. Infections caused by these parasites can lead to plant diseases and lower nectar production and/or sterile flowers.

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6. Saproflorivory: consumption of flowers already shed by the mother-plant. This is unlikely to have major negative impacts on plant fitness.

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This suggests that florivores may have an underappreciated role in impacting female fitness.

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Thus, evaluating the impacts of florivores on plant fitness needs to extend beyond estimating floral

141 tissue loss and start embracing the multitude of paths by which florivores impact host plant fitness,
142 such as their indirect effects.

143 Florivory resembles herbivory in that it involves the consumption of plant tissues with
144 direct (damaging reproductive structures) and indirect (affecting the attractiveness of damaged
145 flowers to pollinators) negative effects on plant performance (see González-Browne *et al.*, 2016;
146 Soper-Gorden & Adler, 2016; Soper-Gorden *et al.*, 2016). However, florivory also resembles
147 pollination in that flowers are a crucial resource to florivores in the similar ways as they are to
148 pollinators (Genini *et al.*, 2010). Many pollinator-attractive traits also attract floral enemies
149 (Jacobsen & Raguso, 2018; Rusman *et al.*, 2019), establishing a complex mosaic of selective forces
150 shaping flower traits. However, florivory goes beyond these mere resemblances, and stand alone
151 as an interaction linking both herbivory and pollination (Table 1). Florivory also shapes interactions
152 with plant herbivores generating cascading and community-wide effects (Soper-Gorden *et al.*,
153 2016).

154 Distinguishing between foliar and floral herbivores as well as generalist consumers of both
155 flowers and leaves is important due to their different consequences for plant fitness (McCall &
156 Irwin, 2006). Induced plant responses to florivory against generalist insects might be common
157 (McCall & Karban, 2006), and may trigger responses in other plant organs. For example, flower
158 damage by generalist flower eating katydids (Orthoptera) induces mechanical defenses in leaves
159 of *Lantana camara* (Goh *et al.*, 2019). Generalist herbivores that simultaneously feed upon leaves
160 and flowers might cause greater reductions in plant performance due to the combined and direct
161 effects of tissue consumption and the indirect effects of flower consumption on pollinators (Cardel
162 & Koptur, 2010). By accounting for both leaf and flower damage, we depict a more realistic picture,
163 as ecological communities involve the simultaneous occurrence of antagonistic and mutualistic
164 species on a single resource (e.g., leaf or flower) rather than as isolated pairwise mutualistic or
165 antagonistic interactions (Tscharrntke & Hawkins, 2002).

166

167 **Table 1** - The tangled nature of the ecology, evolution and biogeography of widely-studied herbivores and pollinators, and overlooked florivores.
168

Characteristic	Herbivores	Pollinators	Florivores	Key References
Interaction outcome	Parasitism, predation, or commensalism	Mutualism, rarely parasitism	Parasitism, rarely mutualism	Herrera & Pellmyr, 2002; McCall & Irwin, 2006
Impact on host	Negative or neutral	Positive	Unknown	Herrera & Pellmyr, 2002; McCall & Irwin, 2006
Plant organ/resource	Roots, stems, leaves, seeds	Pollen, nectar, oil, more rarely petals	Sepals, tepals, stamens, ovules, nectar	Coley & Barone, 1996; McCall & Irwin, 2006
Frequency of interaction	Year-round	During plant reproduction	During plant reproduction	McCall & Irwin, 2006
Specialization degree	From generalists to specialists	From generalists to specialists	Most likely generalists	Forister <i>et al.</i> , 2015; Wash & Ollerton, 2006; McCall & Irwin, 2006; Caruso <i>et al.</i> , 2019
Selective force	On leaves and flowers	On flowers	On flowers?	Lorange <i>et al.</i> , 2013; Schiestl & Johnson, 2013; Caruso <i>et al.</i> , 2019
Latitudinal gradient	Specialization towards tropics	Specialization towards temperate areas	Unknown	Coley & Barone, 1996; Kozlov <i>et al.</i> , 2015b; Waser & Ollerton, 2006; Keith & Mitchell-Olds, 2017

169 Furthermore, the presence of a third agent can affect the likelihood, type or intensity of
170 interactions within a given pairwise interaction, influencing patterns of trait evolution through
171 diffuse selection (Strauss & Irwin, 2004). Generalist ants – which usually defend plants when
172 herbivores are abundant (Rosumek *et al.*, 2009) – turn into florivores when neither leaves or
173 herbivores are present (Smallegange *et al.*, 2007; Bandeili & Müller, 2010). The impacts of
174 herbivory and florivory on plant fitness can lead to different population, community and
175 ecosystem-wide outcomes (McCall & Irwin, 2006), and it is therefore crucial to start distinguishing
176 between damage caused by both generalist and specialist florivores and herbivores.

177 Another knowledge gap preventing a broader understanding of flower-florivore interactions
178 is the understanding on how plant traits affect florivory levels. The remarkable variation in flower
179 traits is generally attributed to pollinator-mediated selection (van der Kooi & Ollerton, 2020).
180 However, floral traits that originally evolved to protect gametes from predators (Chanderbali *et al.*,
181 2016) may also attract flower enemies (see refs in Roddy *et al.*, 2021), creating a mosaic of
182 complex, multilayered interaction networks (Strauss & Irwin, 2004; Lucas-Barbosa, 2016).
183 Unfortunately, morphological flower traits are underrepresented in plant trait databases. A search
184 into TRY database (www.try-db.org) (Kattge *et al.*, 2020) retrieved only 22 morphological flower
185 traits; other traits such as flower rewards, were much more common (more than 3,500 entries) than
186 traits relevant for florivore attraction, including flower size (Teixido *et al.*, 2011; McCall & Barr,
187 2012; Gélvez-Zúñiga *et al.*, 2018), floral display (Krupnick *et al.*, 1999; Oguro & Sakai, 2015),
188 longevity (McCall & Irwin, 2006; Teixido *et al.*, 2011), secondary metabolites in petals (Jacobsen
189 & Raguso, 2018, Kessler *et al.*, 2019) and in pollen and nectar (Adler, 2000).

190 Changes in plant apparency (Feeny, 1976; Smilanich *et al.*, 2016) involve strategies to
191 reduce damage from herbivores and plants may also escape florivory by making flowers less
192 conspicuous. Flower size, longevity and arrangement (single *vs.* inflorescences) are linked with
193 flower appearance and hence may drive florivory levels. In deciduous plant systems, leaf
194 herbivores are expected to act as florivores when leaves are absent or *vice-versa* (Boaventura *et al.*,
195 under review). For instance, in tropical deciduous species (where dry-season blooming occurs
196 in leafless trees) or in temperate ecosystems (where leaves and flowers are often flushed at different
197 times throughout spring and summer), natural selection may favor phytophagous animals capable
198 of using both resources (leaves and flowers), and extend their foraging opportunities and timings
199 (Boaventura *et al.*, under review). In turn, the optimal defense theory (ODT, McKey, 1974) is also

200 evoked as a mechanistic explanation for the variation in the distribution of chemical defenses across
201 plant tissues and organs depending on their vulnerability to attack and fitness value. The ODT is
202 based on two fundamental assumptions: defenses enhance plant fitness when enemies are present
203 and they are costly, decreasing plant fitness when enemies are absent. Natural selection is therefore
204 expected to enhance defense systems with positive benefit-cost ratios, predicting that the most
205 valuable plant organs should be constitutively defended, as is the case of young leaves and flowers
206 or flower parts (McCall & Fordyce, 2010). In contrast, the plant apparency hypothesis (PAH,
207 Feeny, 1976) suggests variation in secondary chemistry investment between and within plants
208 based on the chance of encounter by generalist and specialist herbivores. Because flowers are
209 ephemeral resources and attack levels might be unpredictable and linked to flower longevity, they
210 might be less apparent and also less defended than their counterparts (leaves) within given
211 individuals or species (e.g., McCall & Fordyce, 2010). Both theories – ODT and the PAH – may
212 explain why strict florivores are rare in nature, as coevolution mediated by secondary chemistry
213 might not be the case in florivore-flower systems as commonly reported in herbivore-leaf systems
214 (e.g., Keith & Mitchell-Olds, 2017).

215 Here, we explore overlooked aspects of florivory by using a global, continuously-updated,
216 open-access database to address some issues at the macroevolutionary scale. In particular, we 1)
217 reviewed the evidence for florivory occurrence and addressed its sources of variation using data
218 published in the literature over the past decades; and 2) collected field data with a standardized
219 protocol to a) compare floral damage between tropical and temperate plants; b) examine the effects
220 of plant traits on floral damage, and c) explore the eco-evolutionary dynamics of flower-florivore
221 interactions. Finally, we propose a standardized protocol for florivory measurements that would
222 provide support to macroecological, biogeographic and evolutionary approaches in complex plant-
223 animal interactions.

224

225 **Material and methods**

226

227 *Global trends of naturally occurring florivory*

228

229 Published data was systematically sampled by searching the keywords (Florivor* OR
230 Florivore* OR "Flower enemy" OR "Flower antagonis*" OR "Floral damage" OR "Flower

231 damage") in the title, abstract and keywords of papers published between 1945 and 2020 in the
232 Web of Science and Scopus. Studies found were individually examined following the PRISMA
233 guidelines (Page *et al.*, 2021) for database construction and transparency. We also surveyed the
234 references from recent reviews (González-Browne *et al.*, 2016; Caruso *et al.*, 2019; Moreira *et al.*,
235 2019) and extracted data on floral damage, defined as any kind of damage to any floral part,
236 including petals, sepals and sexual organs. It was not always possible to separate damage within
237 the reproductive structures, so the damage to stamens, pistils and ovules was collectively labelled
238 as reproductive parts.

239 We classified studies according to 1) region (temperate or tropical), 2) country, 3) season
240 of the year in which data were collected), and 4) plant family, and 5) species. We retrieved the
241 following floral traits: 1) symmetry (zygomorphic or actinomorphic), 2) corolla size, 3) color, and
242 4) flower number and arrangement (single or multiple/inflorescence). In our database flower color
243 is reported as perceived by human eyes but we acknowledge that ideally color categories should
244 be addressed from the perspective of animals that use flowers as resources, as has been done for
245 honey bees (Waser *et al.*, 1996). Florivores were classified according to their 1) Order, 2) guild
246 (chewers, suckers, parasites and pollen feeders) and 3) broad taxonomic group (birds, mammals,
247 insects, other invertebrates, and several groups). Lastly, floral damage was classified according to
248 the response variable used as a *proxy* for florivory: 1) frequency of attacked flower or flower parts
249 and 2) percentage of floral area removed [(corolla damage divided by total corolla area) *100].
250 Data recorded as presence or absence of florivory were not included in our dataset as our main goal
251 was to provide quantitative estimates of florivory levels in natural plant populations. When
252 published studies recorded different florivory measurements, different plant species and/or
253 different locations, each record was considered as an outcome (Supporting information, Table S1).
254 Database is available at datadryad.org (<https://doi.org/10.6084/m9.figshare.15106230>).

255 Florivory levels from published studies are reported separately in the database for each
256 response variable (frequency of attacked flowers and percentage of flower area removed by
257 florivores), but only frequency of attacked flowers had enough replicates to allow statistical
258 analysis using phylogenetically least square models (PGLS). We tested for differences in florivory
259 between regions, florivore guilds and floral traits accounting for phylogenetic relatedness among
260 species, adjusting the expected covariance under a multivariate Brownian motion model
261 (Felsenstein, 1985; Martins & Hansen, 1997). In this model we fitted frequency of flower attack as

262 the response variable, and as fixed effects we included region, flower symmetry, number, and color,
 263 florivore broad taxonomic group and florivore guilds. Plant species was included as a random effect
 264 to account for non-independence among data. The phylogeny tree for the 98 plant species
 265 (Supporting information, Fig S1) was constructed using the package *phytools* (Revell, 2012) in R
 266 studio (R Core Team, 2019) and analyses were done using *nlme* (Pinheiro *et al.*, 2018), *lme* and
 267 *lme4* packages (Bates *et al.*, 2015).

268

269 *Field data - florivory levels and trait-mediated florivory*

270

271 To supplement the database using standardized data on florivory levels under natural
 272 conditions, we sampled flower damage from 36 plant in Mexico, Brazil, and Scotland using a
 273 standardized protocol for data sampling and flower measurements. Given that field data on
 274 florivory levels (percentage of corolla damage) is yet rare, we included as many species as possible
 275 to increase our generalization capacity. Up to 20 adult individuals of each plant species were
 276 randomly selected, marked, and georeferenced in the field. The canopy of every selected individual
 277 was sampled in order to haphazardly collect approximately 25 flowers, regardless of signs of
 278 damage by florivores (Supporting information, Fig. S2). For plant species with inflorescences, five
 279 inflorescences were sampled per individual and all flowers were examined for petal or sepal
 280 damage in the laboratory. Due to the ephemeral nature of flowers and to the logistics of digitizing
 281 fresh flowers in the field, we focused our study in floral damage on petals and sepals (or tepals)
 282 and direct damage to sexual organs was not evaluated in these sampled flowers.

283 Flower petals and or sepals were digitized for the 36 plant species sampled (Supporting
 284 information, Table S2) and florivory was estimated according to damage classes by a single
 285 researcher using digital images, a method commonly used for herbivory studies (Dirzo &
 286 Dominguez, 1995). Classes of removed floral area were assigned as: (0)- intact flower, (1) 0.01 to
 287 1%, of floral area removed by florivores, (2) between 1 and 5%, (3) between 5 and 25%, (4)
 288 between 25 and 50%, (5) between 50 and 75% and (6) between 75 and 100% of flower area
 289 removed. A damage index was calculated for each individual, as $IFD = \sum ni(i)/N$, where IFD is
 290 the index of flower damage, “ni” is the number of flowers in category “i” and “N” is the total
 291 number of flowers collected per individual. For 14 of those species (Supporting information, Table
 292 S3), florivory levels were also precisely measured evaluating the ratio of petal area removed by

293 florivores divided by the total corolla or calix area. Each flower (n=3,025) was digitized with a
294 numeric scale and images were processed using the ImageJ® software after calibrating each image
295 to 0.01 mm.

296 We used phylogenetically-adjusted least square models (PGLS) to evaluate the effects of
297 region, floral and plant traits on the variation of florivory levels among related plant species. First,
298 we fitted phylogenetically-adjusted regressions mixed models, where the response variable was the
299 mean level of floral damage, whereas region, flower traits, and plant growth form were considered
300 as explanatory factors. We assumed a Brownian motion model of trait evolution, and all analyses
301 were conducted following the details mentioned above (Supporting information, Figure S3).

302

303 *Eco-evolutionary dynamics of herbivory and florivory*

304

305 To illustrate the spatio-temporal complexity of florivory, we used standardized data from
306 four sympatric, congeneric trees with leafless-blooming during the dry season (Bignoniaceae).
307 Four deciduous species with different blooming lengths (*Handroanthus heptaphyllus*, *H.*
308 *chrysotrichus*, *H. ochraceus* and *Tabebuia roseo-alba*) were examined to evaluate florivory and
309 herbivory levels, and the community of insects associated to flowers or to leaves. Deciduousness
310 promotes spatio-temporal separation of the type of resource available to insects, as either leaves
311 (during summer and spring) or flowers (during fall and winter). In 2017, we monitored flower
312 phenology from beginning to end of by recording the number of days individual plants had open
313 flowers in 60 individual trees [*H. heptaphyllus* (n=17), *H. chrysotrichus*, (n=15), *H. ochraceus*
314 (n=15), *T. roseoalba* (n=13)] in southeastern Brazil (21° 08' 09" S 44° 15' 43" W). We also
315 collected insects associated to either leaves (in the wet season of 2017) or flowers (in the dry season
316 of 2017) using an entomological umbrella (Gullan & Craston, 2017) and a trimmer with a
317 telescopic handle attached to an entomological net. We carried out five beats in the canopy of each
318 individual tree with the entomological umbrella and all the insects that fell into the trap were
319 collected. We also inspected each plant for 10 minutes, collecting insects on it with tweezers. All
320 captured insects were identified to the lowest possible taxonomic group and the resource used
321 (leaves or flowers) was recorded. Simultaneously, twenty leaves and five inflorescences were
322 collected from each plant. Herbivory and florivory levels were calculated for each individual plant
323 as mean percentage of floral or leaf area removed, using leaves or flowers as replicates per

324 individual plant, and correcting for leaf or flower size. Leaf or flower area removed was calculated
325 using digital images processed on ImageJ® after calibrating each image to 0.01 mm (Mendes *et*
326 *al.*, 2021).

327 Because these four plant species are phylogenetically related, we also applied PGLS models
328 to test for differences in insect communities, herbivory and florivory levels among species. The
329 response variables were the level of floral damage, the level of leaf damage and species richness
330 of florivores or herbivores, whereas plant species and number of days plants were blooming were
331 included as explanatory factors. Phylogenetic relatedness was calculated using *phytools* package
332 (Revell, 2012) and included in the model as a random variable assuming Brownian covariance
333 correlation. All analyses were conducted using the *nmle* and *lme4* packages on R software.
334 Goodness of model fit was tested using a maximum likelihood (ML) with link and variance
335 functions.

336

337 *A standardized protocol for measuring florivory*

338

339 Standardized measures of florivory are needed to allow us to address and discuss the effects
340 of florivory on plant fitness and to test predictions on flower-florivore interactions at the
341 macroevolutionary scale, as has been lengthily done for herbivores (Dyer *et al.*, 2007; Turcotte *et*
342 *al.*, 2014; Kozlov *et al.*, 2015a,b; Zhang *et al.*, 2016; Galmán *et al.*, 2018). The available data are
343 yet insufficient to rigorously test predictions at large scales and to be included into meta-analytic
344 frameworks. Having standardized measurements of florivory and flower traits will allow us to run
345 analyses at the macroecological scale and to further identify, for example, the functional
346 significance of flower defensive traits. The results of our review of published data (see below)
347 indicated that frequency of attacked flowers is the most common variable used to quantify
348 florivory. The results of our field data (see below) showed that it is not always possible to assess
349 the levels of florivory using the same methods used to evaluate leaf herbivory, such as categorical
350 measurements of floral area removed. Therefore, we propose a standardized protocol that includes
351 easy-to-measure steps that can be applied to species with different floral traits (Supporting
352 information, Methods S1), and a method to precisely quantify the amount of corolla area removed
353 due to florivore damage.

354

355 **Results**

356

357 *Global trends of naturally occurring florivory*

358

359 Eighty studies on florivory were published between 1976 and 2019, generating 458
360 independent comparisons regarding flower damage for 144 plant species from 47 families, in both
361 temperate and tropical areas (Supporting Information, Fig. S4). Flower damage was inflicted by
362 animals from different lineages (Supporting Information, Fig. S5, Fig. S6).

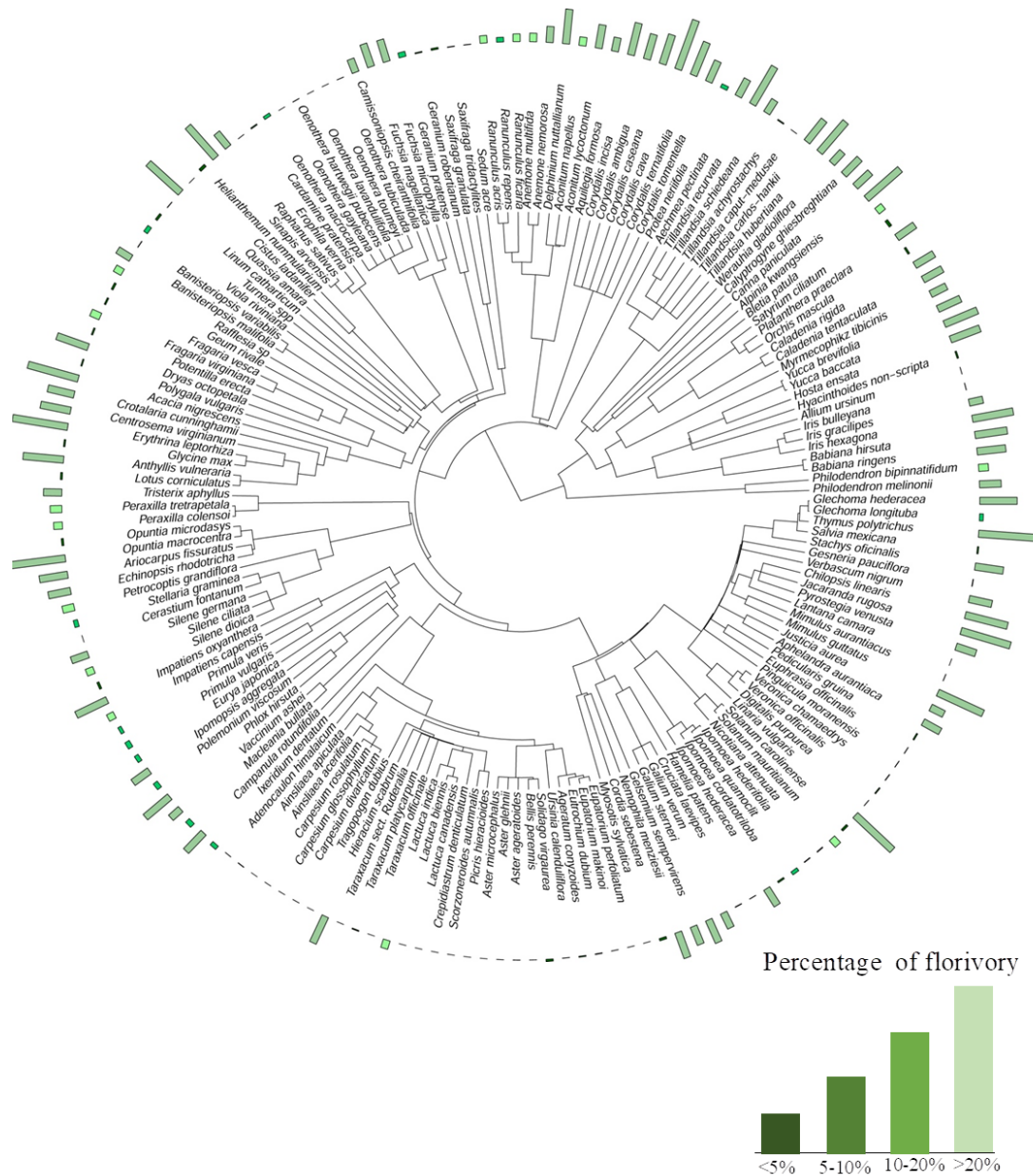
363 Florivory was widespread among plant lineages (Fig. 1) and levels of attack are variable at
364 both plant and flower scales. Frequency of attacked flowers was the most common response
365 variable used to estimate florivory (85% of cases), with values ranging from 0.2 to 97%. Frequency
366 of attacked flowers was significantly higher in the tropics ($F_{1,293}=8.35$, $P=0.004$), in single flowers
367 compared to inflorescences ($F_{1,294}=4.076$, $P=0.007$), and higher for flowers attacked by chewers
368 compared to other guilds and for mammals or birds compared to insects and other invertebrates
369 ($F_{7,293}=9.371$, $P<0.0001$) (Fig. 2). Neither flower color nor flower symmetry influenced levels of
370 florivore attack. We did not detect a phylogenetical significant signal in any of the variables
371 evaluated to modulate florivore attack (Supporting information, Table S4).

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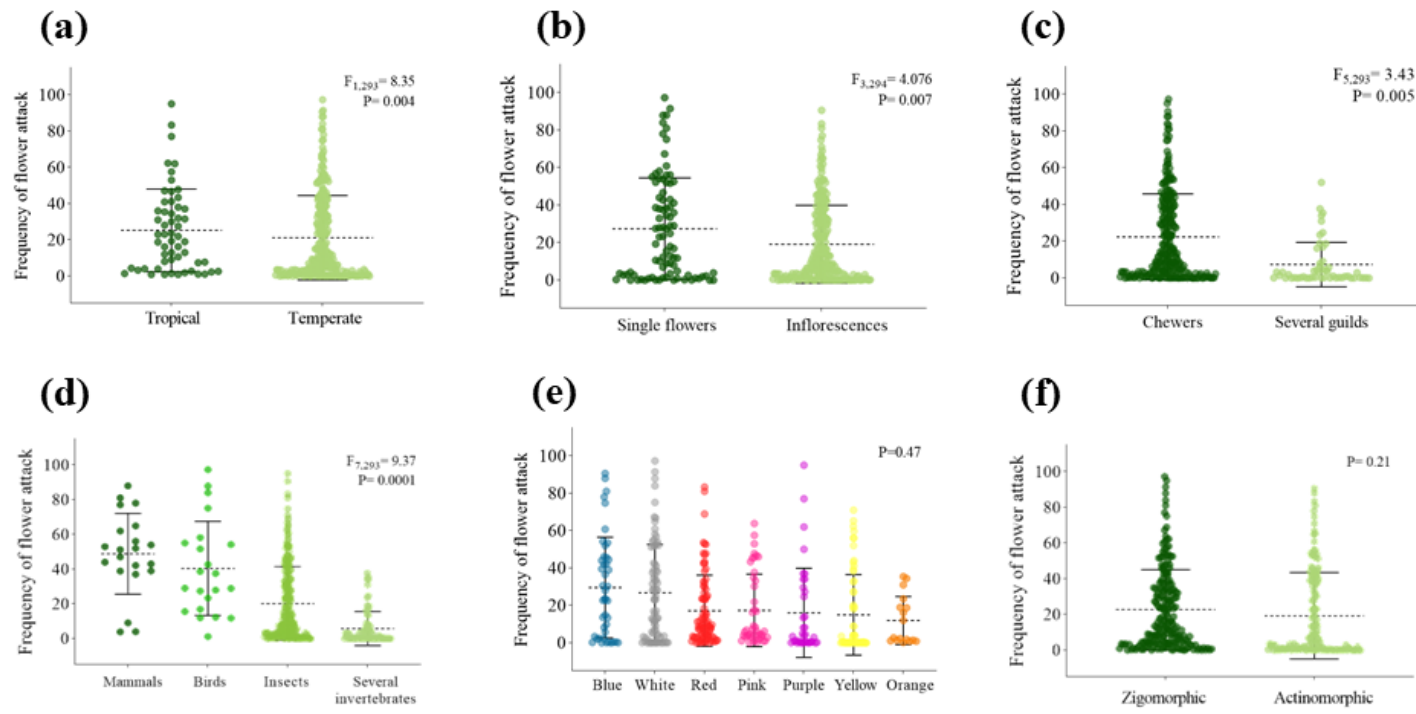
378 **Figure 1.** The phylogenetic distribution of floral damage in the global florivory database mapped onto the
 379 phylogeny of 144 angiosperms. The phylogeny was reconstructed using Qian & Jin (2016) database. See
 380 Supporting Information Table S3 for detailed species name, family and florivory levels.

381

382

383 The percentage of floral area removed was measured in 15% of florivory studies. Average
 384 floral area removed by florivores in attacked flowers was 7.21% (± 1.53 SE, $n=66$). The
 385 phylogenetic distribution of florivory indicates that flower damage is widespread and variable

386 across angiosperm lineages (Fig. 1), but two times higher in the tropics than in temperate regions
387 (10.1 % \pm 2.3 SE for tropical flowers vs. 4.56% \pm 0.97 SE for temperate flowers; $F_{1,18}=4.71$,
388 $P=0.003$). Neither flower symmetry, flower number or florivore guild had a significant effect on
389 levels of damage on flowers. Our assessment on the distribution of floral damage clearly indicates
390 that florivory is not homogeneously distributed across lineages, ecosystems or functional groups
391 despite the phylogenetic and biogeographic biases (Supporting information, Fig. S4), opening new
392 research avenues on the evolutionary ecology of florivory.



393

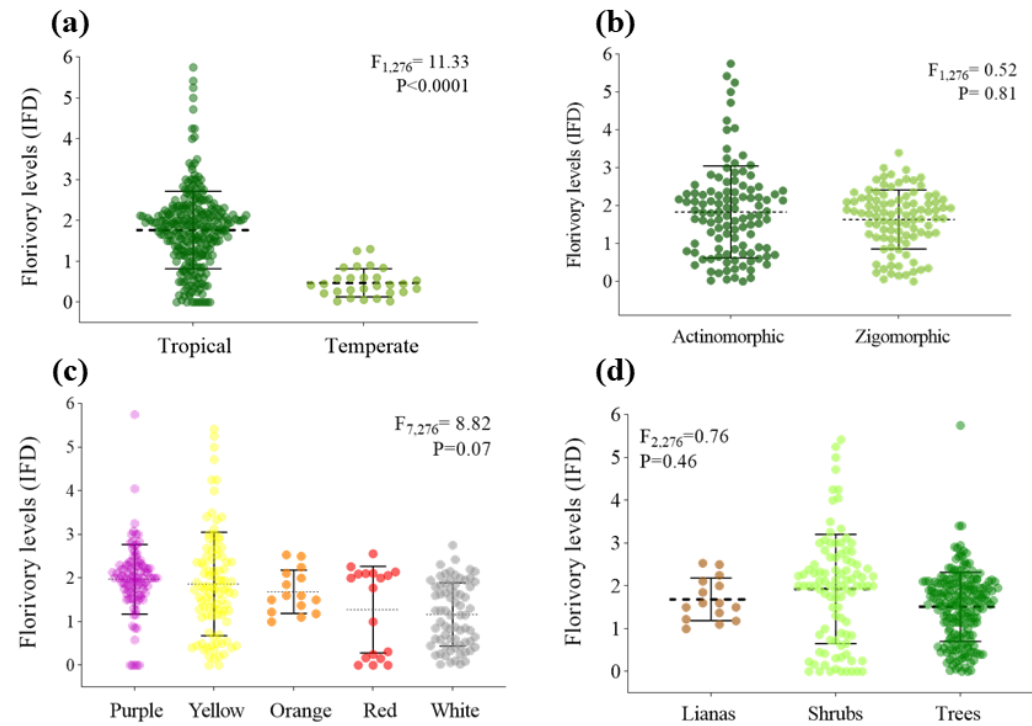
394 **Figure 2.** Comparison of florivory levels (measured as frequency of attacked flowers) according to a) region, b) flower color, c) florivore guild, d)
 395 flower arrangement, e) florivore group, and f) flower symmetry for 98 plant species. Data were extracted from previously published studies between
 396 1945-2020. The line indicates the mean, and the error bars denote SD. For more details see Supporting Information Table S1.

397 *Field data - Florivory levels and trait-mediated florivory*

398

399 Most of our field sampling was conducted for species blooming in spring (28.9%) and
400 summer (31.4%), with actinomorphic flowers (59.4%), arranged into inflorescences (87%). Almost
401 one third of the species evaluated had yellow flowers (31.1%), followed by purple (25.6%) and
402 white flowers (15.7%; Supporting information, Table S2).

403 Our field data showed that, on average, a quarter of all flowers per individual experience
404 florivory (frequency of occurrence of damaged flowers: $24.3 \pm 1.1\%$ SE). Consistently two-times
405 higher florivory levels were reported in tropical plants (Fig. 3). Flower symmetry, color and growth
406 form did not significantly influence the levels of floral damage on these 36 plant species. Florivory
407 levels estimated via the IFD showed an average of 1.65 (± 0.05 SE) of flower area removal, which
408 is equivalent to having up to 5% of petal area consumed by florivores (Supporting information,
409 Methods S1). Based on precise estimates of the floral area removed by florivores using digital
410 images of each flower (n= 14 woody species, 3,025 flowers measured), the floral area lost was
411 7.6% (± 0.21 SE) and this pattern was not influenced by plant phylogeny (Supporting information,
412 Table S5).



413

414

415 **Figure 3.** Florivory levels (percentage of floral area removed) measured by the Index of Floral Damage (IFD) for the 36 sampled plant species
 416 according to a) region, b) growth form, c) flower color and d) flower symmetry. Plants and flowers were sampled in the field using a standardized
 417 protocol (detailed in Supporting Information). Classes of removed floral area were assigned as: (0)- intact flower, (1) 0.01 to 1% of floral area
 418 damage, (2) between 1 and 5%, (3) between 5 and 25%, (4) between 25 and 50%, (5) between 50 and 75% and (6) between 75 and 100% of flower
 419 area damage. The line indicates the mean and the error bars denote SD.

420 *Eco-evolutionary dynamics of herbivory and florivory*

421

422 Florivory levels did not significantly vary among the four species of Bignoniaceae
423 (Supporting information, Fig. S7), but herbivory and florivory were positively correlated within
424 each species, indicating that plants that were more attacked by herbivores were also more attacked
425 by florivores. Interspecific levels of herbivory varied between 2.6% (± 0.2) to 5.95% (± 1.02 ;
426 Supporting information, Fig. S7), and flower area removed ranged from 4.29% (± 1.13) to 7.3% (\pm
427 1.87). A total of 78 species of florivores and 99 species of herbivores were sampled in these host
428 plants and similarity between herbivore and florivore communities was low (Supporting
429 information, Fig. S8). However, there was no relationship between richness of florivores or
430 herbivores and damage level on flowers ($F_{1,58}=1.76$, $P=0.190$) or leaves ($F_{1,58}=2.206$, $P=0.143$).
431 Longer blooming periods of flowers, however, were positively correlated ($r=0.543$) to a greater
432 diversity of florivores ($F_{1,58}=24.29$, $P<0.001$; Supporting information, Fig. S8).

433

434 *A standardized protocol for measuring florivory*

435

436 In brief, our protocol called for field sampling of 20 fresh flowers from at least five
437 individuals of five different species from natural areas. Flowers should be digitized with a scale for
438 reference, measured on image processing software to precisely measure floral area removed (ratio
439 between flower tissue removed and total flower tissue area). Flower traits should be registered
440 along with data from each location. This protocol is already in use (Mendes *et al.*, 2021) and is
441 intended to facilitate sampling of florivory measures across biogeographic regions and plant
442 lineages, which will improve representativeness in our continuously-updated database. We provide
443 data for more than 13,500 flowers in 36 plant species from 17 families across sites spanning over
444 38 degrees of latitude.

445

446 **Discussion**

447

448 *Biogeography of flower-florivore interactions*

449

450 Our global survey using both published and standardized field data collection suggests
451 florivory is more common than previously conceived (see also Xiao *et al.*, 2021), and provides
452 novel insights about the biogeographical, phylogenetic and ecological distribution of floral tissue
453 consumption by florivores both at the plant and flower scales. Our results for published data show
454 that patterns of damage to flowers were influenced by the florivore guild, flower arrangement and
455 florivory metric accounted for. We found higher levels of florivory on tropical plants caused mainly
456 by petal-chewing caterpillars, but higher frequency of attack on flowers when florivores were birds
457 and mammals and flowers were singly arranged. The higher intensity of florivory in tropical plants
458 - observed in the published data and also in our field sampled species - suggests a potential higher
459 impact of floral damage on plant performance in the tropics and sheds light into the long-standing
460 discussion on latitudinal gradients in insect-plant interactions (Forister *et al.*, 2015; Roslin *et al.*,
461 2017). Surprisingly, data gathered from the published literature and our field sampling both show
462 that global florivory levels are almost twice as high as global herbivory levels reported (Kozlov *et*
463 *al.*, 2015a), with higher impact on tropical plants compared to temperate ones. Tropical species
464 might accumulate more specialized communities of herbivores, florivores and also natural enemies
465 of both guilds (Basset *et al.*, 2012), and higher levels of florivory on tropical plants might indicate
466 higher specialization of, for example, Lepidoptera larvae on this particular resource. Caterpillars
467 are amongst the most common chewing herbivores in the tropics (Rinker & Lowman, 2004), and
468 we suggest some species might be specialized in flowers as is the case of several species occurring
469 in Asteraceae flower heads (Oguro & Sakai, 2014; Nobre *et al.*, 2016).

470 Florivory is inconsistently reported in the literature, but frequency of damaged flowers in
471 plants or plant populations was the most commonly used metric, followed by the level of florivory
472 on flowers or flower parts, measured as percentage of area removed by florivores. Each metric
473 provides different information on florivory levels, including either prevalence of damage or
474 damage level, and therefore are non-comparable measurements of florivory. In contrast,
475 sophisticated and widely-accepted methods and metrics are used to estimate herbivory levels on
476 plants (e.g., Machado *et al.*, 2016; Getman-Pickering *et al.*, 2020). The IFD (Index of Floral
477 Damage) analogous to the IHD (index of herbivory damage – the most widespread method to
478 quickly evaluate levels of leaf herbivory in the field) – seems to flatten and obscure the estimates
479 of flower area removed by florivores. Data from our field sampling provides empirical evidence of
480 this, showing that flowers classed in the same IFD category in reality had very different levels of

481 damage when measured quantitatively through image processing (Supporting information, Figure
482 S9). This result strongly suggests the need for switching from semi-quantitative to quantitative
483 measures to allow the identification of thresholds of tissue loss that can be tolerated without
484 hindering plant fitness, as done in parallel of studies on leaf-herbivore interactions (Kozlov &
485 Zvereva, 2017). Flower morphology and structure strongly differ from the usually flat-shaped
486 leaves, posing additional challenges to estimate florivory levels, and impairing quick assessment
487 of florivory in the field as proposed by the IFD. Our results indicate that the quantification of floral
488 area removed with the aid of digital images and precise measurements of petal loss, corrected by
489 flower size, will contribute to a deeper understanding of the effects of flower antagonists on plants
490 and will fill this gap we identified in the published literature.

491

492 *Trait-based flower-florivore interactions*

493

494 Besides regional and latitudinal variation in florivory levels, another promising research
495 avenue is the examination of flower functional traits that influence flower damage, either via
496 attraction or repellence of florivores (Table 2). Species floral traits and time of flowering have been
497 suggested as major drivers of pollinator attraction (Albor *et al.*, 2020) but, despite some recent
498 efforts (e.g., Paterno *et al.*, 2020) flower traits are rarely sampled in global plant trait databases and
499 have rarely been linked to florivory levels experienced by plants. Our results showed that some
500 flower and plant traits influenced the frequency of attack on flowers but also the levels of florivory
501 experienced by plants.

502

503 **Table 2** - Floral traits suggested to have effects on florivory levels on plants.

504

Flower trait	Putative effects on florivory
Flower colour	Floral pigments are often linked to biosynthetic pathways of metabolites that can have defensive properties and also support the idea that florivores, just like pollinators, can use flower colour as cues for locating resources.
Flower longevity	Longer-lived flowers often sustain higher levels of florivory, except when they are better protected by either constitutive or induced defenses. The relationships between defenses and flower longevity in terms of costs and resource allocation remain poorly explored.
Flower size	Smaller flowers might escape florivory due to lower apparency compared to larger flowers, which are conspicuous and attractive to both pollinators and florivores. Although some data support a negative relationship between flower size and floral damage, data from tropical plants are still lacking.
Flower arrangement /number	The number of open flowers and the arrangement of flowers (single flowers or inflorescences) can potentially make flowers more or less conspicuous for florivores and levels of florivory may increase with the increase of the floral display.
Floral symmetry	Floral symmetry has been widely studied in terms of pollinator attraction or repulsion, but is still unclear whether floral symmetry (actinomorphic or zygomorphic) influences levels of damage by floral enemies.
Floral phenology	Plants can delay or advance flowering phenology to reduce overlapping with abundance of florivores. It is not clear, however, whether this strategy would be effective against generalist florivores.
Floral scent	Flowers that produce volatile organic compounds used as cues by florivores for oviposition and/or feeding might suffer higher floral

damage. On the other hand, floral scents might be used by natural enemies of florivores, thereby reducing florivory levels.

Plant growth form

Plant growth form can make flowers more or less apparent to florivores, so that herbs, shrubs or lianas may experience lower levels of florivory compared to trees, where flowers might be less accessible. This functional trait has already been extensively studied for leaf herbivory, but studies that have investigated its relationship with florivory are still scarce.

505

506

507 Flowers tend to be the most diversely pigmented plant structures, displaying a wide range
508 and combination of colors often both between and within species (Willmer, 2011). Although plant
509 pigments have been generally interpreted as means to attract mutualists, they are often linked to
510 biosynthetic pathways of metabolites that can have defensive properties and support the idea that
511 florivores, similarly to pollinators, can use flower color as a cue for either locating resources or for
512 flower consumption. Surprisingly, in our current analyses flower color did not influence the levels
513 of petal damage, both for the published and sampled field data, even though floral color has been
514 shown to drive the foraging decisions of several pollinators (Reverté *et al.*, 2016), and florivores
515 have been flagged as potential drivers influencing flower color polymorphisms (McCall *et al.*,
516 2013). In species with flower color polymorphism where florivory has been assessed, florivores
517 prefer certain color morphs over others (McCall *et al.*, 2013). However, the evidence is still scarce
518 and based on single study cases in experimental settings. We do wonder to what extent this lack of
519 effect in our database is reflecting a general biological pattern or is the result of lack of power to
520 describe a highly variable and complex process, given the limited data on both florivory levels and
521 floral traits. Documenting flower color and florivory levels in wider censuses, along with
522 experiments testing florivore preferences in controlled experimental settings would help fill
523 building a broader picture on the role of florivores in shaping flower color.

524 Flowers arranged into inflorescences experienced overall lower florivory than single
525 flowers and these results are consistent with the hypothesis that escape may be a mechanism to
526 reduce florivory levels, as documented for herbivory. This underscores similar patterns in resource

527 usage by herbivores and florivores. The plant apparency hypothesis (PAH; Feeny, 1976) suggests
528 that conspicuous plants, such as trees, are 'bound to be found' (Smilanich *et al.*, 2016) and have
529 evolved higher concentrations of chemical and/or physical defenses against herbivores due to
530 greater tissue longevity and exposure. We suggest that the PAH might also apply to flowers that
531 are less conspicuous when singly arranged and therefore less 'bound to be found' by florivores,
532 compared to those bundled in dense resource patches as is the case for inflorescences. The
533 evaluation of the frequency of attack in flowers arranged into inflorescences and the levels of
534 florivory experienced by single flowers within an inflorescence will help to understand how flower
535 arrangement, time of blooming, and longevity influence florivore choice.

536 Flower size is another floral trait that influences mutualistic and antagonistic interactions.
537 Larger-flowered species (or populations within species) may be more attacked following the PAH
538 (Feeny, 1976), although this is contingent upon particular ecological and/or phylogenetic contexts.
539 Thus, if populations with larger flowers suffer greater florivory levels, florivores may ultimately
540 act as a selective pressure to reduce flower size (McCall & Irwin, 2006; Teixido *et al.*,
541 2011). Flower size, however, was not commonly or precisely reported in the published studies
542 compiled in our database, appearing in less than a quarter of the studies that described flowers
543 consumed by florivores.

544 The understanding of flower color, arrangement, size, and flower secondary compounds
545 and their effects on both herbivores and florivores is crucial to evaluate the role of generalist plant
546 consumers and their effects as drivers of selection and flower evolution. We here provide a first
547 list of general florivore-related flower traits (Table 2) hoping it will inspire the collection and
548 inclusion of such data in global databases in order to facilitate florivory research in the near future.
549 Overall, if one wishes to understand the effects of multiple and simultaneous forces on floral
550 attraction to both florivores and pollinators, then floral traits should also be included in such
551 studies. Moreover, patterns of co-variance between floral traits and florivory across multiple
552 species using phylogenetic comparative approaches – as done here – are required for a better
553 understanding of the evolution of plant-florivore interactions.

554

555

556

557

558 *Eco-evolutionary dynamics of flower-florivore interactions*

559

560 For almost 90% of the animal-pollinated flowering plants (Ollerton *et al.*, 2011), complex,
561 multispecies interactions with floral enemies and herbivores often result in conflicting selection on
562 plant traits (Jacobsen & Raguso, 2018) including floral display, phenology, flower size, color,
563 morphology, nectar quality and fragrance (Galen, 1999b; Caruso *et al.*, 2010; Agren *et al.*, 2013;
564 Ramos & Schiestl, 2019). Our empirical data is in line with previous studies indicating that leaf
565 herbivores might feed on flowers of the same plant species (e.g., McCall, 2008; McCall *et al.*,
566 2018). Our field data using four related Bignoniaceae species as a study system showed correlated
567 levels of florivory and herbivory within species, but low similarity in the communities of insect
568 florivores and herbivores (Fig. S8). Ants were important components of both communities, acting
569 as florivores in leafless plants, and as predators of herbivores when plants were flushing new leaves
570 and associated herbivores were abundant. Although greater florivore richness did not translate into
571 greater florivory levels for these plant species, florivores were more diverse and abundant in plants
572 with longer blooming periods. Florivory was 2-fold higher than to herbivory for species with the
573 shortest blooming periods (5-7 days). The highest similarity between florivore and herbivore
574 communities was also found for the species that had the longest blooming (~30 days), indicating
575 that plant phenology, flower apparency and phenophase duration may also modulate the levels of
576 flower damage (Albor *et al.*, 2020). Temporal segregation of flowers and leaves in these deciduous-
577 blooming species might be driving the dissimilarity of insects based upon different resources, but
578 we suggest that the determination of insect identity as well as resource use over time are crucial
579 for future studies of florivore-plant interactions.

580 Florivory impacts on plants have to be addressed beyond the effects of floral damage on
581 pollinator attraction and we suggest florivory should be considered as a 'stand-alone' interaction in
582 studies involving flowers and animals. It has the potential to directly and indirectly influence plant
583 fitness via the direct consumption of costly tissues and through impacts on plant attractiveness to
584 pollinators (e.g., abundance, number and duration of visits), and jointly altering fruit and seed
585 production (Moreira *et al.*, 2019). As most research has focused on the positive (mutualistic)
586 animal-flower interactions, negative interactions between animals and flowers, like florivory, have
587 been overlooked or evaluated out of the context of antagonistic interactions. This knowledge gap
588 prevents a better understanding of the multi-layered interaction networks (Genini *et al.*, 2010;

589 Pilosof *et al.*, 2017), hinders its applied facets for mitigating economic losses associated with
590 flower destruction by animals (Kelley *et al.*, 2001), and, because of the inherent importance of
591 flowers in plant fitness, hampers a comprehensive understanding of plant evolution (Chalker-Scott,
592 1999; Tanaka & Brugliera, 2007).

593 We revisited florivory indicating global levels of floral damage, the identity of florivores
594 and a first list of pivotal florivore-related flower traits that influenced floral damage, hoping it will
595 inspire the sampling and incorporation of these key elements into global databases. We hope that
596 our protocol for standardized measurements of florivory, coupled with the examination of flower
597 traits that mediate damage incidence and level would allow the generation of robust, realistic and
598 comparable estimates on the amount of floral tissue removed by to florivores and its impacts on
599 plant fitness.

600

601 **Conclusion**

602 Our study enhances the understanding of the role of floral antagonisms on plant fitness
603 (McCall, 2008) and on flower form and function (Roddy *et al.*, 2021). Multiple ecological factors
604 driving plant fitness and flower traits shape ecological communities and macroevolutionary
605 patterns. A broad understanding of these factors should extend beyond the role of pollinators and
606 include florivores as a major driver of the ecology and evolution of angiosperms. Insects are critical
607 components of communities and ecosystems, providing important ecosystem services such as
608 pollination and decomposition (Elizalde *et al.*, 2020). Many insect taxa act as both pollinators and
609 florivores, but the consequences of their global population declines (van Klink *et al.*, 2020) in
610 flower trait evolution remain largely unexplored. Incorporating florivores into complex multi-
611 layered interactions networks on both natural and modified landscapes will provide a more realistic
612 picture on their role in shaping plant traits, plant demography, community assembly and
613 macroevolutionary patterns (see Burin *et al.*, 2021, Xiao *et al.*, 2021). We look forward to a time
614 when knowledge on florivory is comparable to that of herbivory and pollination, so we can have a
615 broader and realistic understanding of the relative role of phytophagous animals in shaping plant
616 traits.

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619

620 Acknowledgements

621 We thank UFMG, UFSJ, CSEC, PPG-ECMVS and CAPES (Financial code 001) for
622 continuous support. TC and FAOS received grants from CNPq (313007/2020-9) and FAPEMIG.
623 We thank D Negreiros for assistance with Figure 1 and B Soares for assistance with PGLS analyses.
624 Data collection by N Villamil was funded with a Davis Trust, Edinburgh University research grant.
625 This study was partially financed by the Coordenação de Aperfeiçoamento de Pessoal de Nível
626 Superior – Brasil (CAPES) – Finance Code 001. We thank the excellent comments by three
627 anonymous reviewers which significantly improved the previous versions of this paper.

628

629 Author contribution

630 Design of the research (TC, MGB, FAOS), performance of the research (all authors), data
631 collection (all authors), data analysis and interpretation (led by MGB with contributions from all
632 authors), manuscript writing and editing (MGB and TC with input from all authors).

633

634

635 Data availability statement

636 All data that support our findings are available as Supporting Information, and the database
637 has been deposited in Figshare (<https://doi.org/10.6084/m9.figshare.15106230>).

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652 **References**

- 653 **Adler LS. 2000.** The ecological significance of toxic nectar. *Oikos* **91**:409-420.
- 654 **Agren J, Hellström F, Toräng P, Ehrlén J. 2013.** Mutualists and antagonists drive among-
655 population variation in selection and evolution of floral display in a perennial herb. *Proceedings of*
656 *the National Academy of Sciences* **110**: 18202– 18207.
- 657 **Albor C, Arceo-Gómez G, Parra-Tabla V. 2020.** Integrating floral trait and flowering time
658 distribution patterns help reveal a more dynamic nature of co-flowering community assembly
659 processes. *Journal of Ecology*, **108**: 2221– 2231.
- 660 **Bandeili B, Müller C. 2010.** Florivory versus folivory – adaptiveness of flower feeding.
661 *Naturwissenschaften*, **97**:79-88.
662
- 663 **Basset Y, Cizek L, Cuénoud P, Didham RK, Guilhaumo F, Missa O, et al., 2012.** Arthropod
664 diversity in a tropical forest. *Science*, **338**:1481-1484.
- 665 **Bates D, Mächler M, Bolker B, Walker S. 2015.** Fitting Linear Mixed-Effects Models Using
666 lme4. *Journal of Statistical Software*, **67**:1–48.
- 667 **Boaventura MG, Pereira CC, Costa FV, Dáttilo W, Cornelissen TG. 2021.** Insect-plant
668 ecological networks in phylogenetically related hosts. *Under review*.
- 669 **Burin G, Guimarães PRJr, Quental TB. 2021.** Macroevolutionary stability predicts interaction
670 patterns of species in seed dispersal networks. *Science* **372**:733-737.
- 671 **Cardel YJ, Koptur S. 2010.** Effects of florivory on the pollination of flowers: An experimental
672 field study with a perennial plant. *International Journal of Plant Sciences* **171**:283-292.
- 673 **Caruso CM, Scott SL, Wray JC, Walsh CA. 2010.** Pollinators, herbivores, and the maintenance
674 of flower color variation: a case study with *Lobelia siphilitica*. *International Journal of Plant*
675 *Sciences* **171**:1020-1028.
- 676 **Caruso CM, Eisen KE, Martin RA, Sletvold N. 2019.** A meta-analysis of the agents of selection
677 on floral traits. *Evolution* **73**: 4–14.
- 678 **Chalker-Scott L. 1999.** Environmental significance of anthocyanins in plant stress responses.
679 *Photochemistry and Photobiology* **70**: 1-9.
- 680 **Chanderbali AS, Berger BA, Howarth DG, Soltis PS, Soltis DE. 2016.** Evolving ideas on the
681 origin and evolution of flowers: new perspectives in the genomic era. *Genetics* **202**: 1255–1265.
- 682 **Coley PD, Barone JA. 1996.** Herbivory and plant defenses in tropical forests. *Annual Review of*

- 683 *Ecology and Systematics* **27**:305–335.
- 684 **Dirzo R, Dominguez C. 1995.** Plant–herbivore interactions in Mesoamerican tropical dry forests.
685 In: Bullock SH, Medina E, Mooney HA, eds. *Seasonally Dry Tropical Forests*. Cambridge:
686 Cambridge University Press, 304–325.
- 687 **Dyer LA, Singer MS, Lill JT, Stireman JO, Gentry GL, Marquis RJ, Ricklefs RE, Greeney**
688 **HF, Wagner DL, Morais HC, et al. 2007.** Host specificity of Lepidoptera in tropical and
689 temperate forests. *Nature* **448**: 696–699.
- 690 **Elizalde L, Arbetman M, Arnan X, Eggleton P, Leal IR, Lescano MN, Saez A, Werenkraut**
691 **V, Pirk GI. 2020.** The ecosystem services provided by social insects: traits, management tools and
692 knowledge gaps. *Biological Reviews* **95** : 1418–1441.
- 693 **Feeny P. 1976.** Plant apparency and chemical defense. In: Wallace JW, Mansell RL, eds.
694 *Biochemical Interaction Between Plants and Insects*. Boston, MA: Springer US, 1–40.
- 695 **Felsenstein J. 1985.** Phylogenies and the Comparative Method. *The American Naturalist*, **125**:1-
696 15.
- 697 **Forister ML, Novotny V, Panorska AK, Baje L, Basset Y, Butterill PT, Cizek L, Coley PD,**
698 **Dem F, Diniz IR, et al. 2015.** The global distribution of diet breadth in insect herbivores.
699 *Proceedings of the National Academy of Sciences* **112**:442-447.
- 700 **Galen C. 1999a.** Flowers and enemies: predation by nectar-thieving ants in relation to variation in
701 floral form of an alpine wildflower, *Polemonium viscosum*. *Oikos* **85**: 426–434.
- 702 **Galen C. 1999b.** Why do flowers vary? The functional ecology of variation in flower size and
703 form within natural plant populations. *Bioscience* **49**: 631–640.
- 704 **Galmán A, Abdala-Roberts L, Zhang S, Berny-Mier JCT T Rasmann S, Moreira X. 2018.** A
705 global analysis of elevational gradients in leaf herbivory and its underlying drivers: effects of plant
706 growth form, leaf habit, and climatic correlates. *Journal of Ecology* **106**: 413– 421.
- 707 **Gélvez-Zúñiga I, Teixido AL, Neves ACO, Fernandes GW. 2018.** Floral antagonists counteract
708 pollinator-mediated selection on attractiveness traits in the hummingbird-pollinated *Collaea*
709 *cipoensis* (Fabaceae). *Biotropica* **50**: 797–804.
- 710 **Genini J, Morellato P, Guimarães P, Olesen J. 2010.** Cheaters in mutualism networks. *Biology*
711 *Letters* **6**: 494–497.
- 712 **Getman-Pickering ZL, Campbell A, Aflitto N, Grele A, Davis J, Ugine TA. 2020.** LeafByte: A
713 mobile application that measures leaf area and herbivory quickly and accurately. *Methods in*
714 *Ecology and Evolution* **11**: 215– 221.

- 715 **Goh CPS, Tan MK, Tan HTW. 2019.** Florivory by a floriphilic katydid, *Phaneroptera brevis*,
716 induces changes in a leaf trait in *Lantana camara*. *Ecological Entomology* **44**: 577-580.
- 717 **González-Browne C, Murúa MM, Navarro L, Medel R. 2016.** Does plant origin influence the
718 fitness impact of flower damage? A meta-analysis. *PlosOne* **11**: e0146437–e0146437.
- 719 **Gottsberger G. 1999.** Pollination and evolution in neotropical Annonaceae. *Plant Species Biology*
720 **14**: 143-152.
- 721 **Gullan PJ, Cranston PS. 2017.** Métodos em entomologia: coleta, preservação, curadoria e
722 identificação. In: Gullan PJ, Cranston PS, eds. *Insetos: fundamentos da entomologia*. Rio de
723 Janeiro, RJ: Roca 742-773.
- 724 **Herrera CM, Pellmyr O. 2002.** Plant-animal interactions: An evolutionary approach. Oxford,
725 Blackwell Science.
- 726 **Jacobsen DJ, Raguso RA. 2018.** Lingering effects of herbivory and plant defenses on pollinators.
727 *Current Biology* **28**: R1164–R1169.
- 728 **Kattge J, Bönisch G, Díaz S, Lavorel S, Prentice IC, Leadley P, Tautenhahn S, Werner GDA,
729 Aakala T, Abedi M, et al. 2020.** TRY plant trait database - enhanced coverage and open access.
730 *Global Change Biology* **26**: 119–188.
- 731 **Keith RA, Mitchell-Olds T. 2017.** Testing the optimal defense hypothesis in nature: variation for
732 glucosinolate profiles within plants. *PlosOne* **21**: e0180971.
- 733 **Kelley KM, Behe B, Biernbaum JA, Poff KL. 2001.** Consumer ratings of edible flower quality,
734 mix, and color. *HortTechnology* **11**: 644-647.
- 735 **Kessler D, Bing J, Haverkamp A, Baldwin IT. 2019.** The defensive function of a pollinator-
736 attracting floral volatile. *Functional Ecology* **33**: 1223– 1232.
- 737 **Knauer A, Bakhtiari M, Schiestl F. 2018.** Crab spiders impact floral-signal evolution indirectly
738 through removal of florivores. *Nature Communications* **9**: 1367.
- 739 **Kozlov M, Lanta V, Zverev V, Zvereva E. 2015a.** Background losses of woody plant foliage to
740 insects show variable relationships with plant functional traits across the globe. *Journal of Ecology*
741 **103**: 1519-1528.
- 742 **Kozlov MV, Lanta V, Zverev V, Zvereva EL. 2015b.** Global patterns in background losses of
743 woody plant foliage to insects. *Global Ecology and Biogeography* **2**: 1126-1135.
- 744 **Kozlov M, Zvereva E. 2017.** Background insect herbivory: impacts, patterns and methodology.
745 *Progress in Botany* **79**: 313-355.

- 746 **Krupnick GA, Weis AE Campbell DR. 1999.** The consequences of floral herbivory for pollinator
747 service to *Isomeris arborea*. *Ecology* **80**: 125-134.
- 748 **Liao K, Gituru RW, Guo YH, Wang QF. 2013.** Effects of floral herbivory on foraging behaviour
749 of bumblebees and female reproductive success in *Pedicularis gruina* (Orobanchaceae). *Flora -*
750 *Morphology, Distribution, Functional Ecology of Plants* **208**: 562– 569.
- 751 **Loranger J, Meyer S, Shipley B, Kattge J, Loranger H, Roscher C, Wirth C, Weisser W.**
752 **2013.** Predicting invertebrate herbivory from plant traits: polycultures show strong nonadditive
753 effects. *Ecology* **94**: 1499–1509.
- 754 **Lucas-Barbosa D. 2016.** Integrating studies on plant–pollinator and plant–herbivore interactions.
755 *Trends in Plant Science* **21**: 125-133.
- 756 **Machado BB, Orue JP, Arruda MS, Santos CV, Sarath DS, Goncalves WN, Silva GG, Pistori**
757 **H, Roel AR, Rodrigues-Jr JF. 2016.** BioLeaf: A professional mobile application to measure foliar
758 damage caused by insect herbivory. *Computers and Electronics in Agriculture* **129**:44–55.
- 759 **Martins EP, & Hansen TF.1997.** Phylogenies and the comparative method: a general approach
760 to incorporating phylogenetic information into the analysis of interspecific data. *The American*
761 *Naturalist*, **149**:646-667.
- 762 **McCall A. 2008.** Florivory affects pollinator visitation and female fitness in *Nemophila menziesii*.
763 *Oecologia* **155**: 729–737.
- 764 **McCall A, Barr CM. 2012.** Why do florivores prefer hermaphrodites over females in *Nemophila*
765 *menziesii* (Boraginaceae)? *Oecologia* **170**: 147–157.
- 766 **McCall A, Case S, Espy K, Adams G, Murphy S. 2018.** Leaf herbivory induces resistance against
767 florivores in *Raphanus sativus*. *Botany* **96**: 337-343.
- 768 **McCall A, Fordyce JA. 2010.** Can optimal defence theory be used to predict the distribution of
769 plant chemical defences? *Journal of Ecology* **98** :985–992.
- 770 **McCall A, Karban R. 2006.** Induced defense in *Nicotiana attenuata* (Solanaceae) fruit and
771 flowers. *Oecologia* **146**:566-71.
- 772 **McCall A, Irwin RE. 2006.** Florivory: the intersection of pollination and herbivory. *Ecology*
773 *Letters* **9**: 1351–1365.
- 774 **McCall A, Murphy S, Venner C, Brown M. 2013.** Florivores prefer white versus pink petal color
775 morphs in wild radish *Raphanus sativus*. *Oecologia* **172**: 189-195.
- 776 **McKey D. 1984.** Interaction of the ant-plant *Leonardoxa africana* (Caesalpiniaceae) with its

- 777 obligate inhabitants in a rainforest in Cameroon. *Biotropica* **16**: 81-99.
- 778 **Mendes GM, Silveira FAO, Oliveira C, Dáttilo W, Guevara R, Ruiz-Guerra B, Boaventura**
 779 **MG, Sershen, Ramdhani S, Phartyal SS, et al. 2021.** How much leaf area do insects eat? A data
 780 set of insect herbivory sampled globally with a standardized protocol. *Ecology* **102**:e03301.
 781 <https://doi.org/10.1002/ecy.3301>.
- 782 **Missagia CCC, Alves MAS. 2017.** Florivory and floral larceny by fly larvae decrease nectar
 783 availability and hummingbird foraging visits at *Heliconia* (Heliconiaceae) flowers. *Biotropica* **49**:
 784 13-17.
- 785 **Moreira X, Castagneyrol B, Abdala-Roberts L, Traveset A. 2019.** A meta-analysis of herbivore
 786 effects on plant attractiveness to pollinators. *Ecology* **100**: e02707.
 787
- 788 **Nobre P, Bergamini LL, Lewinsohn TM, Almeida-Neto JM. 2016.** Host-Plant specialization
 789 mediates the Influence of plant abundance on host use by flower head-feeding insects.
 790 *Environmental Entomology*, **45**:171-177.
 791
- 792 **Oguro M, Sakai S. 2014.** Difference in defense strategy in flower heads and leaves of Asteraceae:
 793 multiple-species approach. *Oecologia* **174**: 227-239.
- 794 **Oguro M, Sakai S. 2015.** Relation between flower head traits and florivory in Asteraceae: a
 795 phylogenetically controlled approach. *American Journal of Botany* **102**: 407-416.
- 796 **Ollerton J, Winfree R, Tarrant S. 2011.** How many flowering plants are pollinated by animals?
 797 *Oikos* **120**: 321–326.
- 798 **Page M J, Moher D, Bossuyt PM, Boutron I, Hoffmann TC, Mulrow CD, Shamseer L,**
 799 **Tetzlaff JM, Akl EA, Brennan SE, et al. 2021.** PRISMA 2020 explanation and elaboration:
 800 updated guidance and exemplars for reporting systematic reviews. *BMJ*, **372**:71.
 801 [doi:10.1136/bmj.n160](https://doi.org/10.1136/bmj.n160).
- 802 **Paterno GB, Silveira LC, Kollmann J, Westoby M, Fonseca CR. 2020.** The maleness of larger
 803 angiosperm flowers. *Proceedings of the National Academy of Sciences* **117**: 10921-10926.
- 804 **Pilosof S, Porter MA, Pascual M, Kéfi S. 2017.** The multilayer nature of ecological networks.
 805 *Nature Ecology & Evolution* **1**: 0101.
- 806 **Pinheiro J, Bates D, DebRoy S, Sarkar D, R Core Team. 2018.** nlme: linear and nonlinear mixed
 807 effects models. *R package* version 3.1-137. //CRAN.R project.org/package=nlme>.
- 808 **Qian H, Jin Y. 2016.** An updated megaphylogeny of plants, a tool for generating plant phylogenies

- 809 and an analysis of phylogenetic community structure. *Journal of Plant Ecology*, **9**:233–239.
- 810 **R Core Team. 2019.** R: A language and environment for statistical computing. *R Foundation for*
811 *Statistical Computing*, Vienna, Austria. Version 3.5.3. URL <https://www.R-project.org/>.
- 812 **Ramírez-Barahona S, Sauquet H, Magallon S. 2020.** The delayed and geographically
813 heterogeneous diversification of flowering plant families. *Nature Ecology & Evolution* **4**: 1332-
814 1238.
- 815 **Ramos S, Schiestl F. 2019.** Rapid plant evolution driven by the interaction of pollination and
816 herbivory. *Science* **364**: 193-196.
- 817 **Revell L J. 2012.** Phytools: an R package for phylogenetic comparative biology (and other things).
818 *Methods Ecology and Evolution*, **3**:217-223.
- 819 **Reverté S, Retana J, Gómez JM, Bosch J. 2016.** Pollinators show flower colour preferences but
820 flowers with similar colours do not attract similar pollinators. *Annals of Botany*, **118**:249-257.
- 821 **Rinker HB, Lowman MD. 2004.** Insect herbivory in tropical forests. In: Lowman MD, Rinker
822 HB, eds. *Forest canopies*. San Diego, CA: Elsevier Academic Press 359– 387.
- 823 **Roddy AB, Jiang GF, Cao K, Simonin KA, Brodersen CR. 2019.** Hydraulic traits are more
824 diverse in flowers than in leaves. *New Phytologist* **223**: 193-203.
- 825 **Roddy A, Martínez-Pérez C, Teixido A, Cornelissen T, Olson M, Oliveira R, Silveira F. 2021.**
826 Towards the flower economics spectrum. *New Phytologist* **229**: 665-672.
- 827 **Roslin T, Hardwick B, Novotny V, Petry WK, Andrew NR, Asmus A, Barrio IC, Basset Y,**
828 **Boesing AL, Bonebrake TC, et al. 2017.** Higher predation risk for insect prey at low latitudes
829 and elevations. *Science* **356**: 742-744.
- 830 **Rosumek FB, Silveira FAO, de S. Neves F, de U. Barbosa NP, Diniz L, Oki Y, Pezzini F,**
831 **Fernandes GW, Cornelissen T. 2009.** Ants on plants: a meta-analysis of the role of ants as plant
832 biotic defenses. *Oecologia* **160**: 537–549.
- 833 **Rusman Q, Lucas-Barbosa D, Poelman EH, Dicke M. 2019.** Ecology of Plastic Flowers. *Trends*
834 *in Plant Science* **24**: 725–740.
- 835 **Schiestl F, Johnson S. 2013.** Pollinator-mediated evolution of floral signals. *Trends in Ecology &*
836 *Evolution* **28**: 307-315.
- 837 **Smallegange RC, van Loon JJA, Blatt SE, Harvey JA, Agerbirk N, Dicke M. 2007.** Flower vs.
838 leaf feeding by *Pieris brassicae*: glucosinolate-rich flower tissues are preferred and sustain higher
839 growth rate. *Journal Chemical Ecology* **33**:1831-1844.

- 840 **Smilanich A, Fincher R, Dyer L. 2016.** Does plant apparency matter? Thirty years of data provide
841 limited support but reveal clear patterns of the effects of plant chemistry on herbivores. *The New*
842 *Phytologist* **210**: 1044-57.
- 843 **Soper-Gorden N, Adler L. 2016.** Florivory shapes both leaf and floral interactions. *Ecosphere* **7**:
844 e01326.
- 845 **Soper-Gorden N, Zlonis K, Jahnke M, Marshall E, Horky J, Huddelson C, Etterson J. 2016.**
846 Geographic patterns of seed mass are associated with climate factors, but relationships vary
847 between species. *American Journal of Botany* **103**: 60-72.
- 848 **Strauss SY, Irwin RE. 2004.** Ecological and evolutionary consequences of multispecies plant-
849 animal interactions. *Annual Review of Ecology, Evolution, and Systematics* **35**: 435–466.
- 850 **Tanaka Y, Brugliera F. 2007.** Flower Colour. In: Ainsworth C. eds. *Flowering and its*
851 *Manipulation. Annual Plant Reviews* **20**: 199–239.
- 852 **Teixido AL, Méndez M, Valladaresa F. 2011.** Flower size and longevity influence florivory in
853 the large-flowered shrub *Cistus ladanifer*. *Acta Oecologica* **37**: 418-421.
- 854 **Teixido AL, Valladares F. 2014.** Disproportionate carbon and water maintenance costs of large
855 corollas in hot Mediterranean ecosystems. *Perspectives in Plant Ecology, Evolution and*
856 *Systematics* **16**:83–92.
- 857 **Tscharntke T, Hawkins B. 2002.** Multitrophic level interactions: An introduction. In: Tscharntke
858 T, Hawkins B, eds. *Multitrophic Level Interactions*. Cambridge: Cambridge University Press, 1-7.
- 859 **Turcotte M, Thomsen C, Broadhead G, Fine P, Godfrey R, Lamarre G, Meyer S, Richards**
860 **L, Johnson M. 2014.** Percentage leaf herbivory across vascular plant species. *Ecology* **95**: 788.
- 861 **van Klink R, Bowler DE, Gongalsky KB, Swengel AB, Gentile A, Chase JM. 2020.** Meta-
862 analysis reveals declines in terrestrial but increases in freshwater insect abundances. *Science* **368**:
863 417-420.
- 864 **van der Kooi CJ, Ollerton J. 2020.** The origins of flowering plants and pollinators. *Science* **368**:
865 1306-1308.
- 866 **Waser NM, Chittka L, Price MV, Williams NM, Ollerton J. 1996.** Generalization in pollination
867 systems, and why it matters. *Ecology* **77**: 1043–1060.
- 868 **Waser NM, Ollerton J. 2006.** Plant-pollinator interactions: from specialization to generalization.
869 *University of Chicago Press, Chicago*.
- 870 **Willmer P. 2011.** *Pollination and Floral Ecology*. Princeton: University Press.

- 871 **Xiao L, Labandeira C, Dilcher D, Ren D. 2021.** Florivory of early cretaceous flowers by
872 functionally diverse insects: implications for early angiosperm pollination. *Proceedings of the*
873 *Royal Society B.* **288**:20210320.20210320. //doi.org/10.1098/rspb.2021.0320.
- 874 **Ye Z-M, Jin X, Wang Q-F, Yang C-F, Inouye D. 2017.** Pollinators shift to nectar robbers when
875 florivory occurs, with effects on reproductive success in *Iris bulleyana* (Iridaceae). *Plant Biology*
876 **19**: 760-766.
- 877 **Zhang S, Zhang Y, Ma K. 2016.** Latitudinal variation in herbivory: Hemispheric asymmetries
878 and the role of climatic drivers. *Journal of Ecology*, **104**:1089– 1095.
879

***New Phytologist* Supporting Information**

Article title: **Revisiting florivory: an integrative review and global patterns of a neglected interaction**

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Article acceptance date: 02 August 2021

The following Supporting Information is available for this article:

Florivory data of the 98 species were extracted from previously published studies (for more details see Supporting Information Table S1) and the phylogenetic tree was built using the Qian & Jin (2016) database.

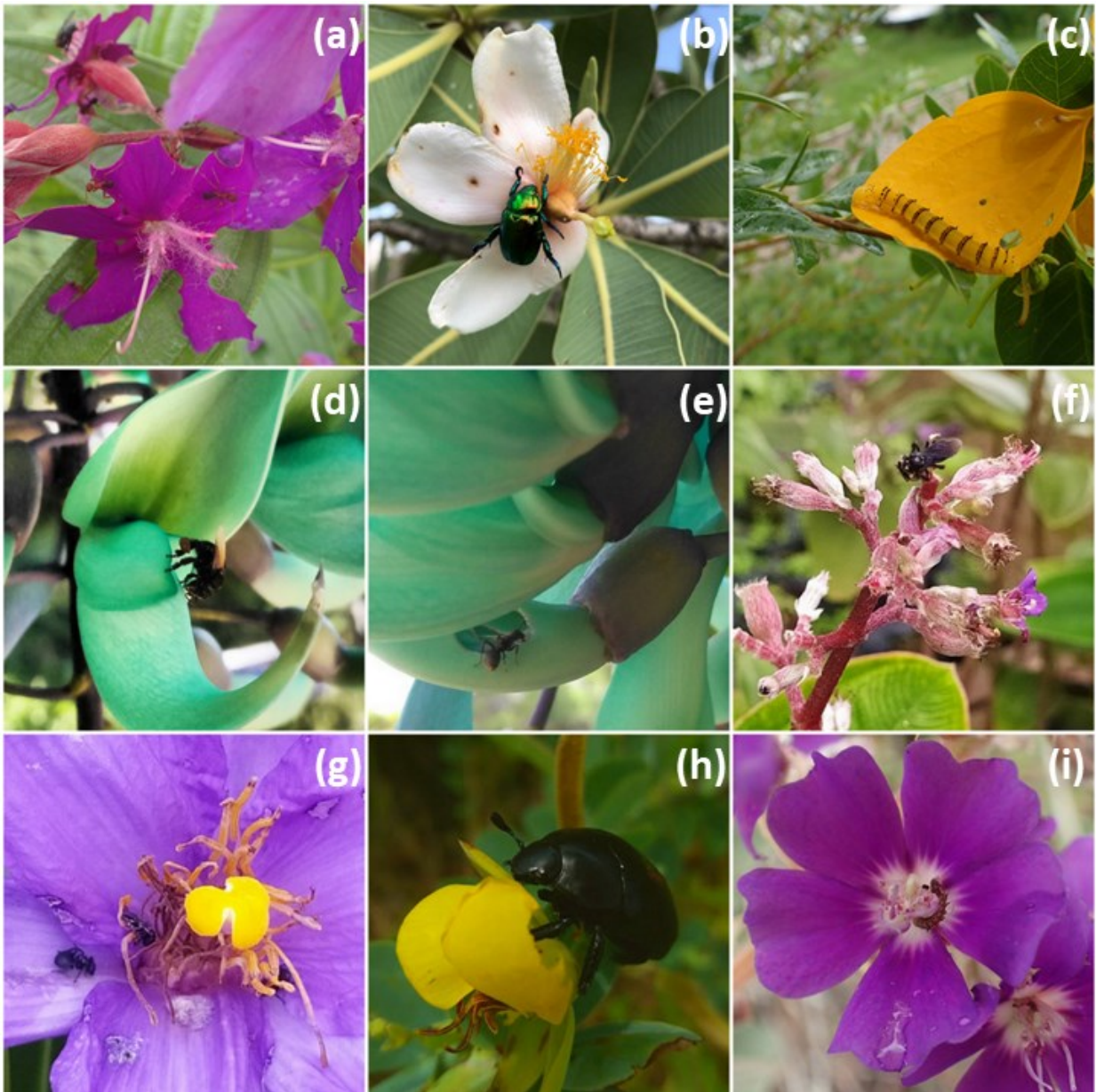


Fig. S2- Examples of floral enemies on tropical plant species. (a) Ants chewing on petals of *Pleroma* (Melastomataceae). (b) A beetle eating reproductive parts of a *Kielmeyera coriacea* (Clusiaceae) flower (Photo by Cassio Cardoso). (c) A caterpillar, one of the most common florivores, chewing on flowers of *Senna spectabilis* (Fabaceae). (d) A *Trigona* bee, one of the most common primary nectar robbers in the Neotropics, stealing nectar from *Strongylodon macrobotrys* (Fabaceae). (e) Secondary nectar robbing by ants on flowers of *Strongylodon macrobotrys* (Fabaceae). (f) A *Trigona* nectar robbing bee on flower buds of *Pleroma heteromallum* (Melastomataceae). (g) *Cephalotes* ants can act as florivores, nectar thieves or nectar robbers on a *Vellozia* flower (Velloziaceae). (h) A beetle thieving nectar from a flower of *Chamaecrista desvauxii* (Fabaceae) (Photo by Lorena Valadão-Mendes). (i) A tiny geometrid caterpillar feeding upon reproductive parts of *Pleroma heteromallum* flower (Melastomataceae).

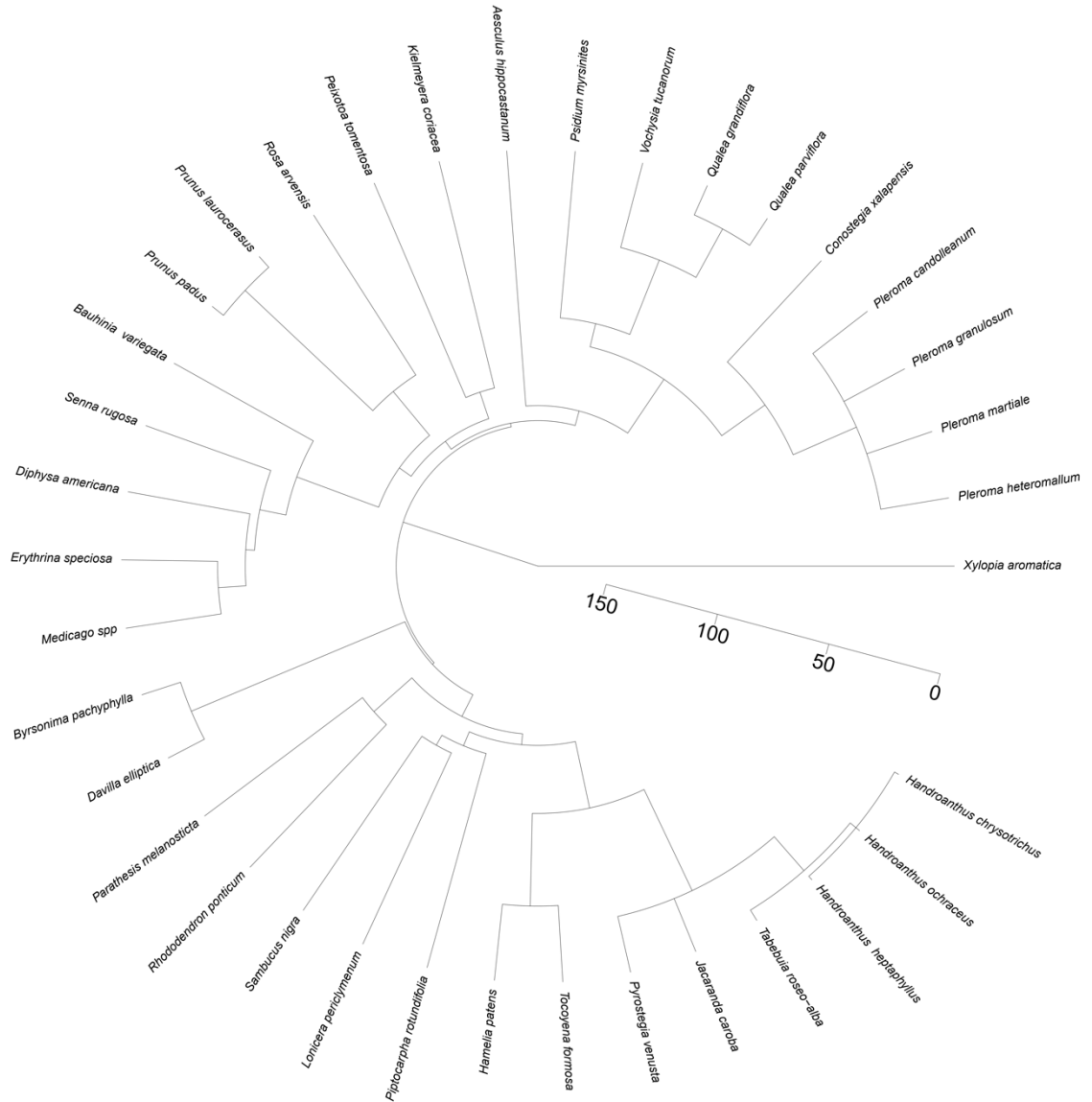


Fig S3 - Phylogenetic tree of the 36 plant species collected using a standardized protocol. The phylogenetic tree was built using the Qian & Jin (2016) database.

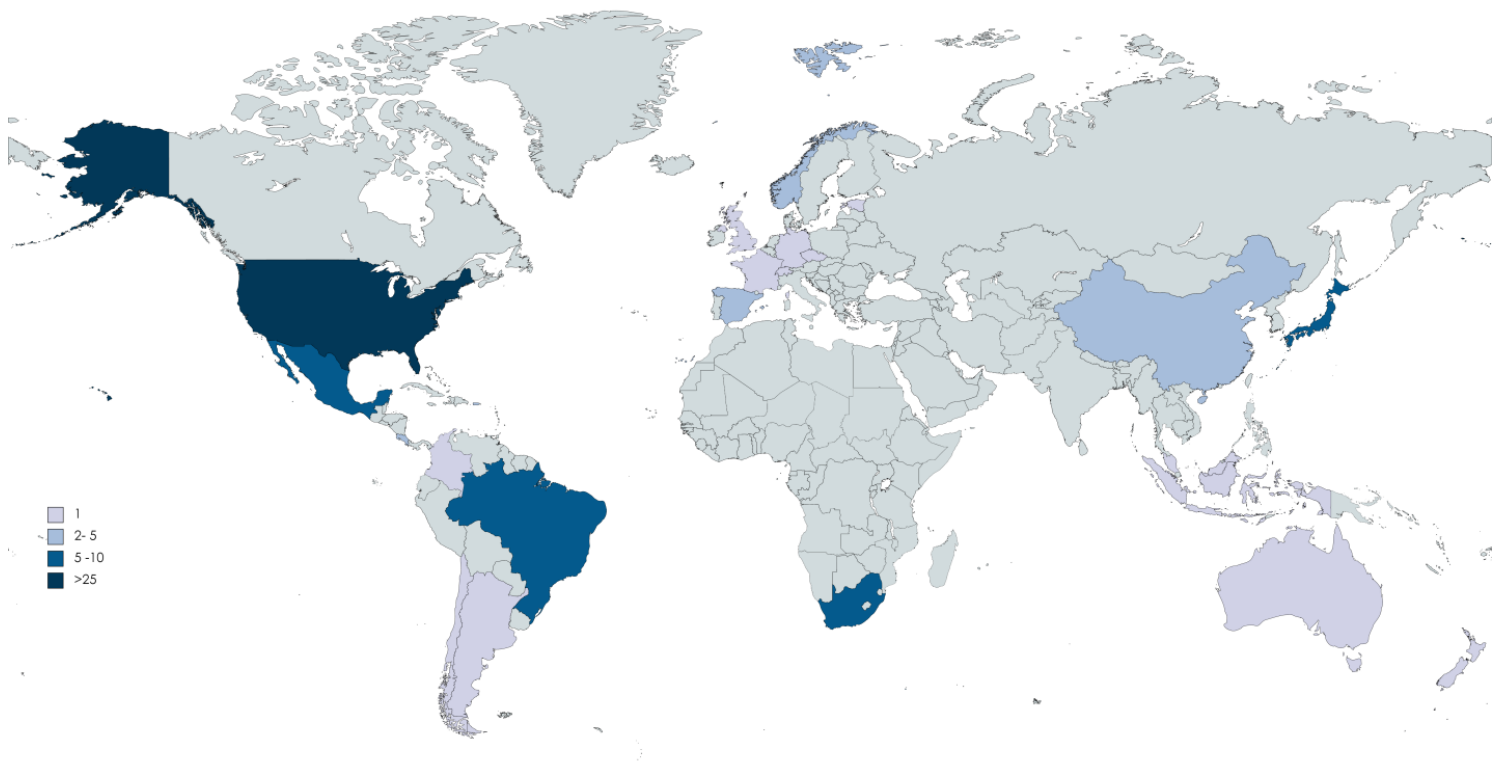


Fig. S4- Distribution and number of florivory studies in the florivory global database per country as a heat map in hues of blues. From the countries where studies are available, USA had the greatest number of studies and New Zealand had the least.

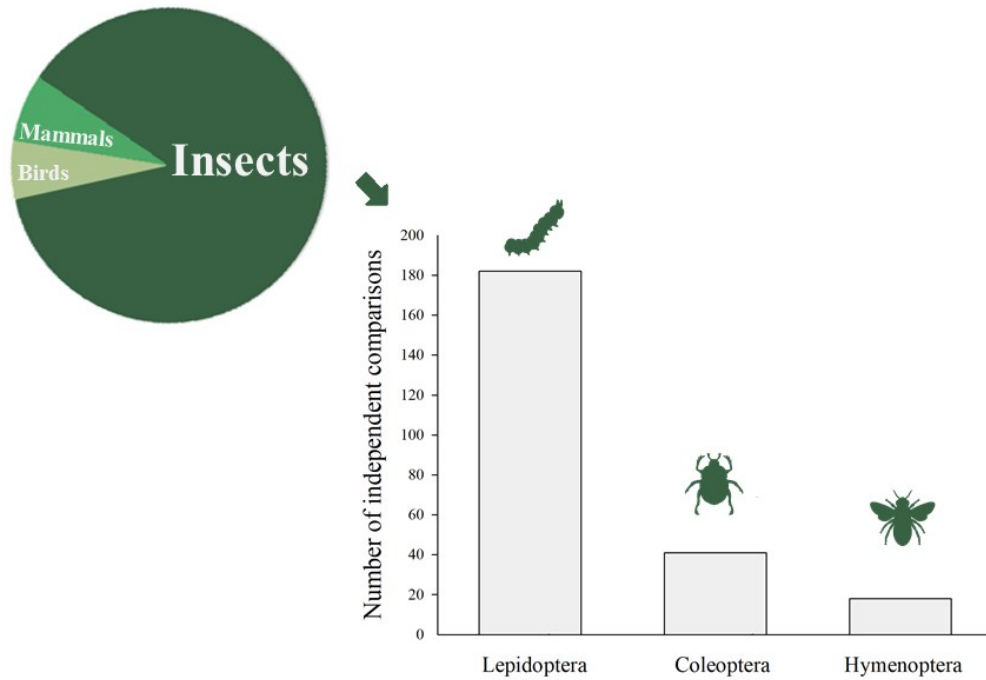


Fig S5 - Main florivore taxa according to broad taxonomic categories in the florivory global database. Insects were the most frequent taxa reported attacking flowers and chewing insects are responsible for inflicting the greatest amount of damage to the plants evaluated.



Fig. S6 - The distribution of florivory studies by region and by botanical families in (a) temperate regions and (b) tropical regions. Data come from published studies included in our global database.

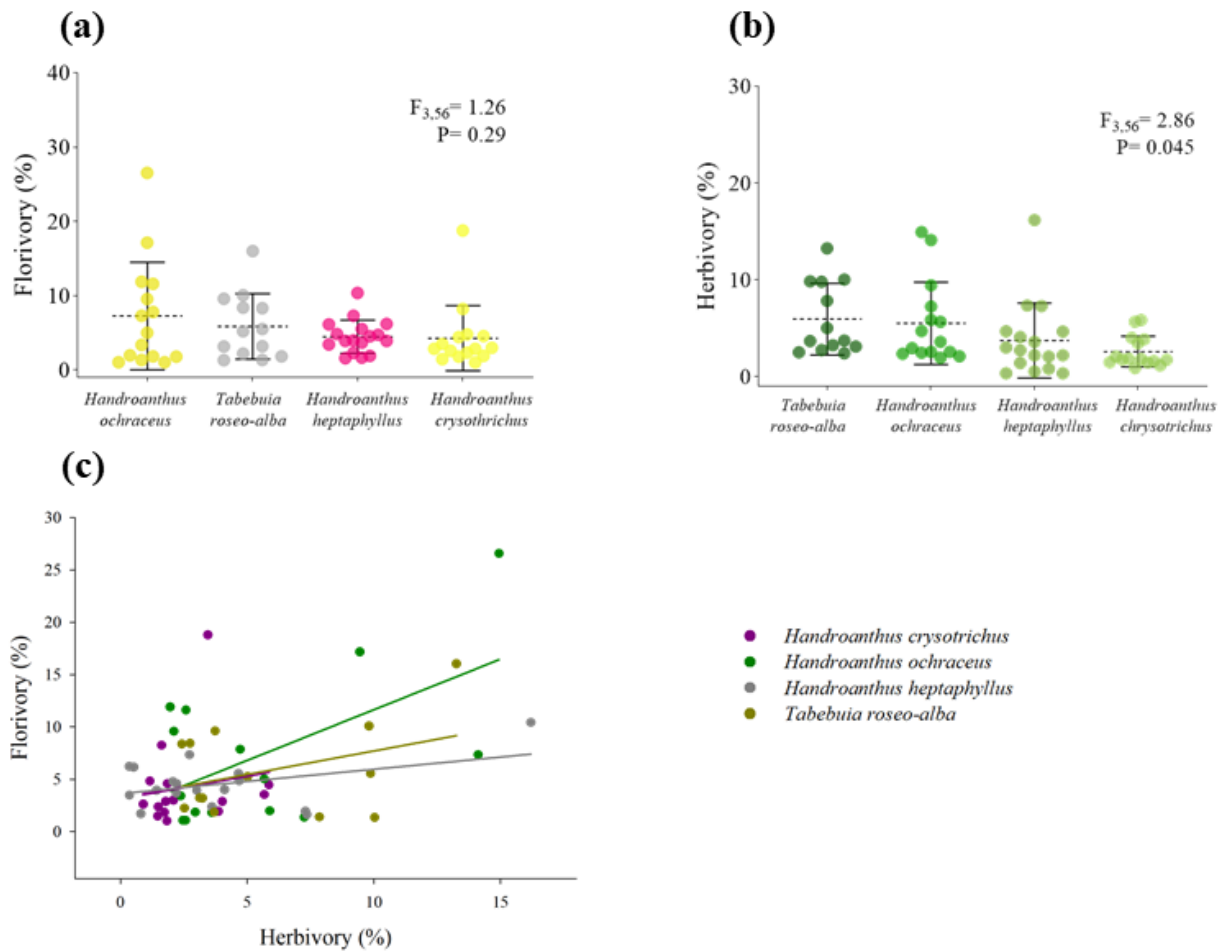


Fig. S7 - Levels of a) florivory and b) leaf herbivory, measured by the percentage of average damage, for four plant species of the Bignoniaceae family and c) relationship between florivory and herbivory for each species of Bignoniaceae. Each point represents individual replicated trees. The line indicates the mean and the error bars denote SD.

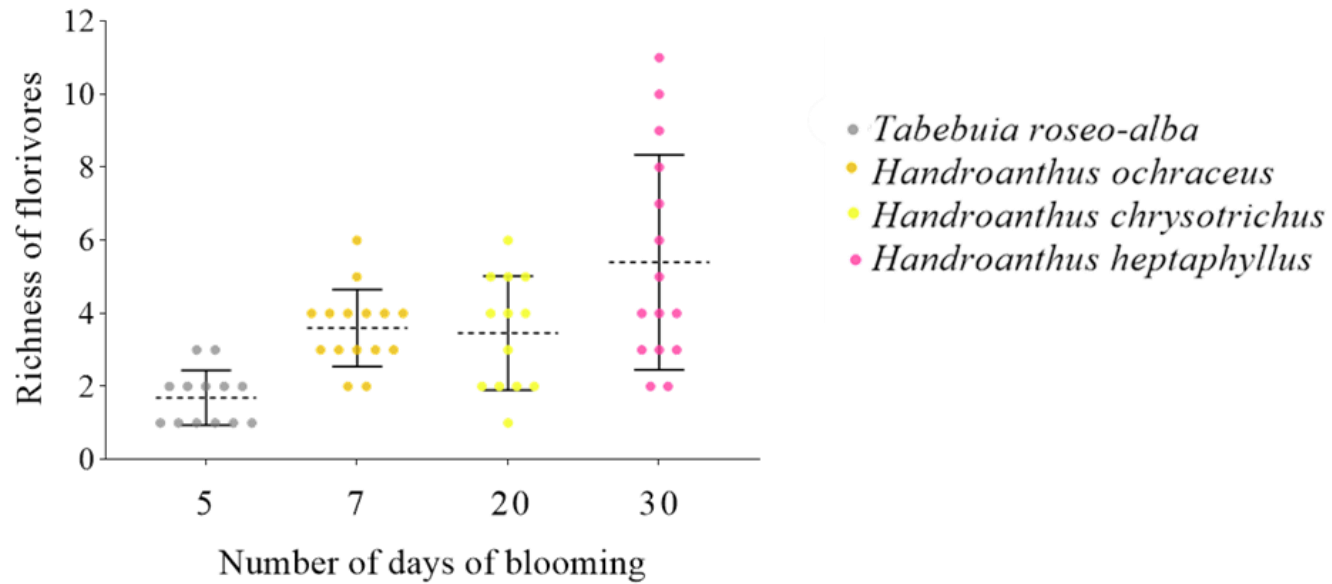
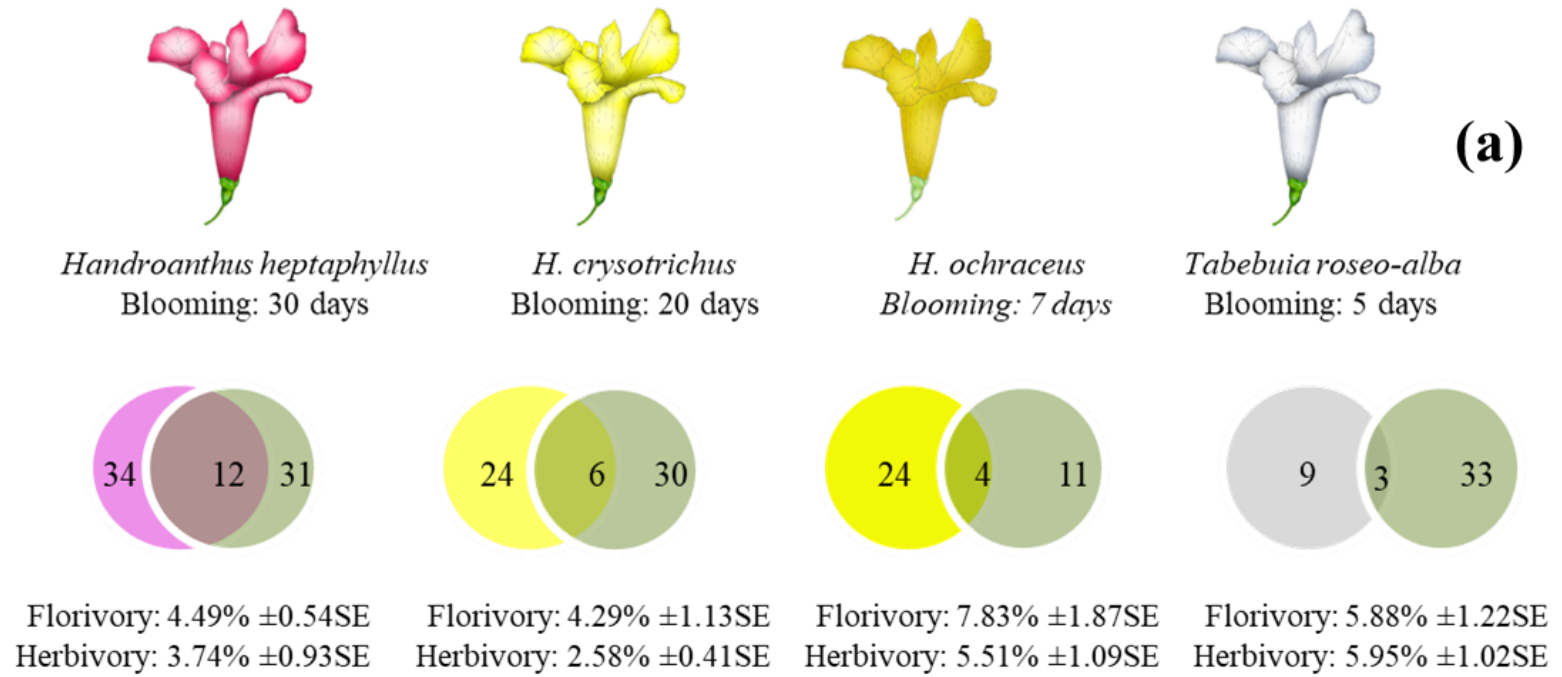
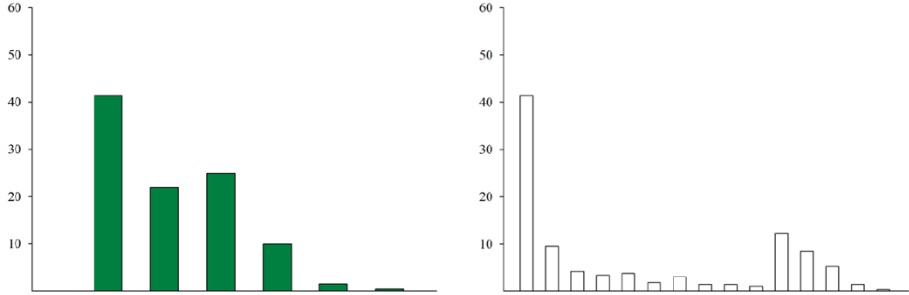


Fig S8 - a) Species richness of florivores and herbivores in four species of Bignoniaceae with corresponding levels of florivory and herbivory and b) the relationship between number of days plants were blooming and richness of florivores. Each dot on panel represents an individual plant sampled and colors represent the colors of the flowers from each species. The line indicates the mean and the error bars denote SD.

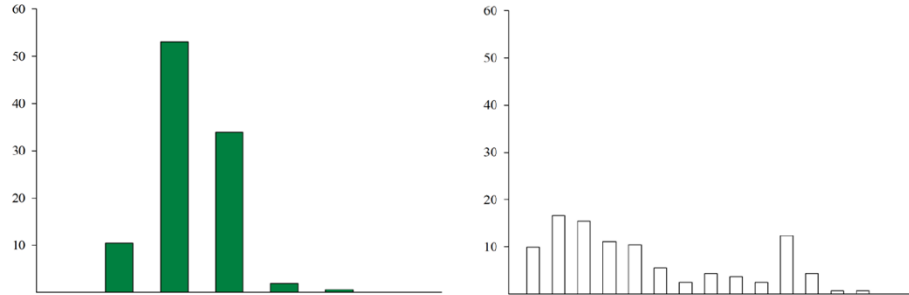
IFD

ImageJ

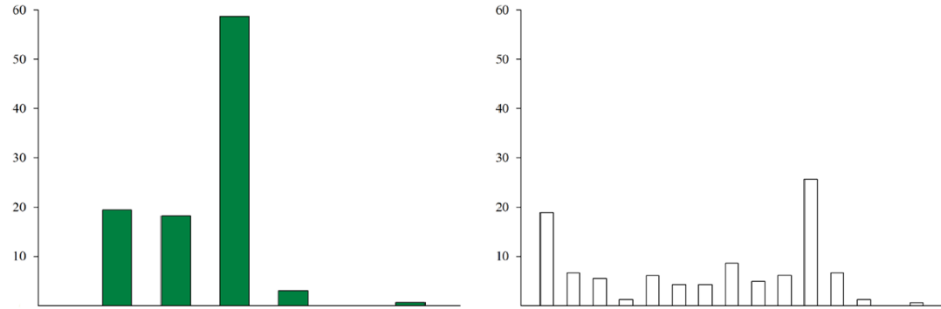
Handroanthus ochraceus



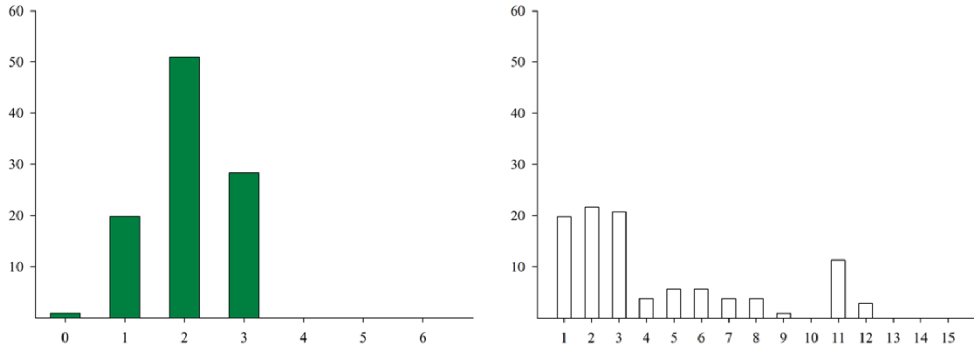
Pleroma candolleianum



Jacaranda caroba



Pleroma martiale



Frequency distribution of damage

Damage classes

Fig S9. Distribution of flower damage into damage classes according to the IFD (left panels) and digital measurements of the same flowers using ImageJ software. The x-axis on both groups of panels indicate the classes of damage in percentage of floral area removed.

Methods S1 – A protocol for florivory evaluation (sampling, measuring damage and general guidelines).

1) Samples should be conducted in natural areas, protected areas or in modified areas, which should be described in the appropriate column.

2) Get GPS data (exact latitude and longitude), as well as altitude for the selected locality.

3) Select at least 5 leaf-bearing plant species that are flowering (preferably during the flowering peak) in the period you will sample. Record Latin name, the geographic coordinates, elevation, the biome where they were collected, growth form, flower color, and data of collection.

4) For each species, randomly choose 5 individual plants as replicates, at least 10m apart. From a distance of 5 m from you, choose up to 20 flowers to be sampled by plant, 5 from each corner around the canopy. Flowers should be fully developed and open. Researchers should record if anthers and/or stigma(s) are mature.

5) Store flowers in plastic bags (in the fridge) if you cannot examine the flowers immediately

6) Note that the flowers from the 5 individual replicates per species should be stored in plastic bags separately. You will end up with 25 samples per locality (5 species X 5 replicates, each replicate with 20 flowers).

7) For inflorescences/bouquets, collect 5 inflorescences/bouquets per replicate plant, but count and register all flowers in the bouquet.

8) The flowers must be digitized as soon as possible (n=500 per locality in the case of single flowers or 125 inflorescences per locality in case of bouquets). The flowers can be digitized through a scanner or digital camera, with a scale for measurements in millimeters and/or centimeters (e.g., a plastic or metal ruler). You must also identify which species, individual and flower each image belongs to.

10) If the flower has with tubular or buliform corolla, gently cut the corolla laterally so that the whole corolla area is digitized. You can use a double-sided tape so that the sides of the corolla are perfectly stretched or a piece of transparent glass can be placed over the flower.

11) If the flower has a radially symmetrical corolla, detach the petals carefully and use the double-sided tape to stretch them when necessary.

12) All flowers must be measured in the ImageJ® software. It is a free software and you can download it at <https://imagej.nih.gov/ij/download.html>.

13) In ImageJ®, open the image to be measured. Using the measuring line, carefully draw a 1 cm or 10mm line, using the ruler as scale. Click “Analyze” > “Set scale”. The program will open a box where you must standardize the scale at which your data will be measured. In the "known distance" box, add 1 cm or 10 mm and select the option "Global". All your measurements will be calibrated to this set line.

14) You must measure the original images of flowers. You will need to redraw the flower, contouring it with the pencil to account for flower areas lost to florivores. Use the zoom tool (Ctrl+) to precisely see the holes in the image.

15) Select the area icon on the top of the screen on ImageJ® and click on each area you have measured (total area of flower and "filled" lost area) and press the "M" key on the keyboard. This will return the value of the areas that you have measured. Record these values on a spreadsheet.

16) To account for intraspecific variation, data should be recorded for each plant species, each individual replicate and each individual flower in a spreadsheet.

Table S1 – Global database of florivory in plants.

Uploaded separately and deposited in figshare under d.o.i [10.6084/m9.figshare.15106230](https://doi.org/10.6084/m9.figshare.15106230)

Table S2 – Plant species collected for this study with a standardized protocol. Data indicates plant species collected in three countries for measurements of florivory, grouped by species, family, country, region, floral symmetry, inflorescence type, flower color and plant growth form. TROP= Tropical, TEMP= Temperate. Z= zygomorphic, A= actinomorphic. I= two or more flowers in an inflorescence, S= single flowers.

Plant species	Plant Family	Country	Region	Flower symmetry	Flower number	Flower color	Growth form
<i>Pyrostegia venusta</i>	Bignoniaceae	Brazil	TROP	Z	I	orange	liana
<i>Jacaranda caroba</i>	Bignoniaceae	Brazil	TROP	Z	I	purple	shrub
<i>Handroanthus chrysotrichus</i>	Bignoniaceae	Brazil	TROP	Z	I	yellow	Tree
<i>Handroanthus ochraceus</i>	Bignoniaceae	Brazil	TROP	Z	I	yellow	Tree
<i>Handroanthus heptaphyllus</i>	Bignoniaceae	Brazil	TROP	Z	I	purple	Tree
<i>Tabebuia roseo-alba</i>	Bignoniaceae	Brazil	TROP	Z	I	white	Tree
<i>Bauhinia variegata</i>	Fabaceae	Brazil	TROP	A	S	white	Tree
<i>Senna Rugosa</i>	Fabaceae	Brazil	TROP	A	I	yellow	Tree
<i>Erythrina Speciosa</i>	Fabaceae	Brazil	TROP	Z	I	red	Tree
<i>Pleroma granulatum</i>	Melastomataceae	Brazil	TROP	A	I	purple	Tree
<i>Pleroma candolleianum</i>	Melastomataceae	Brazil	TROP	A	I	purple	shrub
<i>Pleroma martiale</i>	Melastomataceae	Brazil	TROP	A	I	purple	shrub
<i>Pleroma</i>	Melastomataceae	Brazil	TROP	A	I	purple	shrub

<i>heteromallum</i>							
<i>Peixotoa tomentosa</i>	Malpighiaceae	Brazil	TROP	A	I	yellow	shrub
<i>Piptocarpha rotundifolia</i>	Asteraceae	Brazil	TROP	Z	I	yellowish	Tree
<i>Kielmeyera coriácea</i>	Calophyllaceae	Brazil	TROP	A	I	white	Tree
<i>Davilla elliptica</i>	Dilleniaceae	Brazil	TROP	A	I	yellow	Tree
<i>Byrsonima pachyphylla</i>	Malpighiaceae	Brazil	TROP	Z	I	yellow	Tree
<i>Vochysia tucanorum</i>	Vochysiaceae	Brazil	TROP	A	I	yellow	Tree
<i>Xylopia aromática</i>	Annonaceae	Brazil	TROP	A	IS	white	Tree
<i>Psidium myrsinites</i>	Myrtaceae	Brazil	TROP	Z	I	white	Tree
<i>Tocoyena formosa</i>	Rubiaceae	Brazil	TROP	Z	I	yellow	Tree
<i>Qualea grandiflora</i>	Vochysiaceae	Brazil	TROP	Z	S	yellow	Tree
<i>Qualea parviflora</i>	Vochysiaceae	Brazil	TROP	Z	S	purple	Tree
<i>Medicago sp.</i>	Fabaceae	Mexico	TROP	A	I	violet	shrub
<i>Hamelia patens</i>	Rubiaceae	Mexico	TROP	Z	I	red	shrub
<i>Diphysa americana</i>	Fabaceae	Mexico	TROP	A	I	yellow	Tree
<i>Parathesis melanosticta</i>	Primulaceae	Mexico	TROP	A	I	pink	Tree
<i>Conostegia xalapensis</i>	Melastomataceae	Mexico	TROP	A	I	pink	tree

<i>Prunus padus</i>	Rosaceae	Scotland	TEMP	A	I	white	Tree
<i>Aesculus Hippocastanum</i>	Hippocastanaceae	Scotland	TEMP	A	I	white	Tree
<i>Sambucus nigra</i>	Adoxaceae	Scotland	TEMP	A	I	white	Tree
<i>Rhododendron ponticum</i>	Ericaceae	Scotland	TEMP	A	I	purple	Tree
<i>Prunus laurocerasus</i>	Rosaceae	Scotland	TEMP	A	I	white	Tree
<i>Lonicera periclymenum</i>	Caprifoliaceae	Scotland	TEMP	Z	I	white	shrub
<i>Rosa arvensis</i>	Rosaceae	Scotland	TEMP	A	S	white	shrub

Table S3- Plant species collected for this study with a standardized protocol which was measured using ImageJ. Data indicates plant species collected only in Brazil for measurements of florivory, grouped by species, family, floral symmetry, inflorescence type, flower color, plant growth type and florivory levels. Z= zygomorphic, A= actinomorphic. I= two or more flowers in an inflorescence, S= single flowers, SE = standard error.

Plant species	Plant Family	Flower symmetry	Flower number	Flower color	Growth type	Florivory levels
<i>Pyrostegia venusta</i>	Bignoniaceae	Z	I	orange	liana	3.9±0.98SE
<i>Jacaranda caroba</i>	Bignoniaceae	Z	I	purple	shrub	6.29±1.09SE
<i>Handroanthus chrysotrichus</i>	Bignoniaceae	Z	I	yellow	tree	4.29±1.13SE
<i>Handroanthus ochraceus</i>	Bignoniaceae	Z	I	yellow	tree	7.3±1.87SE
<i>Handroanthus heptaphyllus</i>	Bignoniaceae	Z	I	purple	tree	4.49±0.54SE
<i>Tabebuia roseo-alba</i>	Bignoniaceae	Z	I	white	tree	5.88±1.22SE
<i>Bauhinia variegata</i>	Fabaceae	A	S	white	tree	2.76±0.43SE
<i>Senna Rugosa</i>	Fabaceae	A	I	yellow	tree	7.21±1.3SE
<i>Erythrina Speciosa</i>	Fabaceae	Z	I	red	tree	3.51±0.6SE
<i>Pleroma granulosum</i>	Melastomataceae	A	I	purple	tree	8.79±5.45SE
<i>Pleroma candolleianum</i>	Melastomataceae	A	I	purple	shrub	5.48±0.81SE
<i>Pleroma martiale</i>	Melastomataceae	A	I	purple	shrub	4.08±0.87SE
<i>Pleroma heteromallum</i>	Melastomataceae	A	I	purple	shrub	9.65±2.83SE

<i>Peixotoa tomentosa</i>	Malpighiaceae	A	I	yellow	shrub	32.11±6.41SE
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Table S4 - Results of the Phylogenetically Generalized Least Square (PGLS) models, used to evaluate the relationship between florivory (percentage of attacked flowers) and several factors in 98 plant species from published florivory studies in our database.

Response variable	Fixed factors (number of levels)	Random factors (number of levels)	F-value (d.f.)	P-value	r Brownian (P-value)	AIC value
Frequency of attacked flowers	Region (2)	Plant species (98)	8.35 (1,293)	0.004	-0.155 (0.2128)	3462.29
	Flower symmetry (3)	Plant species (98)	1.55 (2,294)	0.213	-0.08 (0.1230)	3460.06
	Flower number (3)	Plant species (98)	4.07 (2,294)	0.007	-0.071 (0.2285)	3444.98
	Florivore identity (8)	Plant species (98)	9.37 (7, 293)	0.0001	-0.082 (0.223)	3371.80
	Florivore guild (6)	Plant species (98)	3.43 (5,293)	0.005	-0.046 (0.127)	3378.90
	Flower color (7)	Plant species (98)	0.92 (6,293)	0.47	-0.267 (0.061)	3424.70

Table S5 - Results of Phylogenetically Generalized Least Square (PGLS) models, used to evaluate the relationship between florivory levels (Index of flower damage, IFD) and several factors in 36 plant species sampled in the field using a standardized protocol.

Response variable	Fixed factors (number of levels)	Random factors (number of levels)	F-value (d.f.)	P-value	r (Brownian model)	AIC value
Florivory (IFD)	Region (2)	Plant species (36)	11.33 (1,276)	0.001	-0.498 (0.605)	722.35
	Flower symmetry (2)	Plant species (36)	0.052 (1,276)	0.819	-0.104 (0.112)	732.68
	Flower color (7)	Plant species (36)	8.82 (6,276)	0.07	-0.43 (0.287)	729.57
	Growth form (3)	Plant species (36)	0.76 (2,276)	0.46	-0.158 (0.292)	729.61

Capítulo II:

**Patterns of florivory vary with
an urbanization degree
in tropical cities**

1 **Patterns of florivory vary with an urbanization degree in tropical cities**

2

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4

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19

20 **Suggested journals: Urban ecosystem, Ecological Indicators, Journal of urban ecology**

21 Abstract

22

23 Florivory can impact plant reproduction both directly when florivores consume pistils, stamens,
24 and/or ovules or indirectly, when florivores modify floral attributes related to attractiveness by
25 consuming petals, sepals, and floral buds, interfering with plant-pollinator relationships. We still
26 know little about how these direct and indirect impacts affect plants and about the mechanisms and
27 processes that make certain species more susceptible to florivory. Urbanization has been identified
28 as one of the main threats to biodiversity in the last few decades as the conversion of natural areas
29 into urban areas produces significant changes in the biotic and abiotic environment, due to the
30 replacement of native species by exotic ones, formation of heat islands and pollution. Recent
31 studies have indicated that these changes negatively affect both the occurrence of insects in cities
32 and cause changes in the composition and quality of host plants with potential consequences for
33 insect-plant interactions. Here we evaluated if florivory patterns vary with an urbanization degree
34 to examine whether environmental changes caused by urbanization influence florivory levels and,
35 whether florivory levels are driven by plant attributes and functional and phylogenetic traits of
36 flowers and plants. Percentage of attacked flowers (florivory incidence) and area removed from
37 petals (florivory intensity) by florivores were examined in 61 species distributed into 20 botanical
38 families, sampled in areas with different degrees urbanization considering the plant origin (native
39 or exotic), plant function (ornamental or non-ornamental), and floral and vegetative attributes
40 shared across plants lineages as potentially determining factors on florivory levels. Our results
41 indicated that the intensity of florivory decreased with increased urbanization, due to the
42 replacement of native species by exotic ones and the highest frequency of occurrence of ornamental
43 plants for florivores, because of environmental stress in urban environments. On the other hand,
44 urbanization did not modify patterns of florivory incidence. Florivory levels were also influenced
45 by intraspecific variation in the functional traits of the plants in areas over different degrees of
46 urbanization. We believe that the effect of urbanization and functional traits on levels of florivory
47 are mediated mainly by changes in the nutritional quality of host plants for florivores, which varies
48 depending on plant origin and the type of ecological function it performs. We suggest that the
49 abundance and richness of florivores also contributed to our results, and we suggest that this should
50 be better evaluated in future studies.

51

52 **Keywords:** florivory, floral damage, floral herbivory, urbanization, insects

53 **Introduction**

54

55 Antagonisms between animals and plants usually bring negative consequences for plant
56 development, but in addition, they can also mediate negative effects to other ecological
57 interactions, such as pollination (see McCall and Irwin, 2006), which is essential for reproduction
58 of most angiosperms. The consequences of herbivory on plants have been well documented in the
59 last decades (see Strauss 1996; McCall and Irwin 2006; Bronstein et al., 2007; Solga et al., 2014;
60 Mendes et al., 2021), but florivory, on the other hand, has been much less evaluated in the
61 ecological literature (Boaventura et al., 2022). Florivory is defined as any type of damage to flowers
62 involving the consumption of floral tissues by vertebrate or invertebrate animals, but mainly insects
63 (McCall and Irwin, 2006; Boaventura et al., 2022). Florivory can impact plant reproduction both
64 directly when florivores consume floral structures effectively linked to this function - such as
65 pistils, stamens, and ovules, reducing the number of gametes produced - or indirectly, when
66 florivores modify floral attributes related to flower attractiveness by consuming petals, sepals, and
67 floral buds, interfering therefore with plant-pollinator relationships (McCall and Irwin, 2006;
68 Caruso et al., 2019; Rusman et al., 2019). Consequently, florivores cause negative selective
69 pressures on the floral attributes associated with attraction to pollinators, which may result in a
70 reduction in the ecological service provided by those animals (Caruso et al., 2019; Rusman et al.,
71 2019; Boaventura et al., 2022).

72 Despite the potential negative impact of florivory, we still know little about how its direct
73 and indirect effects affect plants and about the mechanisms and processes that make certain species
74 more susceptible to this antagonism (Boaventura et al., 2022). It has been suggested, for example,
75 that abiotic factors such as elevation (Gélvez-Zúñiga, 2021) and seasonal variations in
76 environmental conditions (Ortiz et al., 2023), as well as functional traits of flowers and plants
77 (Feeny, 1976; Boaventura et al., 2022) directly influence levels of florivory. Although the global
78 levels of florivory are about twice as higher as the global levels reported for herbivory (see Kozlov
79 et al., 2015 for herbivory and Boaventura et al., 2022 for florivory levels) it still unclear how and
80 which environmental, phylogenetic, floral and vegetative attributes influence this interaction.
81 Therefore, quantifying florivory and determining the patterns and processes related to it is an
82 important aspect of ecology and is the first step towards improving our understanding of this
83 interaction (Boaventura et al., 2022).

84 Current knowledge about how abiotic and biotic factors that influence plant distribution and
85 development - such as climate and the richness and abundance of florivores (Fenoglio et al., 2020)
86 - affect florivory levels is still quite narrow (Boaventura et al., 2022). Urbanization has been
87 recently identified as one of the main threats to biodiversity (Johnson and Munshi-South, 2017).
88 The conversion of natural habitats into urban areas produces significant changes in both the abiotic
89 and biotic environment due to the reduction of native vegetation cover, increased surface
90 impermeabilization, formation of heat islands, pollution, habitat fragmentation, and replacement
91 of native species with exotic ones (Elmqvist et al., 2013; Wu, 2014; Johnson and Munshi-Shut,
92 2017; Kondratyeva et al., 2020; Fenoglio et al., 2021). Recent studies have also indicated that these
93 changes negatively influence the occurrence of insects in cities (Fenoglio et al., 2020, 2021;
94 Cornelissen et al., 2023) and cause changes in the composition and quality of host plants, with
95 several consequences for insect-plant interactions (Fenoglio et al., 2020, 2021; Moreira et al.,
96 2019a). Although some patterns of florivory in natural habitats have emerged lately, we still have
97 limited information for a narrow number of plant species (e.g., Malo et al., 2001; Irwin et al., 2014;
98 2018) regarding the effects of changes in land use and plant-insect interactions. Although patterns
99 of insect diversity in the cities have been summarized (see Cornelissen et al., 2023), effects of
100 urbanization on ecological interactions have been largely unexplored, except for pollination
101 (Teixido et al., 2022).

102 From the perspective of insect-flower antagonisms, urbanization could lead to an increase
103 in levels of florivory if, for example, plants in cities are more susceptible to florivory compared to
104 plants in natural areas, if urbanization represent a source of abiotic stress which increases plant
105 susceptibility to antagonisms (Cuevas-Reyes et al., 2013), as proposed by the Plant Stress
106 Hypothesis (PSH, White, 1969;1984), and/or because plant species are exotic and have not yet
107 developed adequate anti-florivory defenses (Raupp et al., 2010, Stemmelen et al., 2020). On the
108 other hand, urbanization could also lead to a decrease in florivory levels due to the decrease in the
109 abundance and diversity of florivores in these areas (Fenoglio et al., 2020), or even if
110 exotic/ornamental plants represent a low-quality food resource for these insects (Fenoglio et al.,
111 2021). Studies on *Gelsemium sempervirens* in rural and urban environments indicated that levels
112 of florivory increased with urbanization, due to a decrease in corolla tube length and greater floral
113 apparency (i.e., larger number of flowers produced per day) in an urban environment (Irwin et al.,
114 2014; 2018). However, most of these studies only estimated florivory as a binary variable (presence

115 or not of florivory in flowers in a population), and/or performed only visual estimates of the extent
116 of floral damage, which may not be an adequate measurement of the extent of florivory levels and
117 its impacts on plant populations (Boaventura et al., 2022). To date, we are not aware of any study
118 that has evaluated the effects of urbanization on florivory at the plant community level, or of studies
119 that have evaluated how floral traits influence levels of flower damage in the cities. If florivory
120 really does increase with urbanization, we can expect that urban plants will have an amplified
121 fitness loss due to the combination of high levels of florivory and scarcity of pollinating insects
122 (Fenoglio et al., 2020), which could ultimately lead to a collapse in the urban plant community due
123 to a decrease in the ecological service of pollination.

124 To advance our understanding of impacts of florivory in plants, it is necessary to take into
125 account the biotic and abiotic factors that influence damage levels in different plant lineages and
126 in different environmental contexts. Although urbanization represents one of the most dramatic
127 environmental changes on Earth and although it is expected that urban expansion will continue to
128 occur rapidly in the upcoming decades (Fenoglio et al., 2020;2021), its effect on insect-flower
129 antagonisms are still poorly understood. Even less is known about how floral and vegetative
130 attributes influence levels of florivory in the urban context.

131 In this study we assessed florivory patterns in several plant species sampled in areas within
132 different degrees of urbanization to understand the dynamics of plant-florivore interactions,
133 considering i) environmental and plant attributes related to urbanization (level of urbanization of
134 the collection site, origin and type of plant function), ii) taxonomic relatedness (phylogenetic
135 relationship between species) and iii) biological variation among species and individuals
136 (functional traits of plants and flowers). As the methods for collecting and evaluating florivory
137 have been recently standardized (see Boaventura et al., 2022), damage levels were evaluated both
138 as the percentage of attacked flowers (florivory incidence) and as the percentage of floral area
139 removed (florivory intensity). Following the arguments that florivory can be influenced by
140 urbanization, we hypothesized that: 1) environmental changes caused by urbanization modify
141 patterns of plant-florivore interaction and directly influence florivory levels; 2) levels of florivory
142 will be related to plant attributes related to urbanization, such as plant origin (exotic or native) and
143 its ecological function (ornamental vs. non-ornamental plants); 3) florivory levels will be related
144 to functional and phylogenetic traits of flowers and plants. We also assessed whether incidence and

145 intensity of florivory are related within individual plants and species, and we assessed the identity
146 of the florivores responsible for the damage.

147

148 **Material and Methods**

149

150 *Field sampling*

151

152 Between February 2021 to June 2022, we performed periodic collections campaigns of data
153 collection (every 2-3 months) in several locations in the central region of Minas Gerais state and
154 61 plant species (n= 305 individuals) were evaluated during its flowering period. The plant species
155 were collected in areas under different degrees of human disturbance and the main land use impact
156 evaluated in our study was the degree of urbanization of each collection site. Here we classified as
157 1) natural areas those under environmental protection, such as Federal and State Parks, away from
158 urban centers; 2) rural or periurban areas were those where we detected dirt roads, farms, small
159 towns with less than 5,000 inhabitants and areas in the boundaries between cities and
160 environmental protected areas in which most of the soil coverage was permeable; and 3) urban
161 areas were those composed of cities under stronger human influence, with more than 5,000
162 inhabitants and where most native vegetation cover was replaced by impervious surface cover.

163 All plant species studied here were selected considering taxonomic identity in a way to
164 sample the largest possible number of genera and/or families (Fig.S1). The plant species were
165 chosen and collected according with the protocol for sampling and measurements of florivory
166 proposed for Boaventura et al. (2022) and plants were sampled according to its flowering
167 phenology (i.e., whether they were flowering during the collection period). At least five
168 synchronous individuals with flowers were sampled per plant species and individuals were spaced
169 at least 10m apart. In the case of climbing plants or herbs arranged in intermittent patches, all
170 branches or patches within a radius of 2 m from the main distinguishable branch or patch were
171 considered as a single individual. For herbs arranged in continuous patches, individual flowers
172 spaced at least 3 m apart were collected. We collected 20 flowers for individual and 100 flowers
173 per plant species, totaling 6,100 flowers evaluated in this study. Whenever possible, we also
174 collected and/or identified the florivore insects that were responsible for flower damage at the

175 moment of sampling. Plant species were identified consulting specialists, taxonomic keys and,
176 using an online herbarium (<http://reflora.jbrj.gov.br/reflora/PrincipalUC/PrincipalUC.do>).

177

178 *Florivory levels*

179

180 As proposed by Boaventura et al. (2022) ontogenetic traits intrinsic to plant species and
181 lineages and also attributes related to the site where plant species were collected can potentially
182 influence levels of florivory. We firstly categorized the sampled sites according to the degree of
183 urbanization (preserved areas: n=24, periurban/rural areas: n=19 and urban areas: n=18). For each
184 plant species sampled, we also categorized species according to plant attributes related to
185 urbanization, such as plant origin in the site where it was collected (native or exotic) and type of
186 plant function in each site (ornamental or non-ornamental). Finally, plants were categorized by
187 type of floral arrangement (single flower or inflorescences), flower symmetry (actinomorphic or
188 zygomorphic), flower color (the predominant color of the corolla as perceived by the human eyes),
189 flower size (corolla area in cm²), and plant growth forms (lianas, herbs, shrubs, or trees).

190 Because florivory can be evaluated as both the frequency with which flowers of a plant or
191 a population are attacked by florivores and also by the amount of floral tissue removed from the
192 petals or corolla (see Boaventura, et al., 2022), here we measured and categorized levels of
193 florivory as two distinct metrics: 1) florivory incidence is the percentage of flower attack and 2)
194 florivory intensity is the percentage of floral area removed. To assess the incidence of florivory we
195 counted the number of flowers collected from each species (n=100) with any signs of florivory and
196 the number of intact flowers. If, for example, a species had 45 damaged flowers and 55 intact
197 flowers, we consider florivory incidence as 45%. To evaluate the intensity of florivory, all flowers
198 with some signs of florivory, of all studied species, were laminated with the aid of a laminator or
199 transparent adhesive tape and digitized through a digital camera, using a ruler as a default scale for
200 all images. After being digitized, the flowers were measured using ImageJ® software, where all
201 digital images were calibrated to 1.0 cm and the percentage of floral area removed was determined
202 by adding the areas removed from each flower divided by the total area of the flower. Florivory
203 intensity was multiplied by 100 to express florivory as a percentage of floral tissue lost due to
204 florivory (Boaventura et al., 2022).

205 Flower size was determined by measuring at least 20 intact flowers of each plant species
206 (following the same procedures above for measurements of lost floral area) and the floral size per
207 plant species was determined as the average of the total areas of all the flowers measured (cm²).
208

209 *Data analysis*

210

211 Using generalized linear models (GLMs) we evaluated how site, plant and flower
212 functional floral traits affected florivory. We tested whether the degree of urbanization of the
213 collection area (preserved, rural/periurban and urbanized), plant origin at the collection site (native
214 or exotic), plant function type at the collection site (ornamental vs. non-ornamental) flower size,
215 flower color, floral arrangement (flowers in inflorescences vs. individual), floral symmetry
216 (actinomorphic or zygomorphic), and plant growth type (vine, herbaceous, shrub or tree) were
217 factors that could predict and influence florivory levels. Two distinct models were built for each
218 predictor variable for each of the analyzed florivory metrics (florivory incidence and florivory
219 intensity). Additionally, because florivory incidence and intensity are distinct measurements of
220 florivory that can be taken for the same species at the same time, we evaluated whether they could
221 be related in some way. In this last model, florivory incidence was used as the explanatory variable
222 and florivory intensity as the response variable. All models were run considering the negative
223 binomial error distribution due to overdispersion of data and using R packages "MASS" (Venables
224 and Ripley, 2002) and "vegan" (Oksanen et al., 2022).

225 We use phylogenetic generalized least square (PGLS) to tested differences in florivory
226 levels among collection sites, plant origin and function and plant and flower traits to account for
227 phylogenetic relatedness between species and adjusted for expected covariance under a
228 multivariate model of Brownian motion (Felsenstein, 1985; Martins and Hansen, 1997, Symonds
229 and Blomberg, 2014). In these models we adjusted florivory incidence or florivory intensity as
230 response variables and built two distinct models to separately evaluate these two evaluated
231 florivory metrics. We included as fixed effects the type of collection site (rural, urban, natural), the
232 origin of the plant, function of the plant, flower color, flower symmetry, flower arrangement,
233 flower size and plant growth form. Plant species were included as a random effect to account for
234 non-independence among data. The phylogeny of the 61 plant species was reconstructed using the
235 "V.PhyloMaker" (Jin and Qian, 2019), "ape" (Paridis and Schliep, 2019), "nlme" (Pinheiro et al.,

236 2022) and, “phytools” (Revell, 2012). Analyses were done using the "picante" (Kembel et al.,
237 2010) package. All analyzes were performed in the R 4.2.2 software (R Core Team, 2022).

238

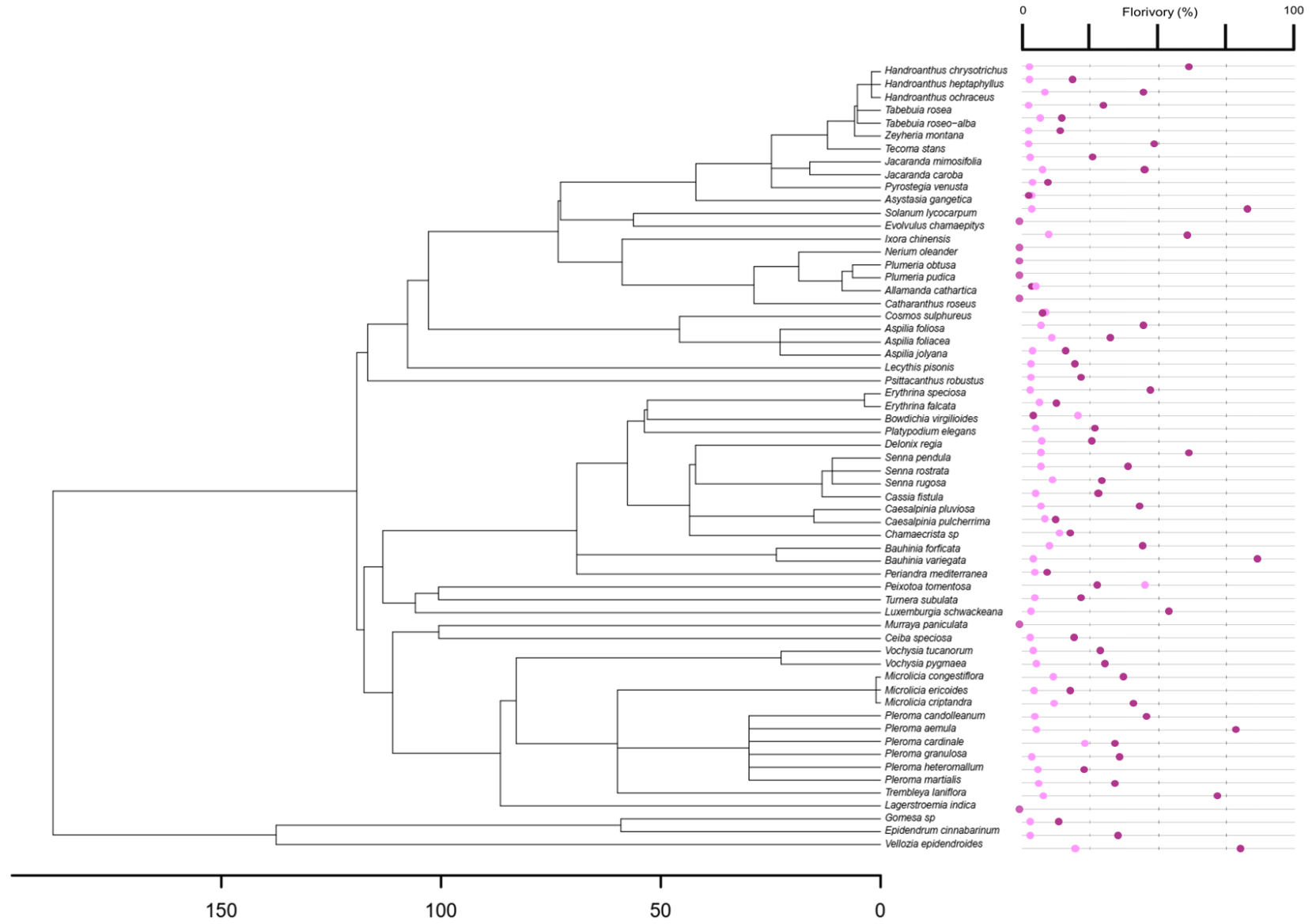
239 **Results**

240

241 *Patterns of florivory across urbanization degrees*

242

243 We evaluated a total of 6,100 flowers from 61 plant species and 20 botanical families
244 collected in areas under different levels of urbanization. Exotic and ornamental plant species were
245 found only in urban or rural/periurban areas. According to the phylogenetic distribution of florivory
246 levels on studied species, both incidence and intensity of floral damage is widespread and variable
247 across plants lineages over different degrees of urbanization (Fig.1).



249 **Fig 1.** Variation of the two estimates of florivory and the phylogenetic relationship of the 61 plant
 250 species evaluated in this study. The pink dots show florivory intensity (percentage of floral area
 251 removed) and the purple dots show florivory incidence (frequency of flower attack) for each sampled
 252 species. The phylogeny was reconstructed using the "V.PhyloMaker" package (Jin and Qian, 2019) in
 253 R software. The scale shown in the tree is thousands of years old.

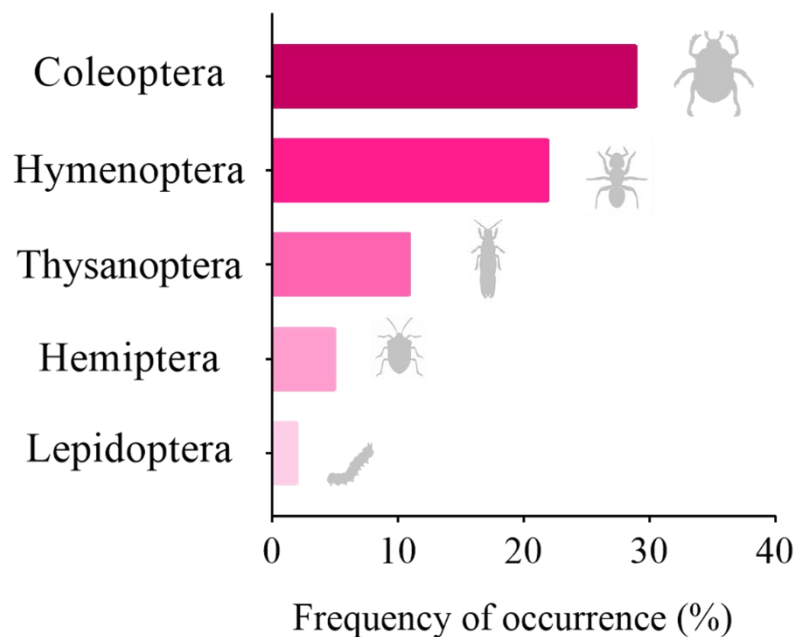
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255

256 Florivory was detected in 89% of the species evaluated here (Fig.1). The seven plant species
 257 (11%) that did not show any sign of florivory were collected in urban or periurban sites and
 258 most of them were exotic and belonged to the Apocynaceae family (n=4). The florivore insects
 259 responsible for floral damage were beetles (Coleoptera) (Fig. 2), mainly from the
 260 Chrysomelidae and Buprestidae families, followed by ants (Hymenoptera) from several tribes
 261 (not yet identified). We also found florivores of the orders Thysanoptera (Thrips), Hemiptera
 262 (bugs) and Lepidoptera (caterpillars) feeding on the floral tissue of the studied species.

263

264



265

266 **Fig 2.** The main groups of florivores responsible for floral damage on the studied plant species in areas
 267 over different degrees of urbanization. Icons indicate each Order of insects.

268

269

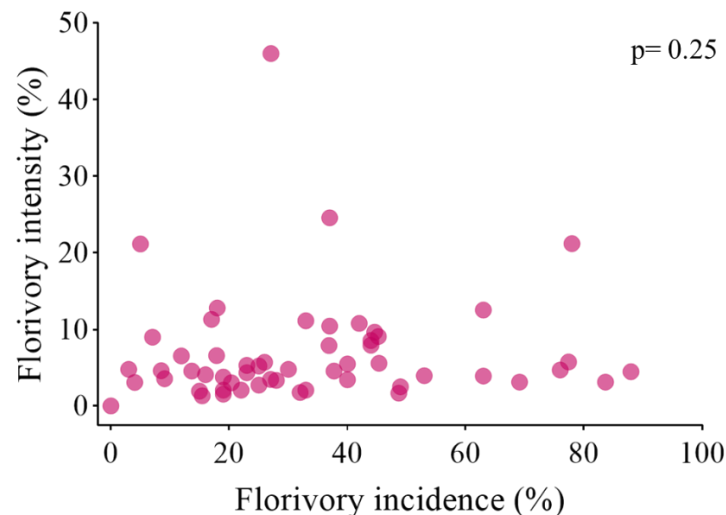
270 About 30.1% ($\pm 2.94SE$) of all evaluated flowers were attacked by florivores (Fig.1).

271 *Solanum lycocarpum* (Solanaceae), a periurban species showed the highest percentage of

272 flowers attacked, as 88% of all flowers (Fig.1) of this species showed some sign of florivory;
 273 *Bauhinia variegata* (Fabaceae), an urban species also had the second highest percentage of
 274 attacked flowers (87%, Fig.1), followed by *Vellozia epidendroides* (Velloziaceae) (78%, Fig.1),
 275 a wild species collected in a natural area. Florivores removed an average of 6.20% (\pm .9SE) of
 276 floral tissue (Fig.1) and all the three most intensely damaged species were collected in natural
 277 areas. *Peixotoa tomentosa* (Malpighiaceae) was the most consumed species, losing about 46%
 278 (Fig.1) of its floral area to florivores, followed by *Pleroma cardinale* (Melastomataceae)
 279 (24.5%, Fig.1), and by *Vellozia epidendroides*, (21.2%, Fig.1). Finally, a higher incidence of
 280 floral damage did not translate into increased intensity of florivory, and our results showed that
 281 these two florivory metrics occurred independently in the studied plants ($r_d = 1.3$, $df = 59$,
 282 $p = 0.25$, Fig.3).

283

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286

287 **Fig 3.** Relationship between the incidence and the intensity of florivory for 61 plant species sampled in
 288 areas over different degrees of urbanization. Dots indicate the average values of each plant sampled.

289

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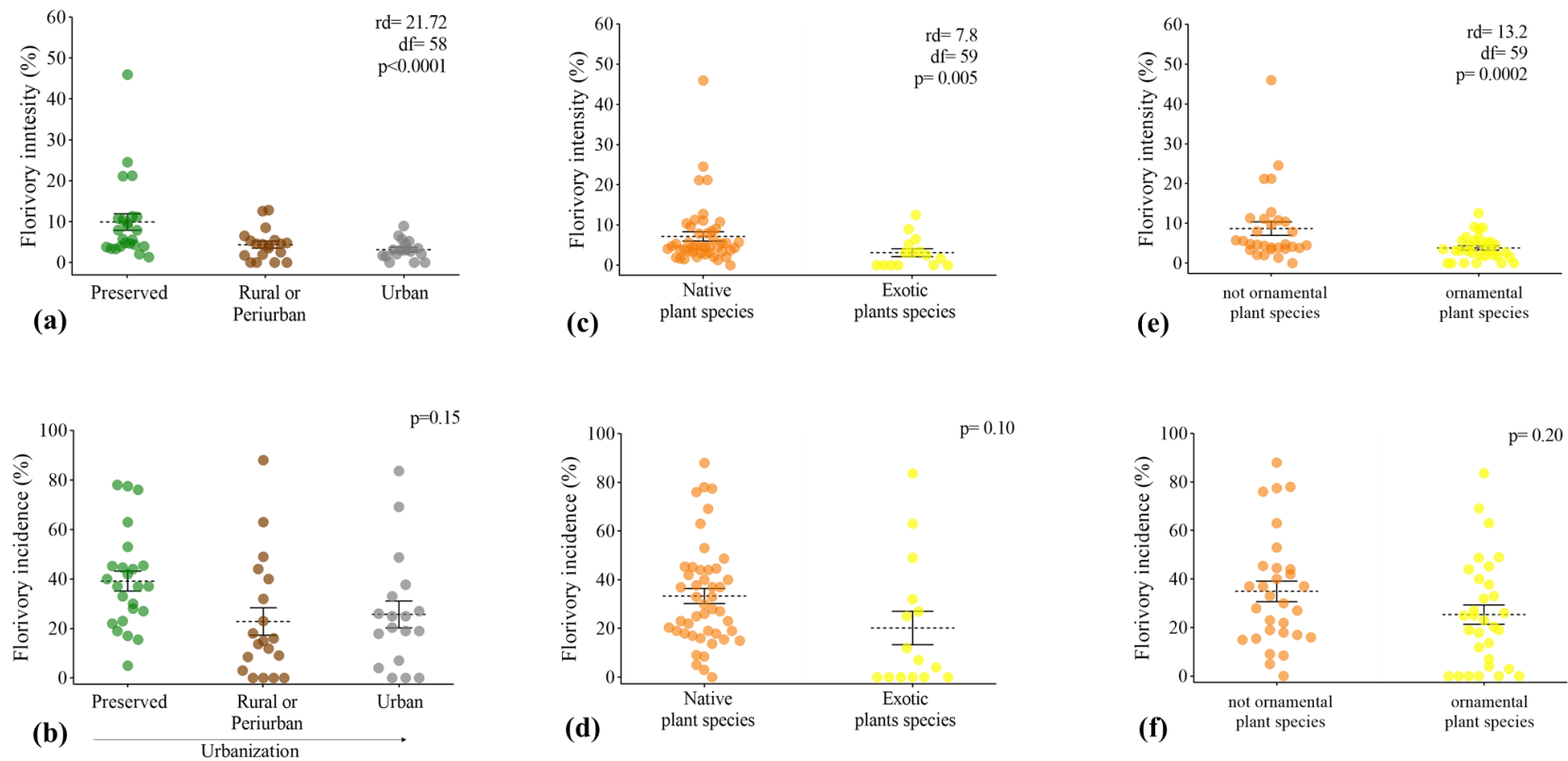
291 *Florivory levels and trait-mediated florivory across urbanization degrees*

292

293 Our results showed that the level of urbanization of the sampling site influenced the
 294 intensity of florivory ($r_d = 21.726$, $df = 58$, $p < 0.0001$) and the levels of floral damage decreased
 295 with the increase of urbanization (Fig.4a). Plants from natural areas showed floral damage

296 intensity three times higher compared to plants collected in urbanized sites ($9.9\% \pm 2SE$, 3.1%
297 $\pm 0.55SE$, respectively) (Fig.4b). The incidence of florivory, despite being higher in plants from
298 preserved sites ($36.1\% \pm 3.9SE$) and lower in plants from periurban/rural ($22.9\% \pm 4.4SE$) and
299 urban ($25.7\% \pm 5.4SE$) sites, was not statistically different among these three groups ($rd=3.75$,
300 $df=58$, $p=0.15$) (Fig.4b), indicating that urbanization does not influence the frequency of floral
301 damage in the studied system.

302 Plant origin (exotic or native) and the type of plant function in the collection site
303 (ornamental or non-ornamental) influenced the intensity of floral damage (origin: $rd=7.8$, df ,
304 59 , $p=0.005$; function: $rd= 13.2$, $df=59$, $p= 0.0002$, Fig 4c-e) and flowers from native and non-
305 ornamental plants were twice as consumed ($7.2\% \pm 1.5SE$ and $8.68\% \pm 1.7SE$, respectively)
306 compared to flowers from exotic and ornamental plants ($3.14\% \pm 0.97SE$ and $3.81\% \pm 0.56SE$,
307 respectively). Although native and non-ornamental plants showed a higher incidence of floral
308 damage ($33.33\% \pm 3SE$ and $34.9\% \pm 4.2SE$, respectively) compared to exotic and ornamental
309 plants ($20.17\% \pm 6.8SE$ and $25.4\% \pm 4SE$, respectively), these differences were not statistically
310 significant (origin: $rd= 2.69$, $df=59$, $p=0.10$; Function: $rd= 1.57$, $df=59$, $p=0.20$; Fig 4d-f).
311



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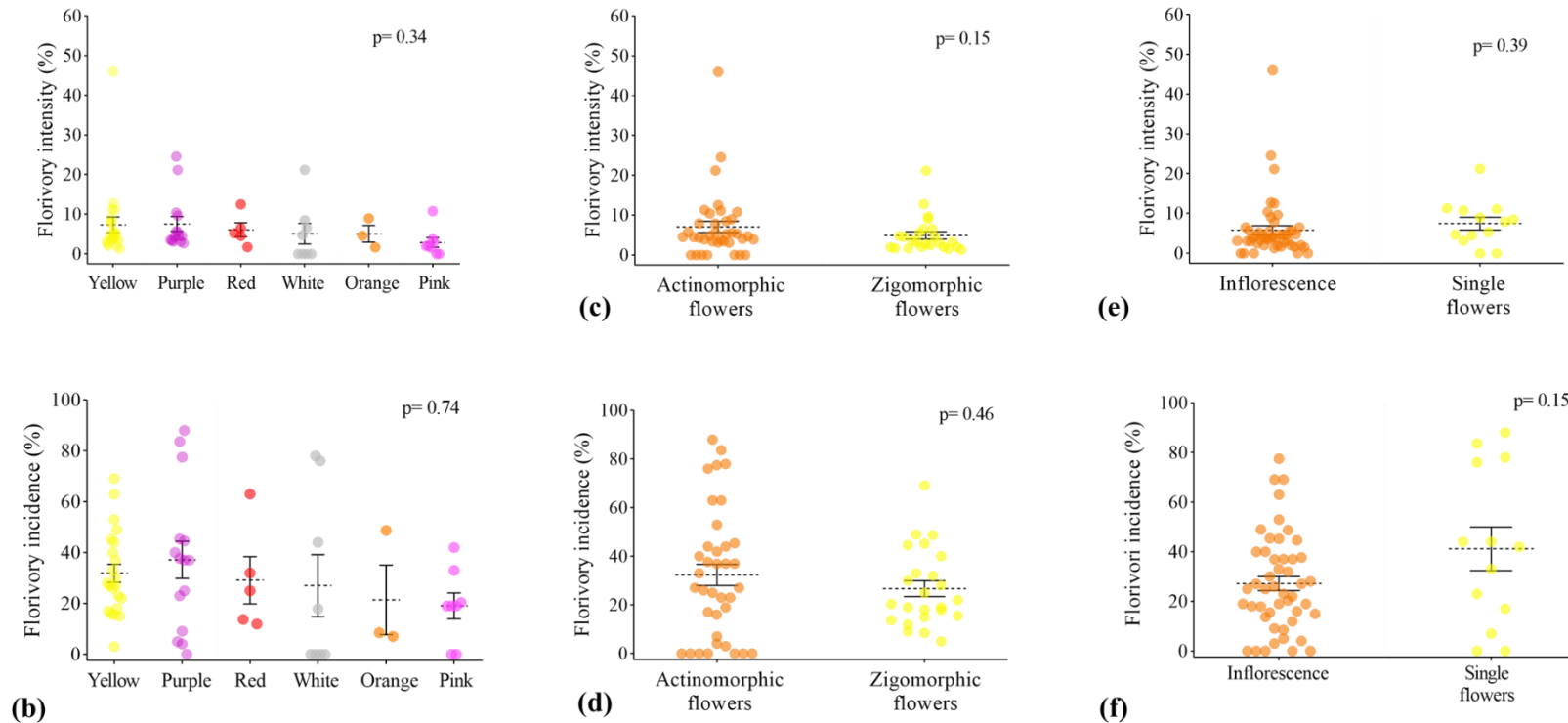
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314 **Fig 4.** Effects of functional traits on florivory metrics (incidence and intensity) in n areas over different degrees of urbanization. Florivory levels were measured
 315 in two different metrics: florivory intensity (percentage of floral area removed) and florivory incidence (percentage of attacked flowers) for all 61 plant species
 316 sampled according to (a,b) degree of urbanization of their collection site, (c,d) plant origin in the site where it was collected and (e,f) type of plant function.
 317 Plants and flowers were sampled in the field using a standardized protocol (see Boaventura et al., 2022). The lines on each panel indicate the mean and the error
 318 bars denote SE.

351 Contrary to our expectations, our results showed that none of the functional traits of the
352 flowers and plants evaluated here influenced florivory levels (Figs 5 and 6). Although floral
353 damage levels varied across plant lineages and species functional traits, our results did not
354 detect any signs that flower size (incidence: $rd= 1.96$, $df=59$, $p=0.16$, intensity: $rd=0.75$, $df=59$,
355 $p=0.38$), flower color (incidence: $rd= 2.73$, $df=55$, $p=0.74$, intensity: $rd=5.63$, $df=55$, $p=0.34$),
356 floral symmetry (incidence: $rd= 0.53$, $df=59$, $p= 0.46$, intensity: $rd=2.02$, $df=59$, $p=0.15$), floral
357 arrangement (incidence: $rd=2$, $df=59$ $p=0.15$, intensity: $rd= 0.72$, $df=59$, $p=0.39$) and plant
358 growth form (incidence: $rd=4.56$, $df=57$, $p=0.20$, intensity: $rd=7.44$, $df=57$, $p=0.058$)
359 influenced the incidence or intensity of florivory.

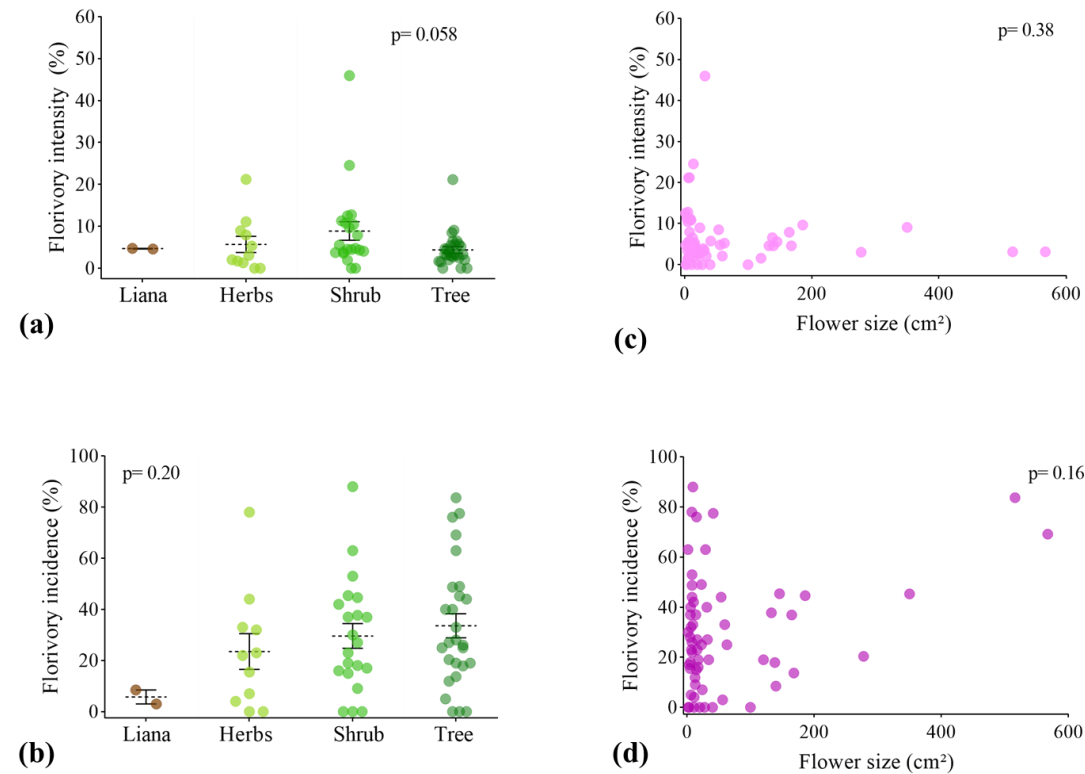
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362

363 **Fig 5.** Functional traits that can influence florivory levels in n areas over different degrees of urbanization. Florivory levels were measured using two different
 364 metrics: florivory intensity (percentage of floral area removed) and florivory incidence (percentage of attacked flowers) for all 61 plant species sampled according
 365 to (a,b) flower color, (c,d) flower symmetry and, (e,f) flower arrangement. Plants and flowers were sampled in the field using a standardized protocol (see
 366 Boaventura et al., 2022). The line indicates the mean, and the error bars denote SE.



367
368

369 **Fig 6.** Functional traits that can influence florivory levels in n areas over different degrees of urbanization. Florivory levels were measured using two different
370 metrics: florivory intensity (percentage of floral area removed) and florivory incidence (percentage of attacked flowers) for all 61 plant species sampled according
371 to (a,b) plant growth form, (c,d) and flower size. Plants and flowers were sampled in the field using a standardized protocol (see Boaventura et al., 2022). The
372 lines on each panel indicate the mean and the error bars denote SE.

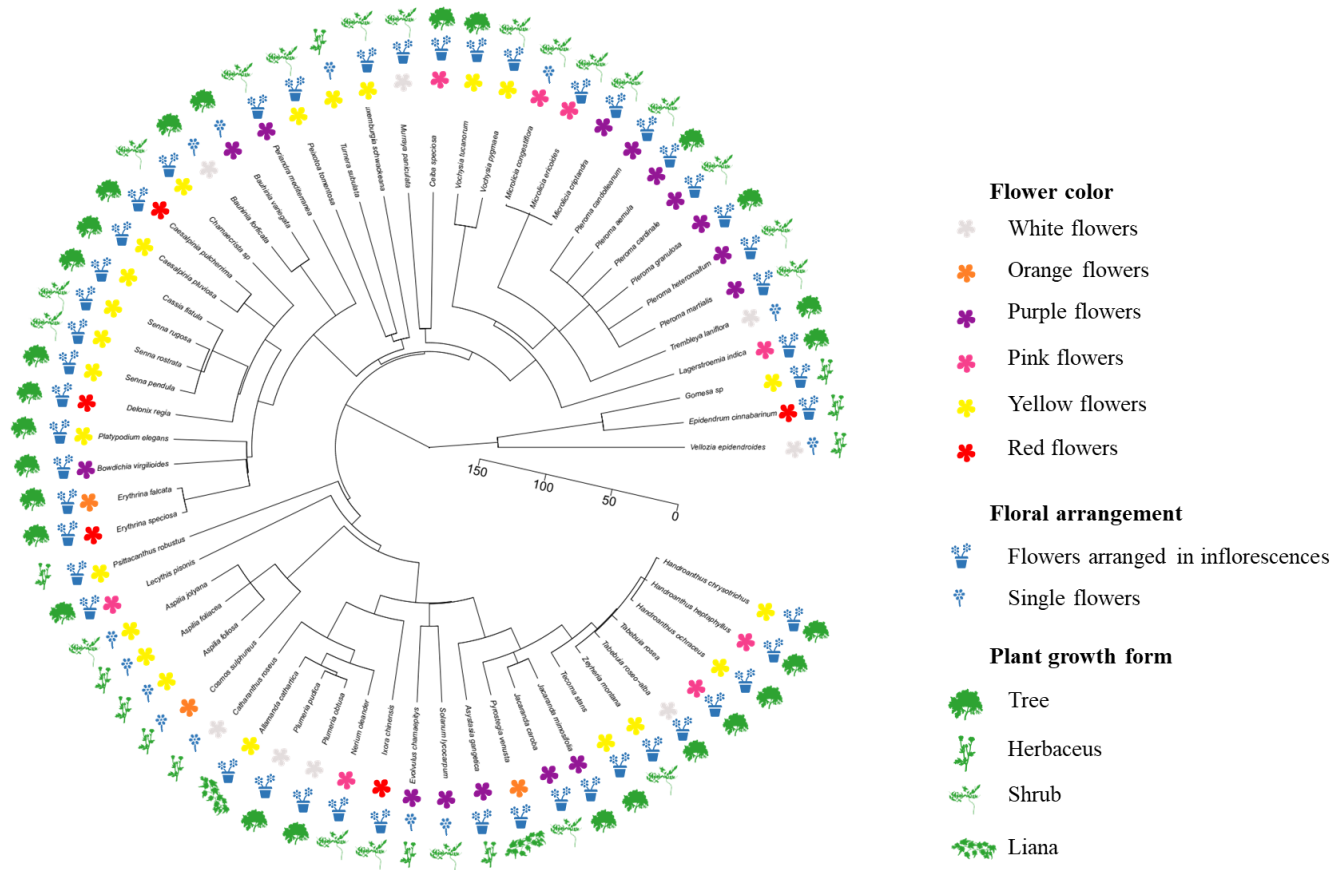
373

374 *Florivory patterns in a phylogenetic context*

375

376 Our results showed that the relationship between levels of florivory and the functional
377 traits of the species changed when we added the weight of phylogeny into our analyses (Fig.7,
378 Tables S1, S2). Considering the phylogeny of the species evaluated here, florivory was
379 positively influenced by plant growth form, flower color, and floral arrangement type and plant
380 relatedness could explain the levels of damage experienced by these species (Fig.7, Tables S1,
381 S2). Flower color and floral arrangement type also influenced the levels of florivory incidence
382 ($F_{5,54} = 3.30$, $p = 0.01$; $F_{1,58} = 6.0$, $p = 0.01$, respectively), and purple and yellow flowers were
383 more frequently attacked by florivores ($37.12\% \pm 7.2SE$ and $31.85\% \pm 3.5SE$, respectively), as
384 well as single flowers ($41.2\% \pm 8.8SE$). Plant growth form, on the other hand, influenced the
385 intensity of florivory ($F_{3,56} = 3.68$; $p = 0.01$), (Tables S1 and S2), with shrubs experiencing the
386 highest levels of floral tissue removal ($8.9\% \pm 2.2SE$), followed by herbaceous plants
387 ($5.7\% \pm 1.9SE$). We found no evidence that phylogeny influenced the relationship between the
388 other variables and levels of florivory.

389



390

391 **Fig 7.** Functional traits of flowers and plants that influence florivory levels when we consider the evolutionary history shared by plant species. When adding the
 392 weight of the phylogeny of species in our analyses, flower color and floral arrangement directly influenced the incidence of florivory, while the plants growth
 393 form influenced the intensity of floral damage. The clipart's around the phylogenetic tree represent the functional traits that influence florivory levels.

394 Discussion

395

396 *Patterns of florivory in areas under different degrees of urbanization*

397

398 This was the first study to evaluate the variation of florivory levels over areas under
399 different degrees of urbanization using multiple plant species sampled with a standardized protocol
400 for collection and evaluation of florivory (Boaventura et al., 2022). We also measured floral
401 damage using two distinct metrics simultaneously. Our results detected florivory in 89% (54 out
402 of 61) of the plant species studied. About 30% of the flowers of these species suffered any kind of
403 floral damage, and florivores, mainly beetles and ants, removed an average of 6.2% of the area of
404 the petals. However, the two measurements of florivory are not related in the same plants nor are
405 influenced by urbanization, as flowers in the cities experienced lower florivory intensity but no
406 effect was detected in florivory incidence. The average of florivory levels found in our study are
407 similar to those previously reported for flowers of several tropical and temperate plant species
408 (24% and 7% for incidence and intensity of florivory, respectively; Boaventura et al., 2022), mainly
409 regarding the intensity of floral damage, despite the florivore community in this study being
410 predominantly composed by caterpillars. However, the average levels of florivory found in our
411 study are about three times lower than those previously reported for *Gelsemium sempervirens* in
412 urban and natural sites (52% and 20% for florivory incidence and intensity, respectively, Irwin et
413 al., 2018). These results suggest that when studying the influence of environmental factors on
414 florivory levels, it is necessary to consider the plant community and evaluate multiple plant species,
415 in order to acquire a realistic picture of the direct and indirect impacts of florivory on plants.
416 Evaluating florivory patterns only at the population level can bring a biased view regarding the
417 magnitude of this interaction, simply because the observed result may come from intraspecific
418 variation in attributes of species that occur naturally in nature and interfere with the interaction
419 with floral enemies, which may or may not be influenced by the external environment. Thus, this
420 study is pioneer in providing a more realistic estimate of the magnitude of the impacts of
421 urbanization on insect-flower antagonisms, especially for tropical plant species, that have never
422 been evaluated in this novel scenario of land use caused by urbanization. Our study has practical
423 implications, as it has been previously shown that there is an indirect interaction between floral
424 antagonists and floral mutualists (McCall and Irwin, 2006; Moreira et al., 2019b) and flowers

425 attacked by insects tend to be less visited by pollinators (Krupnick et al., 1999; McCall, 2008; Liao
426 et al., 2013), with potential negative effects of florivory on plant reproduction via reduced attraction
427 to other beneficial insects such as bees and butterflies. Recent studies (e.g., Theodorou et al., 2020;
428 Wenzel et al., 2020; Ayers and Rehan, 2021) have indicated that pollination is an ecosystem service
429 that has been reduced in several cities around the world due to the negative impacts of urbanization
430 in the populations of pollinators and our data shows lower levels of florivory intensity in the cities
431 but also indicates that most plants studied have signs of florivore damage, which can translate into
432 lowered fitness in urbanized areas where florivores - such as ants and beetles - are present.

433

434 *Processes that drive florivory levels in areas under different degrees of urbanization*

435

436 The results of our study support the hypothesis that urbanization influences florivory levels,
437 but only for florivory intensity. Consistent with our expectations, flowers collected from natural
438 sites had a higher intensity of floral damage and were consumed up to three times more by
439 florivores compared to flowers from urban sites. Our results do not support the findings of previous
440 studies that reported an increase in florivory levels due to urbanization (see Irwin et al., 2014; 2018)
441 and provide a new perspective on the influence of environmental changes on the dynamics of
442 insect-flower antagonisms. The influence of urbanization on insect-plant antagonisms has been
443 evaluated in previous studies (see Nuckols and Connor, 1995; Bode and Gilbert, 2016; Raupp et
444 al., 2010; Cuevas-Reyes et al., 2013; Kozlov et al., 2017; Moreira et al., 2019a; Stemmelen et al.,
445 2020). For foliar herbivory, for example, some of these studies have already pointed to a similar
446 pattern to what was found here; a marked decrease in leaf damage levels due to urbanization
447 (Nuckols and Connor, 1995; Bode and Gilbert, 2016; Kozlov et al., 2017; Moreira et al., 2019a).
448 Explanations for these patterns found include both a decrease in the abundance and diversity of
449 insects, as well as changes in the secondary chemistry composition of the plants and leaves
450 (Nuckols and Connor, 1995; Kozlov et al., 2017; Moreira et al., 2019a).

451 For florivory, the results of our study indicated that the patterns of intensity of flower
452 consumption found in areas under different degrees of urbanization are related to the host plants
453 attributes associated with urbanization, corroborating our second prediction. Our results showed
454 that the origin of the species is determinant for the levels of floral damage, and native plants
455 experienced levels of florivory about twice as high than exotic plants, which would partially

456 explain the patterns of damage found. One of the reasons exotic plants suffered lower levels of
457 floral damage might be associated to the fact that these species sustain a lower abundance and
458 diversity of insects because they provide a low-quality food resource, making flowers less attractive
459 (Fenoglio et al., 2020;2021). In an experimental study that evaluated the herbivore community on
460 native and exotic plant species, a decrease in the abundance of chewing insects (which were the
461 main florivores) was reported in exotic plants, especially if those species were not phylogenetically
462 close to local host species, which confirms our assumptions (Burghardt and Tallamy, 2013). In our
463 study, six of the seven plant species in which we did not detect any signs of floral damage are
464 exotic, and most of them (n=4) belong to the Apocynaceae family, which is widely known for its
465 high toxicity due to the presence of glycosides and alkaloids in all parts of the plants, but mainly
466 in the flowers (Larrea et al., 2014). Therefore, we believe that florivores neglected the flowers of
467 exotic species because they offer low-quality food resources, that can often become toxic and
468 unpalatable for them.

469 In addition to the origin of the species, another attribute related to urbanization that
470 influenced the intensity of floral damage was the fact that the species were used for urban
471 ornamentation. Ornamental species are those artificially planted in cities and rural gardens for
472 urban afforestation and ornamentation and can be native or non-native to that region (Nuckols and
473 Connor, 1995; Smith et al., 2006; Potgieter et al., 2017). In our study, ornamental species
474 experienced levels of florivory about two times lower than non-ornamental species. Similar results
475 have been reported by Nuckols and Connor (1995) when assessing levels of foliar herbivory for
476 the same tree species in artificially planted gardens and natural environments. These findings
477 contradict previous assumptions that plants in urban environments are more susceptible to attack
478 by natural enemies (Dreistadt et al., 1990; Watson and Ware, 1994; Cuevas-Reys et al., 2013) and
479 indicate that the environmental stress caused by increased temperature, decreased availability of
480 water and nutrients, and increased CO₂ emissions in cities may mediate physiological changes in
481 these plants, making them less attractive to florivores. Previous studies have shown, for example,
482 that herbivores tend to avoid trees with water deficit (Hanks and Denno, 1993) and that host plants
483 growing under high CO₂ conditions have low nutritional quality for antagonists due to a decrease
484 in nitrogen concentration and an increase in tannin concentration in leaves (Stiling and Cornelissen,
485 2007). In fact, it has already been demonstrated that florivores preferentially consume flowers with
486 higher concentrations of nitrogen and water (Ogura and Sakai, 2015). Although we have not yet

487 evaluated physiological parameters of flowers, we believe that flowers may have decreased
488 nitrogen concentrations and water in response to stress, as already demonstrated for leaves.
489 Additionally, the identity and abundance of florivores may also influence the patterns of damage
490 found in ornamental and exotic species, and this should be addressed in future studies.

491 Although the incidence of florivory was higher in flowers from natural sites, native and
492 non-ornamental plants, we found that this florivory metric was not influenced by urbanization.
493 Florivory incidence has been the most widely used measure to estimate floral damage (see
494 Boaventura et al., 2022), and although it assesses the frequency with which flowers are attacked in
495 populations and communities, this measure seems unable to represent the potential that
496 environmental conditions must interfere with plant susceptibility to natural enemies. In contrast,
497 the amount of floral tissue removed (florivory intensity) proved to be a much more effective
498 measure for assessing levels of florivory, as in our study, because it indicates the degree of
499 susceptibility to which plants are subjected to damage and varied consistently with different
500 urbanization degrees. Thus, we believe that the evaluation of the magnitude of floral damage,
501 represented by florivory intensity, is a much more reliable measure to estimate levels of florivory
502 in flowers, and we recommend that future studies preferably use this measure to avoid bias and
503 underrepresentation of floral herbivory.

504

505 *Trait-mediated florivory in areas under different degrees of urbanization*

506

507 In addition to plant attributes related to urbanization, our study has found that levels of
508 florivory were also influenced by plant functional traits, although this influence was only detected
509 when we added the weight of the phylogenetic relationship among the species, supporting our third
510 hypothesis. As suggested by Boaventura et al. (2022), floral and vegetative traits inherited through
511 plant lineages - which are often only studied in the lieu of pollinator attraction - can also attract
512 florivores and consequently influence levels of florivory. Flower size (see Gallen, 1999; Oguro and
513 Sakai, 2015; Gélvez-Zúñiga et al., 2018), corolla symmetry (McCall, 2008; Boaventura et al.,
514 2022), flower color (Reverté et al., 2016; McCall et al., 2013), floral arrangement (Ortiz et al.,
515 2023; Boaventura et al., 2022), and even plant growth form (Boaventura et al., 2018) have been
516 reported in the literature as traits that can attract both pollinators and floral enemies and impact the
517 dynamics of these interactions. According to the Plant Apparency Hypothesis (PAH, Feeny, 1976),

518 when any plant trait makes both flowers and species more conspicuous, these attributes could
519 contribute to an increase in florivory levels, and plants may adopt strategies to escape florivory by
520 making flowers less visible and/or less attractive (Boaventura et al., 2022).

521 In our study we found that flower color influenced levels of florivory incidence, mainly
522 because purple and yellow flowers that are phylogenetically closely showed similar levels of
523 florivory incidence (see Fig.7). We now know that plant pigments, which give color to flowers,
524 have the potential to influence florivory levels because, like pollinators, florivores can use the
525 flower color as a cue to locate flowers and consume floral tissue, since these pigments are often
526 related to biosynthetic and metabolic pathways that may have defensive properties (Boaventura et
527 al., 2022). In plant species that have color polymorphism, for example, it has been demonstrated
528 that florivores such as caterpillars, consume yellow and/or white flowers more intensely than pink
529 and purple flowers because they have lower concentrations of anthocyanins, which are one of the
530 main secondary defense compounds of flowers (Irwin et al., 2003; McCall et al., 2013). Here, the
531 secondary chemistry related to flower color does not seem to have influenced the incidence of
532 damage, as the most frequently attacked species have purple flowers, which theoretically would be
533 better protected against natural enemies (Irwin et al., 2003). We suggest that the mechanisms that
534 determine florivory incidence are different from those that determine the intensity of flower
535 consumption and, florivores may use petal color as a cue to locate resources, making certain species
536 more conspicuous and attractive to them (PAH, Feeny, 1976). In our study, the appearance of
537 flowers may be related to their longevity, and purple flowers, for example, may have been more
538 frequently attacked simply because they have a longer exposure time compared to flowers of other
539 morphs. Many species of the genus *Pleroma* (synonymy: *Tibouchina*), which represent most of the
540 purple flowers in our study, produce longer-lasting flowers and have a more prolonged flowering
541 period (*personal observation*), which would increase the chances that these plants are found and
542 attacked by natural enemies, which would help explain the results found here.

543 The type of floral arrangement influenced the incidence of florivory and flowers singly-
544 arranged experienced higher levels of florivory compared to flowers arranged into inflorescences.
545 These findings are consistent with the escape hypothesis proposed by the Plant Appearance
546 Hypothesis (PAH, Feeny, 1976) and suggest that plants that are visually more conspicuous - such
547 as those with flowers arranged in dense resources patches or in inflorescences - are "bound to be
548 found" and therefore may escape florivory by presenting a higher concentration of physical and/or

549 chemical defenses, making them less attractive to florivores, as happens with leaves (Smilanich et
550 al., 2016, Boaventura et al., 2022). We believe that although single flowers are visually less
551 conspicuous, in our study they were more frequently attacked because they have a lower investment
552 in defenses against natural enemies, as previously suggested by Boaventura et al. (2022), and
553 potentially present a better nutritional quality for florivores, since more than two-thirds of plants
554 with solitary flowers were native species (78%) or non-ornamental (61%).

555 Plant growth form, which can also make plants conspicuous (PAH, Feeny, 1976) was the
556 only functional trait that influenced the intensity of florivory considering species phylogeny.
557 Shrubs and herbaceous plants, despite being visually less apparent, showed the highest levels of
558 floral consumption compared to trees and lianas. The plant growth form has been widely evaluated
559 as a functional trait capable of influencing herbivory levels in plants, and there is some consensus
560 that trees suffer more intense levels of damage because they are apparent, compared to other life
561 forms (see Feeny, 1976; Wright et al., 2004; Kozlov et al., 2015). For florivory, we found the
562 opposite pattern, and we suggest that the intensity of consumption is related to some other process
563 that is not the visual appearance itself, but that makes shrubs and herbaceous plants more attractive
564 and apparent to florivores, such as the nutritional quality of flowers. The fact that most shrub and
565 herbaceous species are not ornamental (81% and 54%, respectively) and are native (85% and 63%,
566 respectively) would support our hypothesis that the nutritional quality in these plants would be
567 superior and function as an attractive to florivores, more important than the physical appearance of
568 the plant itself. Furthermore, the fact that most tree species were collected in urban and rural areas
569 (74%) may also help explain our results because beetles, our main florivores, are one of the groups
570 most affected by anthropization (see Fenoglio et al., 2020) and are experiencing a decline in
571 abundance and diversity in urbanized areas. Therefore, lower levels of florivory may also have
572 been detected in trees due to the decline in abundance and diversity of florivores and this point
573 deserves further evaluation and the identification of flower consumers and diet breadth.

574 We did not find evidence that floral size or symmetry influence florivory levels in the
575 evaluated plants, although these functional traits have recently been reported as potentially relevant
576 for attracting florivores and influencing floral damage levels (see Boaventura et al., 2022). Several
577 studies have even reported that large flowers are often more attacked and consumed by florivores
578 (Ogura and Sakai, 2015, Ortiz et al., 2023), as well as flowers with radial symmetry (Boaventura,
579 2018), but in our study, attributes more related to floral attraction and flower quality were more

580 relevant in determining florivory levels than floral size itself. We believe that the importance of
581 functional traits for florivory levels may vary depending on environmental variables and the
582 identity of the evaluated species. Therefore, we emphasize the importance of evaluating florivory
583 for as many species and environmental contexts as possible.

584

585 **Conclusion**

586

587 Our study described how the patterns of florivory vary with an urbanization degree in
588 tropical areas based on the functional and phylogenetic traits of the studied plant species. We
589 provided one of the first assessments of florivory at the community level of species and through a
590 new approach, to understand the combined effects of biotic and abiotic factors that influence the
591 incidence and intensity of florivory, in the context of environmental changes that occur because of
592 urbanization. We found that the florivory levels reported here are consistent with global estimates
593 of floral damage for species collected around the world (Boaventura et al., 2022). We suggest that
594 these findings are due to the use of standardized methods for collecting and evaluating levels of
595 florivory in our study and therefore recommend the use of this standardized protocol in future
596 studies. Our study demonstrated that florivory intensity decreased with increasing urbanization,
597 due to the substitution of native species to exotic ones and because of the environmental stress
598 experienced by ornamental plants in cities, which leads to a decrease in attractiveness of these
599 species to florivores. Florivory incidence was not influenced by urbanization, and we emphasize
600 that this method of evaluating floral damage is not as accurate as determining the amount of floral
601 tissue removed (florivory intensity). We also found that functional traits of species and flowers,
602 such as flower arrangement, flower color, and plant growth form, influenced levels of florivory,
603 but this influence is only detectable when we added the weight of the evolutionary history shared
604 by species in our analyses. Thus, we highlight the need to consider the phylogeny of species in
605 future studies. Overall, our study demonstrated that the effect of urbanization and functional traits
606 on florivory levels are mostly mediated by changes in the occurrence of host plants, which vary
607 depending on the plant origin (native or exotic) and the type of ecological function it has
608 (ornamental or non-ornamental). As most studies on antagonistic interactions in the context of
609 urbanization have been about foliar herbivory, we still do not know which secondary compounds
610 would lead to this observed reduction in flower quality, and we highlight the need for these

611 mechanisms to be evaluated in the future. We believe that the identity, abundance, and richness of
612 florivores in the collection areas may have also contributed to the results found here and suggest
613 that future studies on insect-flower antagonisms in cities jointly evaluate the florivore community
614 and levels of florivory. Finally, we emphasize the need for studies that evaluate the impact of
615 florivory on plants reproductive success in areas under different degrees of urbanization because,
616 even though floral damage levels in cities are lower than previously thought, they could still
617 negatively impact plant fitness, especially in a scenario of pollinator scarcity.

618

619 **Acknowledgements**

620

621 We thank UFMG, CSEC, PPG-ECMVS and CAPES (Financial code 001) for continuous support.
622 We thank all colleagues who helped us in many ways. Our colleagues of CSEC that helped on
623 plants and insects' collections. We thank FAO Silveira and M Zanetti for support with plant species
624 identification. We thank ES Diniz for assistance with PGLS analyses. This study was partially
625 financed by the Coordenação de Aperfeiçoamento de Pessoal de Nível Superior – Brasil (CAPES)
626 – Finance Code 001 and by CNPq grant 313007-2020/9 awarded to T Cornelissen.

627

628 **References**

- 629 Ayers AC, Rehan SM. (2021). Supporting bees in cities: how bees are influenced by local and
630 landscape features. *Insects* 12:128.
- 631 Boaventura MG. (2018). Padrões de florivoria e herbivoria em plantas tropicais e redes de
632 interações tróficas associadas. *Dissertação de mestrado* apresentado ao Programa de Pós-
633 graduação em Ecologia, departamento de Ciências Naturais, Universidade Federal de São João
634 Del-Rei.
- 635 Boaventura MG, Villamil N, Teixido AL, Tito R., Vasconcelos HL, Silveira FAO, Cornelissen T.
636 (2022). Revisiting florivory: an integrative review and global patterns of a neglected interaction.
637 *New Phytologist*, 233:132-144.
- 638 Bode RF, Gilbert AB. (2016). Seed predators, not herbivores, exert natural selection on *Solidago*
639 spp. in an urban archipelago. *Environmental Entomology*, 45:150–154.
- 640 Bronstein JL, Huxman, TE, Davidowitz G. (2007). Plant-mediated effects linking herbivory and
641 pollination. In: Ohgushi T, Craig TG, Price PW eds. *Ecological Communities: Plant Mediation in*
642 *Indirect Interaction Webs*. Cambridge, UK: Cambridge *University Press*, 79–103.
- 643 Burghardt KT, Tallamy DW. (2013). Plant origin asymmetrically impacts feeding guilds and life
644 stages driving community structure of herbivorous arthropods. *Diversity and Distributions*,
645 19:1553–1565.
- 646 Caruso CM, Eisen KE, Martin RA, Sletvold N. (2019). A meta-analysis of the agents of selection
647 on floral traits. *Evolution*, 73: 4–14.
- 648 Cornelissen T, Lourenço GM, Costa FV, Boaventura MG, Pignataro T, Lima J, Vieira S, Dolabella
649 B. (2023). Insects in the city: patterns of biodiversity, interactions, and ecosystem services in urban
650 green areas. In: Angeoletto F, Tryjanowski P, Fellowes M. eds. *Ecology of Tropical Cities: Natural*
651 *and Social Sciences Applied to the Conservation of Urban Biodiversity*. *Springer nature*, Springer,
652 *under review*.
- 653 Cuevas-Reyes P, Gilberti L, González-Rodríguez A, Fernandes GW. (2013). Patterns of herbivory
654 and fluctuating asymmetry in *Solanum lycocarpum* St. Hill (Solanaceae) along an urban gradient
655 in Brazil. *Ecological Indicators*, 24:557–561.
- 656 Dreistadt SH, Dahlsten DL, Frankie GW. (1990). Urban forests and insect ecology. *BioScience*,
657 40:192-198.
- 658 Elmqvist T, Fragkias M, Goodness J, Güneralp B, Marcotullio PJ, McDonald RI, Parnell S,
659 Schewenius M, Sendstad M, Seto KC, Wilkinson C. (2013). *Urbanization, Biodiversity and*
660 *Ecosystem Services: Challenges and Opportunities*. London: *Springer*.
- 661 Feeny P. (1976). Plant apparency and chemical defense. In: JW Wallace, RL Mansell, Eds.
662 *Biochemical interaction between plants and insects*. Boston, MA, USA: *Springer US*, 1– 40.

- 663 Felsenstein J. (1985). Phylogenies and the comparative method. *American Naturalist*, 125: 1–15.
- 664 Fenoglio MS, Calviño A, González E, Salvo A, Videla M. (2021). Urbanization drivers and
665 underlying mechanisms of terrestrial insect diversity loss in cities. *Ecological Entomology*, 46:757-
666 771.
- 667 Fenoglio MS, Rossetti MR, Videla M. (2020). Negative effects of urbanization on terrestrial
668 arthropod communities: A meta-analysis. *Global Ecology and Biogeography*, 29:1412–1429.
- 669 Galen C. (1999). Why do flowers vary? The functional ecology of variation in flower size and form
670 within natural populations. *BioScience*, 49: 631–640.
- 671 Gélvez-Zúñiga I, Teixido AL, Neves ACO, Fernandes GW. (2018). Floral antagonists counteract
672 pollinator-mediated selection on attractiveness traits in the hummingbird-pollinated *Collaea*
673 *cipoensis* (Fabaceae). *Biotropica*, 50: 797–804.
- 674 Gélvez-Zúñiga I. (2021). Disentangling the intensity and consequences of floral antagonists in a
675 threatened mountain top ecosystem. *Tese de doutorado* apresentada ao Programa de pós-graduação
676 em ecologia, conservação e manejo da vida silvestre, Instituto de Ciências Biológicas,
677 Universidade Federal de Minas Gerais.
- 678 Hanks LM, Denno RF. (1993). Natural enemies and plant water relations influence the distribution
679 of an armored scale insect. *Ecology*, 74:1081-1091.
- 680 Irwin RE, Strauss SY, Storz S, Emerson A, Guibert G. (2003). The role of herbivores in the
681 maintenance of a flower color polymorphism in wild radish. *Ecology*, 84:1733-1743.
- 682 Irwin, RE, Warren PS, Adler LS. (2018). Phenotypic selection on floral traits in an urban landscape.
683 *Proceedings of the Royal Society B: Biological Sciences*, 285:20181239.
- 684 Irwin RE, Warren PS, Carper AL, Adler LS. (2014). Plant–animal interactions in suburban
685 environments: implications for floral evolution. *Oecologia*, 174:803–815.
- 686 Jin Y, Qian H. (2019), V.PhyloMaker: an R package that can generate very large phylogenies for
687 vascular plants. *Ecography*, 42: 1353-1359.
- 688 Johnson MTJ, Munshi-South J. (2017). Evolution of life in urban environments. *Science*, 358:1–
689 11.
- 690 Kembel SW, Cowan PD, Helmus MR, Cornwell WK, Morlon H, Ackerly DD, Blomberg SP, Webb
691 CO. (2010). Picante: R tools for integrating phylogenies and ecology. *Bioinformatics*, 26:1463-
692 1464.
- 693 Kondratyeva A, Knapp S, Durka W, Kühn I, Vallet J, Machon N, Martin G, Motard E, Grandcolas
694 P, Pavoine S. (2020). Urbanization effects on biodiversity revealed by a two-scale analysis of
695 species functional uniqueness vs. redundancy. *Frontiers in Ecology and Evolution*, 8:73.

- 696 Kozlov M, Lanta V, Zverev V, Zvereva E. (2015). Background losses of woody plant foliage to
697 insects show variable relationships with plant functional traits across the globe. *Journal of Ecology*,
698 103: 1519–1528.
- 699 Kozlov MV, Lanta, V, Zverev, V, Rainio, K, Kunavin, MA, Zvereva, EL. (2017). Decreased losses
700 of woody plant foliage to insects in large urban areas are explained by bird predation. *Global*
701 *Changes Biology*, 23:4354-4364.
- 702 Krupnick GA, Weis AE, Campbell DR. (1999). The consequences of floral herbivory for pollinator
703 service to *Isomeris arborea*. *Ecology*, 80:125– 134.
- 704 Larrea MISA, M.D, Larrea SA, Fernández CR. (2014). Plants, poisonous (Animals). Eds, Philip
705 Wexler. *Encyclopedia of Toxicology*. Third Edition, *Academic Press*, Oxford. 960-969.
- 706 Liao K, Gituru RW, Guo YH, Wang QF. (2013). Effects of floral herbivory on foraging behaviour
707 of bumblebees and female reproductive success in *Pedicularis gruina* (Orobanchaceae). *Flora –*
708 *Morphology, Distribution, Functional Ecology of Plants*, 208: 562– 569.
- 709 Malo JE, Leirana-Alcocer J, Parra-Tabla V. (2001). Population fragmentation, florivory, and the
710 effects of flower morphology alterations on the pollination success of *Myrmecophila tibicinis*
711 (Orchidaceae). *Biotropica*, 33:529-534.
- 712 Martins EP, Hansen TF. (1997). Phylogenies and the comparative method: a general approach to
713 incorporating phylogenetic information into the analysis of interspecific data. *American Naturalist*,
714 149: 646–667.
- 715 McCall A. (2008). Florivory affects pollinator visitation and female fitness in *Nemophila menziesii*
716 *Nemophila menziesii*. *Oecologia*, 155: 729– 737.
- 717 McCall AC, Irwin RE. (2006). Florivory: the intersection of pollination and herbivory. *Ecology*
718 *letters*, 9:1351-1365.
- 719 McCall A, Murphy S, Venner C, Brown M. (2013). Florivores preferem morfos de cor de pétala
720 branca versus rosa em rabanete selvagem *Raphanus sativus*. *Oecologia*, 172: 189– 195.
- 721 Mendes GM, Silveira FAO, Oliveira C, Dattilo W, Guevara R, Ruiz-Guerra B, Boaventura MG, et
722 al. (2021). How much leaf area do insects eat? A data set of insect herbivory sampled globally with
723 a standardized protocol. *Ecology*, 102: e03301.
- 724 Moreira, X, Abdala-Roberts L, Teran JCBM, Covelo F, Mata R, Francisco M, Hardwick B, et
725 altem. (2019a). Impacts of urbanization on insect herbivory and plant defences in oak trees. *Oikos*,
726 128:113–123.
- 727 Moreira X, Castagneyrol B, Abdala-Roberts L, Traveset A. (2019b). A meta-analysis of herbivore
728 effects on plant attractiveness to pollinators. *Ecology*, 100: e02707.
- 729 Nuckols MS, Connor EF. (1995). Do trees in urban or ornamental plantings receive more damage
730 by insects than trees in natural forests? *Ecological Entomology*, 20, 253–260.

- 731 Oguro M, Sakai S. (2015). Relation between flower head traits and florivory in Asteraceae: a
732 phylogenetically controlled approach. *American Journal of Botany*, 102: 407–416.
- 733 Ortiz GL, Columbano Y, Melo, MV, Boaventura MG, Aoki C, Cornelissen T, Souza CS, Teixido
734 AL. (2023). Among-species variation in flower size determines florivory in the largest seasonally
735 flooded tropical wetland. *American Journal of Botany*, under review.
- 736 Oksanen J, Simpson G, Blanchet F, Kindt R, Legendre P, Minchin P, O’Hara R, et al. (2022).
737 `_vegan: Community Ecology Package_`. R package version 2.6-4, <[https://CRAN.R-](https://CRAN.R-project.org/package=vegan)
738 [project.org/package=vegan](https://CRAN.R-project.org/package=vegan)>.
- 739 Paradis E, Schliep K. (2019). Ape 5.0: an environment for modern phylogenetics and evolutionary
740 analyses in R. *Bioinformatics*, 35: 526-528.
- 741 Pinheiro J, Bates D, R Core Team (2022). `_nlme: Linear and Nonlinear Mixed Effects Models_`. R
742 package version 3.1-160, <<https://CRAN.R-project.org/package=nlme>>.
- 743 Potgieter LJ, Gaertner M, Kueffer C, Larson BM, Livingstone SW, O’Farrell PJ, et al. (2017).
744 Alien plants as mediators of ecosystem services and disservices in urban systems: a global review.
745 *Biological Invasions*, 19:3571–3588
- 746 Raupp MJ, Shrewsbury PM, Herms DA. (2010). Ecology of herbivorous arthropods in urban
747 landscapes. *Annual Review of Entomology*, 55:19–38.
- 748 Rusman Q, Lucas-Barbosa D, Poelman EH, Dicke M. (2019). Ecology of plastic flowers. *Trends*
749 *in Plant Science*, 24: 725–740.
- 750 Revell LJ. (2012). Phytools: An R package for phylogenetic comparative biology (and other
751 things). *Methods Ecology and Evolution*, 3:217-223.
- 752 Reverté S, Retana J, Gómez JM, Bosch J. (2016). Pollinators show flower colour preferences but
753 flowers with similar colours do not attract similar pollinators. *Annals of Botany*, 118: 249– 257.
- 754 Smilanich A, Fincher R, Dyer L. (2016). Does plant apparency matter? Thirty years of data provide
755 limited support but reveal clear patterns of the effects of plant chemistry on herbivores. *New*
756 *Phytologist*, 210: 1044– 1057.
- 757 Smith RM, Thompson K, Hodgson JG, Warren PH, Gaston KJ. (2006) Urban domestic gardens
758 (IX): composition and richness of the vascular plant flora, and implications for native biodiversity.
759 *Biological Conservation*, 129:312–322.
- 760 Solga, Mj, Harmon, JP, Ganguli, AC. (2014). Timing is Everything: An Overview of Phenological
761 Changes to Plants and Their Pollinators. *Natural Areas Journal*, 34:227-234.
- 762 Stemmelen A, Paquette A, Benot ML, Kadiri Y, Jactel H, Castagneyrol B. (2020). Insect herbivory
763 on urban trees: Complementary effects of tree neighbours and predation. *BioRxiv*, 2:e22.

- 764 Stiling P., Cornelissen T. (2007). How does elevated carbon dioxide (CO₂) affect plant–herbivore
765 interactions? A field experiment and meta-analysis of CO₂-mediated changes on plant chemistry
766 and herbivore performance. *Global Change Biology*, 13: 1823–1842.
- 767 Strauss SY, Conner JK, Rush SL. (1996). foliar herbivory affects floral characters and plant
768 attractiveness to pollinators: implications for male and female plant fitness. *The American*
769 *Naturalist*, 147:1098-1107.
- 770 Symonds M, Blomberg S. (2014). A primer on phylogenetic generalized least squares. In:
771 Garamszeg, L. Ed. Modern phylogenetic comparative methods and their application in
772 evolutionary biology: concepts and practice. Berlin, Heidelberg: *Springer*, 105–130.
- 773 Teixido AL, Fuzessy LF, Souza CS, Gomes IN, Kaminski LA, Oliveira CP, Maruyama PK. (2022).
774 Anthropogenic impacts on plant-animal mutualisms: A global synthesis for pollination and seed
775 dispersal. *Biological Conservation*, 266:109461.
- 776 Theodorou P, Radzevičiūtė R, Lentendu G. et al. (2020) Urban areas as hotspots for bees and
777 pollination but not a panacea for all insects. *Nature Communications*, 11, 576.
- 778 Venables WN, Ripley BD. (2002). Modern Applied Statistics with S. Fourth Edition. *Springer*,
779 New York. ISBN 0-387-95457-0.
- 780 Watson G, Ware G. (1994). What botanical gardens can contribute to urban forestry research.
781 *Public Garden*, 9:24-28.
- 782 Wenzel A, Grass I, Belavadi VV, Tschardt T. (2020). How urbanization is driving pollinator
783 diversity and pollination – A systematic review. *Biological Conservation*, 241, 108321.
- 784 White TCR. (1969). An index to measure weather-Induced stress of trees associated with outbreaks
785 of psyllids in Australia. *Ecology*, 50:905–909.
- 786 White TCR. (1984). The abundance of invertebrate herbivores in relation to the availability of
787 nitrogen in stressed food plants. *Oecologia*, 63:90–105.
- 788 Wu J. (2014). Urban ecology and sustainability: The state-of-the-science and future directions.
789 *Landscape and Urban Planning*, 125:209–221.

Supporting information



Fig S1. Florivory and florivores in some of the studied plant species. A- flowers and florivores of *Solanum lycocarpum*. B- Florivory in *Bauhinia variegata*. C- *Diabrotica speciosa* (Chrysomelidae), a florivore of *Turnera subulata*. D- Ant damaging *Perianthra mediterranea* flowers. E- Chrysomelidae beetle feeding on the petals of *Chamaecrista* sp. F- Florivory in *Ceiba speciosa*. G- *Conognatha compta* beetle feeding on the petals of *Microliceae congestiflora*. H- Florivory in *Trembleya laniflora*. I- Florivory in *Aspilia* sp.

Table S1. Summary of phylogenetic generalized least squares (PGLS) models used to assess the weight of the phylogeny of 61 plant species on florivory incidence. The existence of a significant relation between the variables is indicated when the GLS does not show a relevant relationship, but the PGLS analysis does, which suggests that these relationships are relevant only when considering the impact of phylogeny. The (*) indicate fixed factors that may potentially have an influence the incidence of florivory independently of the influence of phylogeny.

Response variable	Fixed factors (number of levels)	GLS		PGLS	
		F-value	p-Value	F-value	p-Value
Florivory incidence (Frequency of flower attack)	Collected* Sitio (3)	F _{2,57} = 4.96	0.01	F _{2,57} = 4.73	0.01
	Flower* size (61)	F _{1,58} =4.12	0.04	F _{1,58} =9.29	0.003
	Flower Color (6)	F _{5,54} = 1.27	0.28	F_{5,54}= 3.30	0.01
	Flower arrangement (2)	F _{1,58} =1.51	0.22	F_{1,58}= 6.1	0.01
	Flower Symmetry (2)	F _{1,58} = 0.014	0.90	F _{1,58} = 0.005	0.94
	Plant Growthform (4)	F _{3,56} = 1.34	0.26	F _{3,56} = 0.46	0.70
	Plant Origin (2)	F _{1,58} = 10.53	0.001	F _{1,58} = 0.58	0.44
	Plant function (2)	F _{1,58} =4.002	0.05	F _{1,58} = 0.23	0.62
	Florivory Intensity (61)	F _{1,58} =16.99	0.0001	F _{1,58} =3.63	0.061

Table S3. Summary of phylogenetic generalized least squares (PGLS) models used to assess the weight of the phylogeny of 61 plant species on florivory intensity. The existence of a significant relation between the variables is indicated when the GLS does not show a relevant relationship, but the PGLS analysis does, which suggests that these relationships are relevant only when considering the impact of phylogeny. The (*) indicate fixed factors that may potentially have an influence the incidence of florivory independently of the influence of phylogeny.

Response variable	Fixed factors (number of levels)	GLS		PGLS	
		F-value	p-Value	F-value	p-Value
Florivory intensity	Collected* Sitio (3)	F _{2,57} = 8.49	0.0006	F _{2,57} = 11.43	0.001
	Flower size (61)	F _{1,58} =0.04	0.84	F _{1,58} =0.69	0.40
(Floral area removed)	Flower Color (6)	F _{5,54} = 1.41	0.23	F _{5,54} = 2.01	0.09
	Flower arrangement (2)	F _{1,58} =0.79	0.37	F _{1,58} = 1.95	0.16
	Flower Symmetry (2)	F _{1,58} = 0.22	0.63	F _{1,58} = 0.05	0.80
	Plant Growthform (4)	F _{3,56} = 1.44	0.24	F_{3,56}= 3.68	0.01
	Plant* Origin (2)	F _{1,58} = 10.33	0.002	F _{1,58} = 5.7	0.02
	Plant * function (2)	F _{1,58} =10.49	0.002	F _{1,58} = 4.52	0.03

Considerações Finais

Apesar do potencial impacto negativo da florivoria para as plantas, estudos sobre interações flor-florívoro permanecem amplamente negligenciados na literatura ecológica já que a maioria dos estudos com interações inseto-plantas tende a se concentrar na herbivoria como relação antagonística e na polinização como relação mutualística. Aqui, nós fornecemos as primeiras informações sobre os padrões de florivoria a nível global, abordamos suas fontes de variação e identificamos os principais caminhos de pesquisa que ajudarão a preencher lacunas ainda persistentes no conhecimento acerca dessa interação. Ao dar os primeiros passos para avaliar a florivoria em ambientes naturais e modificados na região tropical, nós descobrimos que a ocorrência de dano floral é comum nas angiospermas, mas é amplamente variável entre as linhagens de plantas, apesar dos níveis de florivoria poderem ser até duas vezes maiores em flores de espécies tropicais comparado a espécies temperadas, sugerindo impactos potencialmente negativos ainda mais fortes na aptidão das plantas nos trópicos. Também descobrimos que tanto a identidade dos florívoros (classe e guilda alimentar), quanto o tipo de arranjo floral influenciam a frequência com que as flores são atacadas por inimigos florais. Descobrimos que os níveis de herbivoria e florivoria estão correlacionados dentro das mesmas espécies de plantas, embora as comunidades de florívoros e herbívoros sejam pouco similares, mas que as espécies plantas com períodos de floração mais longos formaram comunidades de florívoros mais similares, sugerindo que a fenologia das plantas, a aparição floral e a duração fenofásica também podem modular os níveis de dano floral em diversas espécies de plantas.

Ao avaliar a florivoria em ambientes modificados, como aqueles gerados pela urbanização, observamos que há um efeito das mudanças no uso da terra causadas pelo crescimento das cidades (e.g., aumento da cobertura de asfalto, diminuição das áreas verdes, avanço da ocorrência de espécies exóticas) no nível de florivoria, mas não na sua frequência de ocorrência, o que pode indicar que os insetos florívoros, ao encontrarem as plantas hospedeiras no ambiente urbano são capazes de infligir maior dano. Tal resultado tem aplicações para o plantio de espécies no ambiente urbano e no planejamento das cidades, uma vez que sabe-se que além dos efeitos negativos diretos da florivoria nas plantas, há também efeitos indiretos causados pela redução na atratividade das flores danificadas a insetos mutualistas, como polinizadores e dispersores. Assim, nossos resultados indicam que a florivoria pode impactar a aptidão de plantas urbanas - tanto nativas

quanto exóticas - e maior número de estudos são necessários para testar a generalidade desse padrão em cidades que variam em termos de grau de urbanização e impactos de uso do solo na ocorrência de insetos florívoros.

Em conjunto, nossos resultados dos dois capítulos começaram a dar os primeiros passos para entender a importância da florivoria nas plantas, seu padrão de ocorrência e distribuição e guiam pesquisas futuras para avaliar como florívoros, herbívoros e mutualistas interagem com os recursos e como impactam a aptidão de plantas em ambientes naturais e modificados.

REFERÊNCIAS

- Agrawal AA, Maron JL. (2022). Long-term impacts of insect herbivores on plant populations and communities. *Journal of Ecology*, 110, 2800–2811.
- Bronstein JL, Huxman, TE, Davidowitz G. (2007). Plant-mediated effects linking herbivory and pollination. In: Ohgushi T, Craig TG, Price PW eds. *Ecological Communities: Plant Mediation in Indirect Interaction Webs*. Cambridge, UK: Cambridge University Press, 79–103.
- Caruso CM, Eisen KE, Martin RA, Sletvold N. (2019). A meta-analysis of the agents of selection on floral traits. *Evolution*, 73: 4–14.
- Coley PD, Barone JA. (1996). Herbivory and plant defenses in tropical forests. *Annual Review of Ecology and Systematics*, 27, 305–335.
- Gorden NLS, Adler LS. (2016). Florivory shapes both leaf and floral interactions. *Ecosphere*, 7:e01326.
- Kozlov MV, Cornelissen T, Gavrikov DE, Kunavin MA, Lama AD, Milligan JR, et al. (2017). Reproducibility of fluctuating asymmetry measurements in plants: Sources of variation and implications for study design. *Ecological Indicators*, 73:733–740.
- Kozlov M, Lanta V, Zverev V, Zvereva E. (2015). Background losses of woody plant foliage to insects show variable relationships with plant functional traits across the globe. *Journal of Ecology*, 103: 1519–1528.
- McCall A. (2008). Florivory affects pollinator visitation and female fitness in *Nemophila menziesii*. *Oecologia*, 155: 729–737.
- McCall AC, Irwin RE. (2006). Florivory: the intersection of pollination and herbivory. *Ecology Letters*, 9:1351–1365.
- Marquis,RJ. (1984). Leaf herbivores decrease fitness of a tropical plant. *Science*, 226:537–539.
- Mendes GM, Silveira FAO, Oliveira C, Dattilo W, Guevara R, Ruiz-Guerra B, Boaventura MG, et al. (2021). How much leaf area do insects eat? A data set of insect herbivory sampled globally with a standardized protocol. *Ecology*, 102: e03301.

Metcalf DB, Asner GP, Martin RE, Silva Espejo JE, Huasco WH, et al. (2014). Herbivory makes major contributions to ecosystem carbon and nutrient cycling in tropical forests. *Ecology Letters*, 17, 324–332.

Rusman Q, Lucas-Barbosa D, Poelman EH, Dicke M. (2019). Ecology of plastic flowers. *Trends in Plant Science*, 24: 725–740.

Strauss SY, Conner JK, Rush SL. (1996). Foliar herbivory affects floral characters and plant attractiveness to pollinators: implications for male and female plant fitness. *The American Naturalist*, 147:1098-1107.

Turcotte MM, Davies TJ, Thomsen CJM, Johnson, MTJ. (2014). Macroecological and macroevolutionary patterns of leaf herbivory across vascular plants. *Proceedings of the Royal Society B: Biological Sciences*, 281:20140555.

Turcotte MM, Thomsen CJM, Broadhead GT, Fine PVA, Godfrey RM, Lamarre GPA, et al. (2014b), Percentage leaf herbivory across vascular plant species. *Ecology*, 95, 788–788.