



OPEN

Diversity, distribution and ecology of fungal communities present in Antarctic lake sediments uncovered by DNA metabarcoding

Láuren Machado Drumond de Souza¹, Juan Manuel Lirio², Sílvia Herminda Coria², Fabyano Alvares Cardoso Lopes³, Peter Convey^{4,5}, Micheline Carvalho-Silva⁶, Fábio Soares de Oliveira⁷, Carlos Augusto Rosa¹, Paulo Eduardo Aguiar Saraiva Câmara⁶ & Luiz Henrique Rosa¹✉

We assessed fungal diversity in sediments obtained from four lakes in the South Shetland Islands and James Ross Island, Antarctica, using DNA metabarcoding. We detected 218 amplicon sequence variants (ASVs) dominated by the phyla *Ascomycota*, *Basidiomycota*, *Mortierellomycota*, *Mucoromycota* and *Chytridiomycota*. In addition, the rare phyla *Aphelidiomycota*, *Basidiobolomycota*, *Blastocladiomycota*, *Monoblepharomycota*, *Rozellomycota* and *Zoopagomycota* as well as fungal-like Straminopila belonging to the phyla *Bacillariophyta* and *Oomycota* were detected. The fungal assemblages were dominated by unknown fungal taxa (Fungal sp. 1 and Fungal sp. 2), followed by *Talaromyces rubicundus* and *Dactylonectria anthuriicola*. In general, they displayed high diversity, richness and moderate dominance. Sequences representing saprophytic, pathogenic and symbiotic fungi were detected, including the phytopathogenic fungus *D. anthuriicola* that was abundant, in the relatively young Soto Lake on Deception Island. The lake sediments studied contained the DNA of rich, diverse and complex fungal communities, including both fungi commonly reported in Antarctica and other taxa considered to be rare. However, as the study was based on the use of environmental DNA, which does not unequivocally confirm the presence of active or viable organisms, further studies using other approaches such as shotgun sequencing are required to elucidate the ecology of fungi in these Antarctic lake sediments.

Antarctica represents one of the most extreme regions of the planet, including polyextremophilic environments and habitats that combine cold, dry and ultra-oligotrophic conditions, and offers unique opportunities to discover and study extremophilic organisms¹. The large majority of Antarctica (>99.6% of its area) is permanently covered by ice and snow. Nevertheless, terrestrial and freshwater ecosystems are present that provide habitats occupied by organisms living in some of the most extreme conditions on the planet². Antarctic lakes offer multiple extreme environmental challenges, including low temperatures, high salinity, pH variation, seasonally high UV radiation and low nutrient availability¹. Various types of lakes are present, of different size, depth, trophic status and age, whose ecosystems are dominated by microorganisms adapted to the extreme conditions³.

Lakes present in the maritime Antarctic are typically shallow systems (<10 m deep), transparent allowing penetration of high levels of light and UV radiation, with low temperatures and low available nutrients⁴. However, the complex geology of the different islands in maritime Antarctica results in variable lake chemistry, which may favor the dominance of different microorganisms able to survive in these ecosystems and play a major role in the transfer of inorganic and organic material and energy⁵. The freshwater aquatic zones of Antarctica, including lakes, are in direct or indirect contact with air, rocks, soil and animals, and are bordered by vegetation;

¹Laboratório de Microbiologia Polar e Conexões Tropicais, Departamento de Microbiologia, Instituto de Ciências Biológicas, Universidade Federal de Minas Gerais, P. O. Box 486, Belo Horizonte, MG 31270-901, Brazil. ²Instituto Antártico Argentino, Buenos Aires, Argentina. ³Laboratório de Microbiologia, Universidade Federal do Tocantins, Porto Nacional, Brasil. ⁴British Antarctic Survey, NERC, High Cross, Madingley Road, Cambridge CB3 0ET, UK. ⁵Department of Zoology, University of Johannesburg, PO Box 524, Auckland Park 2006, South Africa. ⁶Departamento de Botânica, Universidade de Brasília, Brasília, Brasil. ⁷Departamento de Geografia, Universidade Federal de Minas, Gerais, Belo Horizonte, MG, Brasil. ✉email: lhrosa@icb.ufmg.br

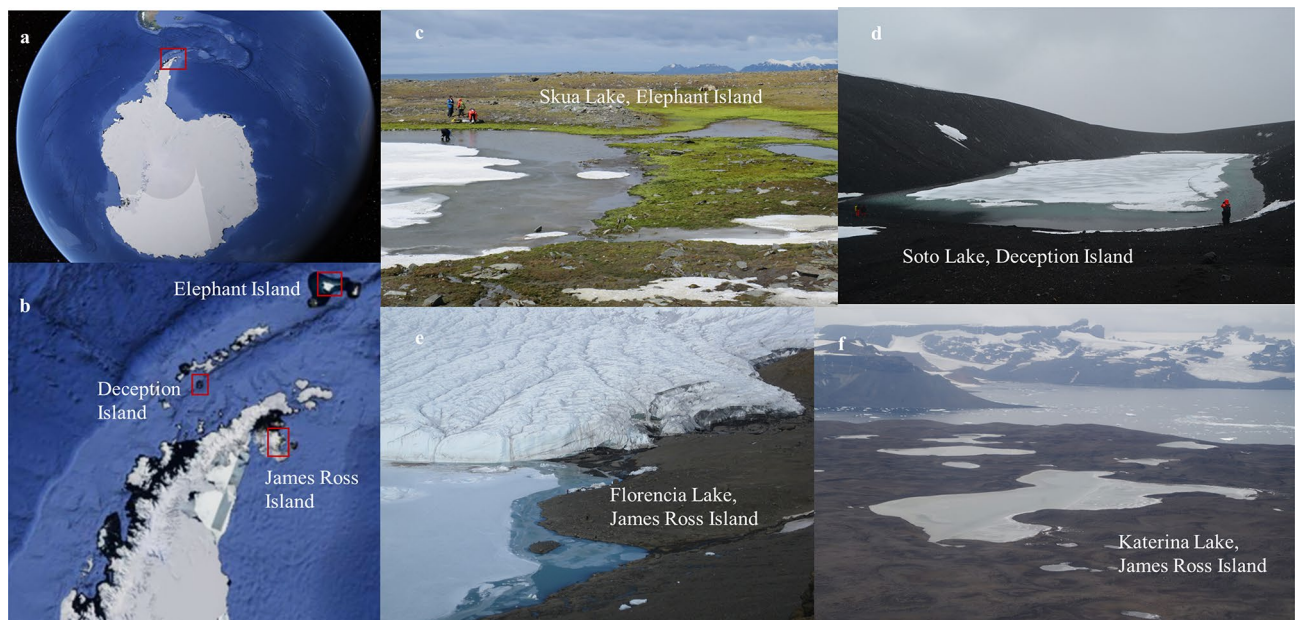


Figure 1. Satellite images (a) and (b) (obtained in Google Earth Pro, 2019). (a) Antarctica with the Antarctic Peninsula inside the red rectangle, (b) Elephant, Deception and James Ross Islands inside the red rectangle, (c) Skua Lake, Elephant Island ($61^{\circ}13'18.2''\text{S}$; $55^{\circ}21'54.3''\text{W}$), (d) Soto Lake, Deception Island ($62^{\circ}58'44.6''\text{S}$; $60^{\circ}33'20.2''\text{W}$), (e) Florencia ($64^{\circ}01'27.8''\text{S}$; $57^{\circ}40'00.6''\text{W}$) and (f) Katerina Lakes ($64^{\circ}01'18.6''\text{S}$; $57^{\circ}43'26.2''\text{W}$), James Ross Island. Photos (c) and (e) taken by Juan M. Lirio, photo (f) taken by Matej Roman and photo (d) taken by Luiz H. Rosa.

consequently, Antarctic lakes represent important sites for the study of microbial diversity and ecology⁶. Antarctic lake sediments can provide regional climatic archives and also shelter unique microbial communities, including bacteria, cyanobacteria, viruses, protists and fungi^{7,8}.

Various Antarctic fungal communities have been characterized, including those in Antarctic lakes, which face various extreme conditions including low temperatures, high salinity, pH variation, seasonally high UV radiation and low nutrient availability¹. To date, most fungal diversity studies in Antarctic lakes have applied traditional culturing methods^{1,3,8–11}, which do not reveal the full diversity of the resident mycobiota. The recent application of approaches using metabarcoding have highlighted that fungal diversity in Antarctic lakes can potentially be considerably greater than previously appreciated^{13–15}.

With this background, we hypothesized that metabarcoding approaches using high throughput sequencing (HTS) may detect cryptic fungal assemblages present in sediments of Antarctic lakes and contribute to better understanding of the complex resident fungal ecological networks involving saprophytic, mutualistic and parasitic taxa living under different extreme environmental conditions. We studied fungal richness, diversity and ecology in sediments sampled from four Antarctic lakes in the South Shetland Islands, north-west of the Antarctic Peninsula and James Ross Island to the east of the tip of the Antarctic Peninsula. We used a metabarcoding approach to assess and characterize fungal DNA sequence diversity present in sediment samples obtained from these four lakes.

Methods

Study sites and sediment sampling. The four Antarctic lakes sampled were Skua Lake (Elephant Island), Soto Lake (Deception Island), both in the South Shetland Islands, and Florencia Lake and Katerina Lake (James Ross Island) (Fig. 1). Skua Lake is the northern-most of these lakes. It is an open basin, up to 1 m deep, fed by glacial meltwater from the Elephant Island Ice Cap. The lake occupies a depression on metamorphic rocks formed by glacial erosion. Soto Lake is a shallow crater depression formed by a phreatomagmatic eruption (maar). Its precise depth has not been measured, but is likely to be several meters. The lake is in a closed basin, fed by precipitation and its basement is formed by volcanic rock of basaltic composition. Florencia Lake is located on a sector of Clearwater Mesa known as Foreland on James Ross Island. It is an open basin, more than 10 m deep, directly abutting Blancmange Glacier. Its basement is formed by basal till deposits composed mainly of volcanic hyaloclastite breccias. Katerina Lake is located on top of Clearwater Mesa, lying on a relatively flat surface formed by a basaltic lava flow. It is an open basin, up to 1 m depth, and is fed by snowmelt. Each of these lakes displays different geological characteristics that may influence the resident mycobiota.

Sediments were collected from Skua Lake and Soto Lake during the austral summer of 2016/17 and from Katerina Lake and Florencia Lake in the austral summer of 2020/21 (Fig. 1; Table 1). Samples from all lakes were collected in triplicate using PVC pipes (60 mm diameter \times 50 mm length), disinfected to avoid contamination as described by Ogaki et al.⁸. Sediments were sampled from three different points (each separated by approximately 50 m) in the littoral region of each lake at a depth of 20–50 cm. Approximately 500 g of each sediment sample

Lake geological characteristics	Lake			
	Skua	Soto	Katerina	Florencia
Location	61° 13' 18.2" S 55° 21' 54.3" W	62° 59' 05.4" S 60° 39' 18.0" W	64° 01' 25.5" S 57° 43' 03.6" W	64° 01' 24.0" S 57° 40' 03.1" W
Altitude (meters above sea level)	59	13	250	25
Total area (m ²)	6009	18,190	126,951	132,000
Perimeter (m)	400	630	2085	1854
Distance to coastline (m)	500	220	1,337	1,429
Depth (m)	0.4	> 3	1	> 10
Lake shape	Elongate (E–W)	Rectangular	Irregular	Circular
Sediment chemical parameters				
pH in H ₂ O	6.1	8.8	8.4	8.5
Exchangeable P—mg dm ⁻³	218.4	19.5	120.5	9.3
Sum of exchangeable bases Ca + K + Mg (SB)—cmol _c /dm ³	0.73	3.51	5.53	7.59
Percentage of base saturation (PBS)—%	35.6	78.0	94.4	79.3
H + Al—potential acidity—cmol _c dm ⁻³	1.32	0.99	0.33	1.98
Cation exchange capacity at pH 7 (CEC)—cmol _c dm ⁻³	2.05	4.50	5.86	9.57
Total organic carbon (TOC)—dag/kg	0.07	0.08	0.54	0.38
Micronutrient Fe—mg dm ⁻³	84.5	231.2	252.2	44.0
Micronutrient Mn—mg dm ⁻³	2.3	11.1	54.2	63.2
Resident fungal diversity indices				
Number of DNA reads	33,429	71,983	54,939	46,813
Number of taxa	50	31	51	171
Fisher's- α (diversity)	21.40	10.27	22.1	559.7
Margalef (richness)	9.25	5.66	9.44	32.09
Simpson's (dominance)	0.76	0.78	0.79	0.92

Table 1. Lake locations, characteristics, sediment physicochemical data and diversity indices of fungal assemblages obtained from Skua Lake (Elephant Island), Soto Lake (Deception Island), Katerina Lake and Florencia Lake (James Ross Island).

was immediately sub-sampled, sealed, placed in sterile Whirl-pack bags (Nasco, Ft. Atkinson, WI) and frozen at -20°C until processing in the laboratory at the Federal University of Minas Gerais, Brazil. There, the sampled was gradually thawed at 4°C for 24 h before carrying out DNA extraction. In all DNA extraction steps, we proceeded under strict control conditions within a laminar flow hood to recover the fungal DNA and avoid contaminations.

Sediment chemical analysis. Sediment chemical analyses were performed following Embrapa¹⁶. pH was determined using a 1:2.5 sediment:deionized water ratio. Potential acidity (H + Al) was extracted with $0.5\text{ mol L}^{-1}\text{ Ca}(\text{OAc})_2$ buffered to pH 7.0 and quantified by titration with $0.0606\text{ mol L}^{-1}\text{ NaOH}$. Exchangeable Ca^{2+} , Mg^{2+} and Al^{3+} were extracted with $1\text{ mol L}^{-1}\text{ KCl}$, and K^{+} and P^{+} were extracted with Melich¹⁶. The element levels in the extracts were determined by ICP (Al^{3+}), flame emission (Na^{+} , K^{+}) and photolorimetry (P) by the ascorbic acid method. Total organic carbon (TOC) was quantified by wet oxidation using the Walkley–Black method. All analyses were performed in triplicate. Total cation exchange capacity (CEC) was calculated as the sum of the bases (Ca^{2+} , Mg^{2+} , K^{+} , Al^{3+}) and potential acidity (H^{+} + Al^{3+}). Chemical analyses of all sediment samples were performed in triplicate.

DNA extraction, Illumina library construction and sequencing. Three replicate sub-samples were taken from the center of each core section under strict contamination control conditions. Total DNA was extracted from these using the FastDNA Spin Kit for Soil (MPBIO, Ohio, USA), following the manufacturer's instructions. DNA quality was analyzed by agarose gel electrophoresis (1% agarose in $1\times\text{Trisborate-EDTA}$) and then quantified using the Quanti-iT[™] Pico Green dsDNA Assay (Invitrogen). The internal transcribed spacer 2 (ITS2) of the nuclear ribosomal DNA was used as a DNA barcode for molecular species identification of Fungi^{17,18} using the universal primers ITS3 and ITS4¹⁹. Library construction and DNA amplification were performed using the Herculanase II Fusion DNA Polymerase Nextera XT Index Kit V2, following Illumina 16S Metagenomic Sequencing Library Preparation protocol (Part #15044223, Rev. B). Paired-end sequencing ($2\times 300\text{ bp}$) was performed on a MiSeq platform (Illumina) by Macrogen Inc. (South Korea). All quality control to avoid contamination of DNA extraction, PCR, sequencing were carried out and analyzed by Macrogen Inc.

Data analysis and fungal identification. Quality analysis was carried out using BBDuk v. 38.87 in BBmap software²⁰ with the following parameters: Illumina adapters removing (Illumina artefacts and the PhiX Control v3 Library); ktrim = 1; k = 23; mink = 11; hdist = 1; minlen = 50; tpe; tbo; qtrim = rl; trimq = 20; ftm = 5;

maq = 20. The remaining sequences were imported to QIIME2 version 2021.4 (<https://qiime2.org/>) for bioinformatics analyses²¹. The qiime2-dada2 plugin was used for filtering, dereplication, turn paired-end fastq files into merged, and remove chimeras, using default parameters²². Taxonomic assignments were determined for amplicon sequence variants (ASVs) in three steps. First, ASVs were classified using the qiime2-feature-classifier²³ classify-sklearn against the UNITE Eukaryotes ITS database version 8.3²⁴. Second, remaining unclassified ASVs were filtered and aligned against the filtered NCBI non-redundant nucleotide sequences (nt) database (October 2021) using BLASTn²⁵ with default parameters; the nt database was filtered using the following keywords: “ITS1”, “ITS2”, “Internal transcribed spacer”, and “internal transcribed spacer”. Third, output files from BLASTn²⁵ were imported to MEGAN6²⁶ and taxonomic assignments were performed using the “megan-nucl-Jan2021.db” mapping file with default parameters and trained with Naive Bayes classifier and a confidence threshold of 98.5%. Taxonomic profiles were plotted using the Krona²⁷. The heatmaps of ASV abundance and clustering analysis were created using Heatmapper²⁸; clustering analysis was performed using the following parameters: Average Linkage, Spearman Rank Correlation and Z-score among samples for each ASV.

Many factors, including extraction, PCR and primer bias, can affect the number of reads obtained²⁹, and thus lead to misinterpretation of absolute abundance³⁰. However, Giner et al.³¹ concluded that such biases did not affect the proportionality between reads and cell abundance, implying that more reads are linked with higher abundance^{32,33}. Therefore, for comparative purposes, we used the number of reads as a proxy for relative abundance. Fungal classification followed Kirk et al.³⁴, Tedersoo et al.³⁵, MycoBank (<http://www.mycobank.org>) and the Index Fungorum (<http://www.indexfungorum.org>).

Fungal diversity and ecology. The relative abundances of the ASVs were used to quantify the fungal taxa present in the total sediments sampled, where fungal ASVs with relative abundance > 10% were considered dominant, those between 1 and 10% as intermediate and those with < 1% as minor (rare) components of the fungal community³⁶. The numbers of reads were used to quantify taxon diversity, richness and dominance, using the following indices: (i) Fisher’s α , (ii) Margalef’s and (iii) Simpson’s, respectively. In addition, species accumulation curves were obtained using the Mao Tao index. All results were obtained with 95% confidence, and bootstrap values were calculated from 1000 replicates using the PAST computer program 1.90³⁷. Functional assignments of fungal ASVs at species and generic levels were prepared using FunGuild³⁸, which can be accessed at <http://www.funguild.org/>.

Results

Fungal taxonomy. A total of 207,164 DNA reads were detected in the sediments from the four lakes, representing 218 ASVs (see Supplementary Table S1 online). The Mao Tao rarefaction curves of the fungal assemblages detected in all four lake sediments are shown in Supplementary Fig. S1 online. The curves did not reach asymptote, indicating that further fungal sequence diversity is likely to be present in these sediments. Supplementary Figs. S2 and S3 show the abundances of fungal and allied organism ASVs at different hierarchical levels, which varied across the four lakes. Unknown fungi dominated the sequence assemblages, followed by the phyla *Ascomycota*, *Basidiomycota*, *Mortierellomycota*, *Mucoromycota* and *Chytridiomycota*. The lake sediment assemblages were dominated by Fungal sp. 1 (Skua, Katerina and Florencia Lakes), *Talaromyces rubicundus* (Soto and Katerina Lakes) and *Dactyloectria anthuriicola* (Soto Lake), in rank order, which displayed the highest relative abundances. However, richness (215 ASVs) of the fungal assemblages detected primarily comprised taxa of intermediate and rare relative abundance, including those of fungal phyla rarely (*Rozellomycota*, *Zoopagomycota*) or unreported (*Aphelidiomycota*, *Basidiobolomycota*, *Blastocladiomycota*, *Monoblepharomycota*) from Antarctica as well as members of the allied kingdom Straminopila (phylum *Oomycota*), which have also been infrequently detected in Antarctica in studies using traditional isolation methods. Ninety-five ASVs (47.57% of the total detected) could only be assigned to higher taxonomic levels (phylum, class, order or family) and may represent taxa not currently included in the available sequence databases or be new Antarctic records and/or previously undescribed taxa.

Influence of physical and chemical properties of lacustrine sediments on the contained fungal communities. The fungal assemblages present in the lake sediments showed variation in their ecological diversity indices but, in general, despite the extreme physicochemical characteristics, they displayed high diversity and richness indices and moderate dominance (Table 1). The sediment fungal assemblage of Florencia Lake displayed the highest number of taxa and diversity indices, followed by those from Katerina, Skua and Soto Lakes.

The four lakes had sediments with similar chemical attributes, with low values for most of the elements. The sediment of Skua Lake differed from the other lakes in having slightly higher acidity (lower pH value and higher H + Al), higher exchangeable P content, and very low values for sum of bases and CEC. It was the only lake with dystrophic sediments (PBS < 50%). In contrast, sediments from Katerina Lake and Florencia Lake had the largest nutrient reserve in the form of exchangeable elements, mainly Ca, K and Mg, with the highest values of base saturation and CEC observed. The total organic carbon content was very low in sediments from all four lakes (< 1 dag kg⁻¹). Of the total fungal ASVs detected, eight were present in all four lake sediments (Fig. 2; Supplementary Table S2). Taxa distribution varied across the four lakes, with sediments in each lake hosting some specific fungal taxa, in particular Florencia Lake with 116 fungal ASVs. The most abundant fungal ASV overall (Fungal sp. 1) was present in all four lake sediments. Beyond the eight cosmopolitan fungal ASVs that were shared amongst all lakes, no other ASVs were shared among Skua Lake (Elephant Island) and Soto Lake (Deception Island). Although both Katerina Lake and Florencia Lake are in relatively close proximity on Clearwater Mesa, James Ross Island (see Supplementary Fig. S4 online), the fungal assemblages detected in each (51 and 171 ASVs,

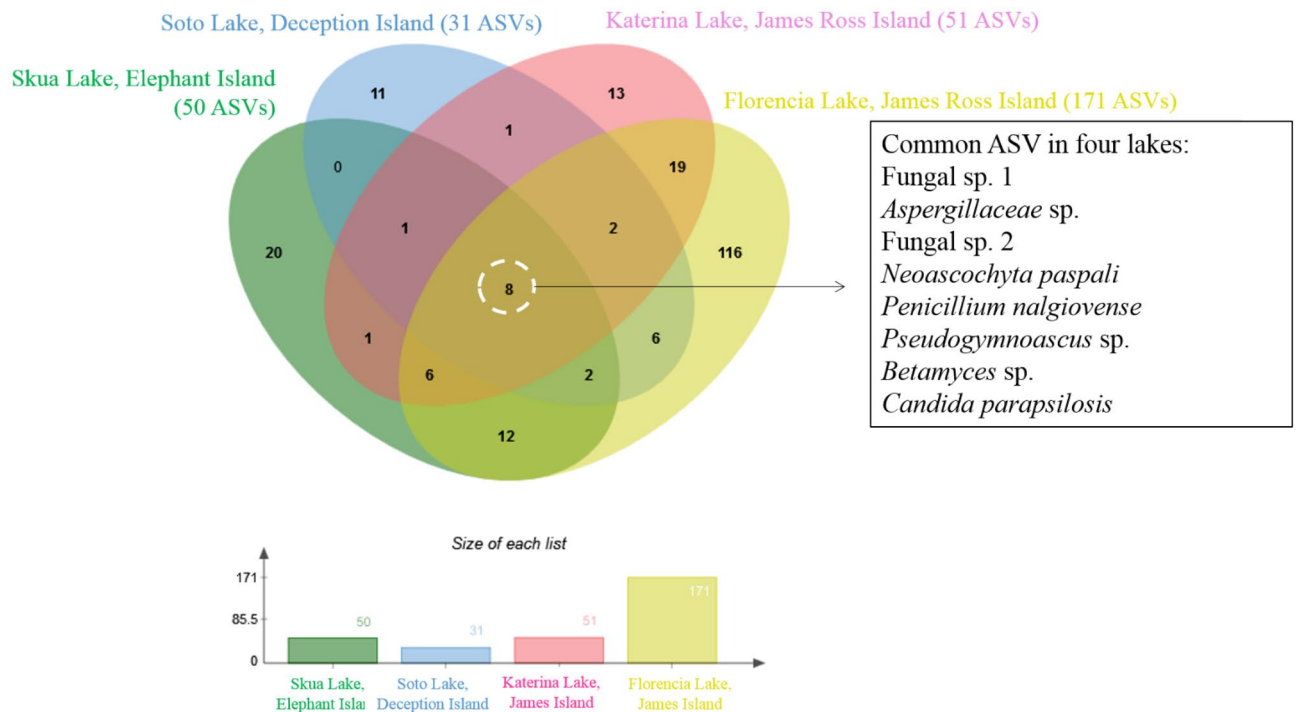


Figure 2. Venn diagram showing the distribution of fungal amplicon sequence variants (ASVs) among the sediment samples obtained from Skua Lake (Elephant Island), Soto Lake (Deception Island), Katerina Lake and Florencia Lake (James Ross Island).

respectively) shared only 19 fungal taxa and displayed distinct ecological diversity indices. Functional ecology assignments of the ASVs detected at generic level are shown in Supplementary Table S3 online, and indicated that the fungal communities present in the sediments of all four lakes were dominated by saprotrophic, plant and animal pathogenic and symbiotic taxa.

Discussion

Taxonomy and occurrence. The availability of studies of fungal diversity in Antarctic lake sediments has increased in recent years, but still remains limited^{8,12}. Most previous studies of Antarctic lake fungal communities have used traditional culture-dependent methods^{1,3,5,8,39}. However, recent applications of metabarcoding approaches have detected the presence of DNA sequences of different fungal taxa in Antarctic lake sediments^{14,15}. The current study similarly confirmed the presence of high fungal sequence richness and diversity in sediments from four lakes sampled in the South Shetland Islands and James Ross Island. The study detected both the commonly encountered phyla *Ascomycota*, *Basidiomycota*, *Mortierellomycota*, *Mucoromycota*, *Chytridiomycota*, *Rozellomycota* and *Zoopagomycota* (cf. reports of da Silva et al.⁴⁰, Ogaki et al.¹⁴ and Rosa et al.¹⁵), and the more rarely detected or unreported *Aphelidiomycota*, *Basidiobolomycota*, *Blastocladiomycota* and *Monoblepharomycota*, as well as members of the fungal-like kingdom Straminopila in the phylum *Oomycota*.

ASVs assigned to Fungal sp. 1, *T. rubicundus* and *D. anthuriicola* were the most abundant fungal taxa detected in the lake sediments. The most abundant ASV was Fungal sp. 1 (24.757%), suggesting a dominance of possibly undescribed fungi present in Antarctic lake sediment ecosystems. This unidentified dominant ASV provides an example of a sequence not currently included in the available sequence databases used in our study. *Talaromyces*, the teleomorphic form of *Penicillium*⁴¹, includes species pathogenic to humans, but also producers of antibiotics and enzymes potentially useful in biotechnological processes⁴². *Talaromyces rubicundus* has been reported to produce various extrolites such as islandicin chromophore, mitorubrin, mitorubrinic acid and rubratoxins⁴². In Antarctica, representatives of *Talaromyces* and *Penicillium* have been detected in different habitats and substrates including soil, marine and lake sediments, freshwater and seawater, plants, macroalgae, rocks, snow, ice⁴³ and air³⁶. *Talaromyces rubicundus* (anamorphic form of *Penicillium rubicundum*) was originally isolated from cultivated soil in Georgia, USA, and displays mesophilic characteristics and an optimal growth temperature of 25°C⁴⁴. The genus *Dactylonectria* includes 14 species, some of which are recognized as soil-borne plant pathogens⁴⁵. *Dactylonectria anthuriicola* (synonym of *Ilyonectria anthuriicola*) was isolated from the roots of *Anthurium* sp. in the Netherlands⁴⁶. Recently, Parkinson et al.⁴⁷ reported that *D. anthuriicola* caused significant root rot on avocado trees in Australia. Our study represents the first records of both *T. rubicundus* and *D. anthuriicola* in Antarctica.

Ninety-five ASVs, representing almost the half of the total community, were identified at phylum, class, order or family levels. These fungi may represent taxa not currently included in the available sequence databases or be undescribed taxa. However, as our study focused on fungal diversity assigned from environmental DNA samples, further specific taxonomic studies are required to elucidate if the apparently high richness of undescribed taxa does indeed represent previously unrecognized fungal species present in Antarctica.

Diversity and ecology. Our fungal sequence diversity data can be compared with studies using both traditional culturing methods and the increasing number of recent metabarcoding studies of Antarctic lakes (see Supplementary Table S5 online). The diversity indices of fungal sequence assemblages detected in the sediments of the four lakes studied here were greater than those reported in culture-based studies but comparable with those using metabarcoding. Ogaki et al.¹² used culture-dependent methods to characterize the same sediments from Skua and Soto Lakes, identifying only 13 and 6 fungal taxa, respectively, mostly at genus and species levels, contrasting with our data obtained using metabarcoding.

The four lakes also differed in their diversity indices, with Soto Lake and Florencia Lake displaying the lowest and highest diversity indices, respectively. No significant correlation was observed between the diversity indices and the sediment physicochemical parameters of the lakes. Considerable differences in the perimeters and depths of these two lakes may be correlated with their sediment accumulation and, hence, DNA deposition into the sediments, underlying the different levels of diversity detected.

Soto Lake is not surrounded by a well-developed community of mosses or lichens and is characterized by low nutrient availability. It represents an ecosystem still in formation and displayed the lowest diversity indices. Skua Lake showed intermediate diversity indices, which may be due to the rich terrestrial vegetation and presence of nesting skuas and giant petrels in its surroundings. Florencia Lake showed the highest diversity and richness indices, which were strongly divergent with those in Katerina Lake, despite their close proximity (separated by approximately 2050 m) (see Supplementary Fig. S4 online). However, these two lakes have geological differences and are surrounded by extensive moss and lichen communities. Florencia Lake has the greatest water inflow of the four studied lakes, being fed mainly by meltwater that carries significant quantities of clastic sediment, as well as receiving water from other lacustrine basin lakes with which it is connected. These are fed by snowmelt and are surrounded by a rich flora of mosses and microbial mats that provide organic matter to the lake. The local gradients in the Florencia Basin also encourage fluvial erosion of the moss carpets and the transport of fragments towards Florencia Lake. In contrast, Katerina Lake is fed by snowmelt and linked with other small lakes also fed by snowmelt through small ephemeral streams with low transport capacity, although it is also surrounded by a relatively rich terrestrial flora of mosses and lichens. The range of potential sources of organic matter input to Florencia Lake perhaps underlies this lake having much higher fungal diversity indices than Katerina Lake.

The chemical attributes of lake sediments reflect their geological contexts and the biological communities in their catchments. These attributes and the low TOC values suggest a mainly mineral composition of the sediments in the studied lakes. The lakes associated with basic igneous rocks, of basaltic composition (Soto, Katerina and Florencia) had higher SB values and other indices (PBS, CEC). These rocks contain minerals rich in Ca, K and Mg, which form a major constituent of the sediments in these lakes. Geochemical studies (personal communication Dr. Siliva Coria—Instituto Antártico Argentino) of Florencia Lake sediments have identified minerals such as plagioclase, potassium feldspar, piroxenes, zeolites and forsterite, that are rich in Ca, K, Na, Fe and Mg.

In contrast, Skua Lake is located in an area of metamorphic rocks that have greater content of aluminosilicate minerals, such as mica, quartz and garnet, with lower base content. Metamorphic rocks are commonly associated with more acidic and less fertile soils than rocks of basic composition⁴⁸. Skua Lake is surrounded by rich vegetation and receives input of excrement from surrounding breeding birds, which leads to higher values of exchangeable P⁴⁹. The sediment samples examined in this study were collected in the littoral regions of each lake at 20–50 cm depth, meaning that there could be a mix of present-day and past fungal communities dating from years to centuries ago contained in the sediments.

The fungal genera detected in the sediments display different ecological roles including saprophytes, mutualists, symbionts and parasites, as more generally reported for Antarctic fungi⁴³. Among the genera with known ecological roles, saprophytes dominated the fungal communities present in all four lake sediments, followed by plant and animal pathogens and symbionts. Studies that have addressed the functional ecological roles of fungi in different Antarctic environments and habitats report saprophytes as the dominant functional group, followed by plant and animal pathogens and symbionts. The same functional ecological profile detected here has also been reported in studies sampling fungi in the air^{36,50}, soil⁵¹, fresh water¹³ and rocks⁵² in Antarctica. Similar functional profiles were reported in metabarcoding studies of lake sediments reported by Ogaki et al.¹⁴ from Vega Island (located close to James Ross Island) and Rosa et al.¹⁵ in Hope Bay at the north-eastern tip of the Antarctic Peninsula. However, in sediment samples obtained from Soto Lake (Deception Island), the recognized phytopathogenic species *D. anthuriicola* dominated the sequence assemblage, although the species has not previously been recorded in Antarctica. Fungi inhabiting polar environments commonly display the capability to degrade organic matter at low temperature and release compounds containing carbon, nitrogen and other elements to other organisms⁵³, suggesting that the saprophytic fungi detected in the four lake sediments studied here might host a complex fungal community that plays a vital role in the decomposition of organic matter under extreme conditions.

Conclusions

Application of a metabarcoding approach revealed that sediments from the four lakes studied contained DNA potentially representing rich, diverse and complex fungal communities. These included both known fungi commonly reported in Antarctica (members of the phyla *Ascomycota*, *Basidiomycota*, *Mortierellomycota* and *Mucoromycota*), as well as others considered rare or not previously reported in Antarctica (*Chytridiomycota*, *Rozellomycota*, *Zoopagomycota*, *Aphelidiomycota*, *Basidiobolomycota*, *Blastocladiomycota*, *Monoblepharomycota*), and the presence of the fungal-like Straminopila. Antarctic lake sediments accumulate over long periods of time, and our data indicate that they may be considered a hotspot of fungal diversity, potentially including new and/or previously unreported species. The dominance of sequences of a small number of saprophytic and phytopathogenic

taxa was notable. This metabarcoding study was based on the use of environmental DNA and further studies for instance using shotgun sequencing are now required to elucidate the ecology of Antarctic lake sediment fungi.

Data availability

All data generated or analysed during this study are included in this published article and its supplementary information files.

Received: 9 January 2022; Accepted: 29 April 2022

Published online: 19 May 2022

References

1. Gonçalves, V. N. *et al.* Diversity and distribution of fungal communities in lakes of Antarctica. *FEMS Microbiol. Ecol.* **82**, 459–471 (2012).
2. Convey, P. Antarctic ecosystems. In *Encyclopedia of Biodiversity* (ed. Levin, S. A.) 179–187 (Elsevier, Amsterdam, 2017).
3. Vincent, W. F. Evolutionary origins of Antarctic microbiota: invasion, selection and endemism. *Antarct. Sci.* **12**, 374–385 (2000).
4. Izaguirre, I., Mataloni, G., Vinocur, A. & Tell, G. Temporal and spatial variations of phytoplankton from Boeckella lake (Hope Bay, Antarctic Peninsula). *Antarct. Sci.* **5**, 137–141 (1993).
5. Ellis-Evans, J. C. Microbial diversity and function in Antarctic freshwater ecosystems. *Biodivers. Conserv.* **5**, 1395–1431 (1996).
6. Ogaki, M. B. *et al.* Diversity and ecology of fungal assemblages present in lakes of Antarctica. In *Fungi of Antarctica: Diversity, Ecology and Biotechnological Applications* (ed. Rosa, L. H.) 69–97 (Springer, Berlin, 2019).
7. Chaparro, M. *et al.* Sedimentary analysis and magnetic properties of Lake Anónima, Vega Island. *Antarct. Sci.* **29**, 429–444 (2017).
8. Ogaki, M. B. *et al.* Diversity, ecology, and bioprospecting of culturable fungi in lakes impacted by anthropogenic activities in Maritime Antarctica. *Extremophiles* **24**, 637–655 (2020).
9. Brunati, M. *et al.* Diversity and pharmaceutical screening of fungi from benthic mats of Antarctic lakes. *Mar. Gen.* **2**, 43–50 (2009).
10. Quesada, A., Camacho, A., Rochera, C. & Velázquez, D. Byers Peninsula: A reference site for coastal, terrestrial land limnetic ecosystem studies in maritime Antarctica. *Polar Sci.* **3**, 181–187 (2009).
11. Vaz, A. B. M. *et al.* The diversity, extracellular enzymatic activities and photoprotective compounds of yeasts isolated in Antarctica. *Braz. J. Microbiol.* **42**, 937–947 (2011).
12. Ogaki, M. B. *et al.* Diversity and bioprospecting of cultivable fungal assemblages in sediments of lakes in the Antarctic Peninsula. *Fungal Biol.* **124**, 601–611 (2020).
13. de Souza, L. D. *et al.* Assessment of fungal diversity present in lakes of Maritime Antarctica using DNA metabarcoding: a temporal microcosm experiment. *Extremophiles* **25**, 77–84 (2021).
14. Ogaki, M. B. *et al.* Diversity of fungal DNA in lake sediments on Vega Island, north-east Antarctic Peninsula assessed using DNA metabarcoding. *Extremophiles* **25**, 257–265 (2021).
15. Rosa, L. H. *et al.* Fungal diversity in a sediment core from climate change impacted Boeckella Lake, Hope Bay, north-eastern Antarctic Peninsula assessed using metabarcoding. *Extremophiles* **26**, 1–10 (2022).
16. Embrapa, S. *Manual de métodos de análise de solo* 3rd edn. (Centro Nacional de Pesquisa de Solos, 2017).
17. Chen, S. *et al.* Validation of the ITS2 region as a novel DNA barcode for identifying medicinal plant species. *PLoS ONE* **5**, e8613 (2010).
18. Richardson, R. T. *et al.* Application of ITS2 metabarcoding to determine the provenance of pollen collected by honey bees in an agroecosystem. *Appl. Plant Sci.* **3**, 1400066 (2015).
19. White, T. J. *et al.* Amplification and direct sequencing of fungal ribosomal RNA genes for phylogenetics. In *PCR Protocols: A Guide to Methods and Applications* (ed. Innis, M. A., Gelfand, D.H., Sninsky, J.J., White, T.J.) 315–322 (1990).
20. Bushnell, B. “BBMap: a fast, accurate, splice-aware aligner”. *Lawrence Berkeley National Lab.* (LBNL), Berkeley, CA (United States). <https://sourceforge.net/projects/bbmap> (2014).
21. Bolyen, E. *et al.* Reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. *Nat. Biotechnol.* **37**, 852–857 (2019).
22. Callahan, B. J. *et al.* DADA2: high-resolution sample inference from Illumina amplicon data. *Nat. Methods.* **13**, 581–583 (2016).
23. Bokulich, N. A. *et al.* Optimizing taxonomic classification of marker-gene amplicon sequences with QIIME 2’s q2-feature-classifier plugin. *Microbiome.* **6**, 90 (2018).
24. Abarenkov, K. *et al.* UNITE QIIME release for eukaryotes. Version 04.02.2020. UNITE Community. doi:<https://doi.org/10.1515/BIO/786386> (2020).
25. Camacho, C. *et al.* BLAST+: architecture and applications. *BMC Bioinform.* **10**, 1–9 (2009).
26. Huson, D. H. *et al.* MEGAN community edition-interactive exploration and analysis of large-scale microbiome sequencing data. *PLoS Comput. Biol.* **12**, e1004957 (2016).
27. Ondov, B. D., Bergman, N. H. & Phillippy, A. M. Interactive metagenomic visualization in a web browser. *BMC Bioinform.* **12**, 385 (2011).
28. Babicki, S. *et al.* Heatmapper: web-enabled heat mapping for all. *Nucl. Acids Res.* **44**, 147–153 (2016).
29. Medinger, R. *et al.* Diversity in a hidden world: potential and limitation of next-generation sequencing for surveys of molecular diversity of eukaryotic microorganisms. *Mol. Ecol.* **19**, 32–40 (2010).
30. Weber, A. A. & Pawlowski, J. Can abundance of protists be inferred from sequence data: a case study of Foraminifera. *PLoS ONE* **8**, e56739 (2013).
31. Giner, C. R. *et al.* Environmental sequencing provides reasonable estimates of the relative abundance of specific picoeukaryotes. *Appl. Environ. Microbiol.* **82**, 4757–4766 (2016).
32. Deiner, K. *et al.* Environmental DNA metabarcoding: transforming how we survey animal and plant communities. *Mol. Ecol.* **26**, 5872–5895 (2017).
33. Hering, D. *et al.* Implementation options for DNA-based identification into ecological status assessment under the European Water Framework Directive. *Water Res.* **138**, 192–205 (2018).
34. Kirk, P. M., Cannon, P. F., Minter, D. W. & Stalpers, J. A. *Dictionary of the Fungi*, 10th ed., CAB International, Wallingford, UK, p. 784 (2008).
35. Tedersoo, L. *et al.* High-level classification of the Fungi and a tool for evolutionary ecological analyses. *Fungal Div.* **90**, 135–159 (2018).
36. Rosa, L. H. *et al.* DNA metabarcoding to assess the diversity of airborne fungi present in air over Keller Peninsula, King George Island, Antarctica. *Microb. Ecol.* **82**, 165–172 (2020).
37. Hammer, Ø., Harper, D. A. T. & Ryan, P. D. PAST: paleontological statistics software package for education and data analysis. *Palaeontol. Electron.* **4**, 9 (2001).
38. Nguyen, N. H. *et al.* FUNGuild: an open annotation tool for parsing fungal community datasets by ecological guild. *Fungal Ecol.* **20**, 241–248 (2016).

39. Ellis-Evans, J. C. Fungi from maritime Antarctic freshwater environments. *Br. Antarctic Surv. B.* **68**, 37–45 (1985).
40. da Silva, T. H. *et al.* Diversity, distribution, and ecology of viable fungi in permafrost and active layer of Maritime Antarctica. *Extremophiles* **24**, 565–576 (2020).
41. Benjamin, C. R. Ascocarps of *Aspergillus* and *Penicillium*. *Mycologia* **47**, 669–687 (1955).
42. Yilmaz, N., Visagie, C. M., Houbraeken, J., Frisvad, J. C. & Samson, R. A. Polyphasic taxonomy of the genus. *Talaromyces*. *Stud. Mycol.* **78**, 175–341 (2014).
43. Rosa, L. H. *et al.* Fungi in Antarctica: diversity, ecology, effects of climate change, and bioprospection for bioactive compounds. In *Fungi of Antarctica: Diversity, Ecology and Biotechnological Applications* (ed. Rosa, L. H.) 1–18 (Springer, Berlin, 2019).
44. Miller, J. H., Giddens, J. E. & Foster, A. A. A survey of the fungi of forest and cultivated soils of Georgia. *Mycologia* **49**, 779–808 (1957).
45. Lombard, L., Van Der Merwe, N., Groenewald, J. & Crous, P. Lineages in *Nectriaceae*: Re-evaluating the generic status of *Ilyonectria* and allied genera. *Phytopathol. Mediterr.* **53**, 515–532 (2014).
46. Cabral, A., Groenewald, J. Z., Rego, C., Oliveira, H. & Crous, P. W. *Cylindrocarpon* root rot: multi-gene analysis reveals novel species within the *Ilyonectria radicola* species complex. *Mycol. Prog.* **11**, 655–688 (2012).
47. Parkinson, L., Shivas, R. G. & Dann, E. K. Pathogenicity of *Nectriaceus* fungi on avocado in Australia. *Phytopathology* **107**, 1479–1485 (2017).
48. Brady, N. C. & Weil, R. R. *The nature and properties of soils* (Pearson Prentice Hall, Upper Saddle River, 2008).
49. Rodrigues, W. F., de Oliveira Soares, F., Schaefer, C. E. G., Leite, M. G. P. & Pavinato, P. S. Phosphatization under birds' activity: ornithogenesis at different scales on Antarctic soils. *Geoderma* **391**, 114950 (2021).
50. Rosa, L. H. *et al.* DNA metabarcoding of fungal diversity in air and snow of Livingston Island, South Shetland Islands, Antarctica. *Sci. Rep.* **10**, 1–11 (2020).
51. Rosa, L. H. *et al.* DNA metabarcoding uncovers fungal diversity in soils of protected and non-protected areas on Deception Island, Antarctica. *Sci. Rep.* **10**, 1–9 (2020).
52. de Menezes, G. C. A. *et al.* Fungal diversity present on rocks from a polar desert in continental Antarctica assessed using DNA metabarcoding. *Extremophiles* **25**, 193–202 (2021).
53. Schütte, U. M. *et al.* Effect of permafrost thaw on plant and soil fungal community in a boreal forest: Does fungal community change mediate plant productivity response?. *J. Ecol.* **107**, 1737–1752 (2019).

Acknowledgements

This study received financial support from CNPq, CAPES, FNDCT, FAPEMIG, INCT Criosfera, PROANTAR and Pró-Reitoria de Pesquisa (PRPq) of Universidade Federal University of Minas Gerais. P. Convey is supported by NERC core funding to the British Antarctic Survey's 'Biodiversity, Evolution and Adaptation' Team. We thank the Instituto Antártico Argentino for logistical and financial support for the field campaign in Antarctica during the 2020/21 austral summer and Lagos Group Team for their help in the field.

Author contributions

L.M.D.S., L.H.R., J.M.L., S.H.C. and P.E.A.S.C. conceived the study. L.M.D.S. and L.H.R. performed DNA extraction from lake sediments. F.A.C.L. performed the metabarcoding analysis. F.S.O. performed the sediment physicochemical analysis. L.M.D.S., L.H.R., J.M.L., S.H.C., P.E.A.S.C., F.A.C.L., P.C., F.S.O., M.C.S. and C.A.R. analyzed the results and wrote the manuscript. All authors read and approved the final manuscript.

Competing interests

The authors declare no competing interests.

Additional information

Supplementary Information The online version contains supplementary material available at <https://doi.org/10.1038/s41598-022-12290-6>.

Correspondence and requests for materials should be addressed to L.H.R.

Reprints and permissions information is available at www.nature.com/reprints.

Publisher's note Springer Nature remains neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Open Access This article is licensed under a Creative Commons Attribution 4.0 International License, which permits use, sharing, adaptation, distribution and reproduction in any medium or format, as long as you give appropriate credit to the original author(s) and the source, provide a link to the Creative Commons licence, and indicate if changes were made. The images or other third party material in this article are included in the article's Creative Commons licence, unless indicated otherwise in a credit line to the material. If material is not included in the article's Creative Commons licence and your intended use is not permitted by statutory regulation or exceeds the permitted use, you will need to obtain permission directly from the copyright holder. To view a copy of this licence, visit <http://creativecommons.org/licenses/by/4.0/>.

© The Author(s) 2022